



# 1 Network complexity of rubber plantations is lower than tropical forests for soil

# 2 bacteria but not fungi

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#### 17 Abstract:

- 18 Soil microbial communities play a crucial role in ecosystem functioning. Past
- 19 research has examined the effects of forest conversion on soil microbial composition
- 20 and diversity, but it remains unknown how networks within these communities
- 21 respond to forest conversion such as when tropical rainforest are replaced with rubber
- 22 plantations. In this study, we used Illumina sequencing and metagenome
- 23 shotgun sequencing to analyze bacterial and fungal community network structure in a
- 24 large number of soil samples from tropical rainforest and rubber plantation sites in
- 25 Hainan Island, China. Our results showed only a few shared network edges were
- 26 observed in both bacterial and fungal communities, which indicates that forest
- 27 conversion altered soil microbial network structure. We found a greater degree of
- network structure and a larger number of network edges among bacterial networks in
- samples from tropical rainforest compared to samples from rubber plantations. The
- 30 difference was especially pronounced during the rainy season and indicates that
- 31 rainforest bacterial networks were more complex than rubber plantation bacterial
- 32 networks. However, rubber plantations soil fungal networks showed more higher
- 33 links and higher network degree, suggesting that forest conversion does not reduce
- 34 fungal network complexity. We found that some groups of Acidobacteria were
- 35 keystone taxa in our tropical rainforest soils, while Actinobacteria were keystone taxa
- 36 in rubber plantation soils. In addition, seasonal change had a strong effect on network
- 37 degree, the complexity of soil bacterial and fungal network structure. In conclusion,
- 38 forest conversion changed soil pH and other soil properties, such as available





| 39 | potassium (AK) and total nitrogen (TN), which resulted in changes in bacterial and |
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| 40 | fungal network composition and structure.  |
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| 43 | Keyword: Tropical rainforest, Rubber plantations, Networks, Soil microbial         |
| 44 | community, Forest conversion   |
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# 49 **1. Introduction**

| 50 | Tropical rainforest have the highest biodiversity of any ecosystem and harbor more        |
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| 51 | than 60% of all known plant and animal species (Dirzo and Raven, 2003). However,          |
| 52 | over the past several decades, logging, mining, slash and burn agriculture have caused    |
| 53 | widespread deforestation and forest degradation. Of these, the conversion of forest to    |
| 54 | agriculture has caused the most forest loss (Li et al., 2007).                            |
| 55 | Hainan is home to a large area of tropical rainforest rich in biodiversity. It is a part  |
| 56 | of the Indian-Malay rainforest system and the northern edge of the world's rainforest     |
| 57 | distribution. However, rubber plantations now account for almost a quarter of the total   |
| 58 | extent of vegetated areas on Hainan Island (Lan et al., 2020a).                           |
| 59 | The soil microbiome is highly diverse and comprises up to one quarter of Earth's          |
| 60 | diversity (Wagg et al., 2019). Soil microbes play a critical role in the maintenance of   |
| 61 | soil quality and function, and they represent the majority of biodiversity in terrestrial |
| 62 | ecosystems (Philippot et al., 2013). A number of studies have investigated the impact     |
| 63 | of the conversion of tropical forests to rubber plantations on soil microbial             |
| 64 | composition and diversity (Schneider et al., 2015; Kerfahi et al., 2016, Lan et al.,      |
| 65 | 2017a; 2017b; 2017c; Lan et al., 2020a; 2020b; 2020c). Studies conducted in               |
| 66 | Indonesia (Schneider et al. 2015), Malaysia (Kerfahi et al. 2016) and South China         |
| 67 | (Lan et al. 2017a; Lan et al. 2017b; Lan et al. 2017c) have found significant             |
| 68 | differences between rubber plantations and tropical forests, specifically that the        |
| 69 | diversity of soil bacteria was higher in rubber plantations than in rainforest. Compared  |
| 70 | to primary forests, agricultural systems tend to have higher bacterial richness but       |





| 71 | lower fungal richness (Lan et al., 2017a; Cai et al., 2018; Tripathi et al., 2012; Kerfahi |
|----|--|
| 72 | et al., 2016). However, there are few studies on the effects of forest conversion on soil  |
| 73 | microbial network structure.   |
| 74 | Network analysis is an increasingly popular tool for investigating microbial               |
| 75 | community structure, as it integrates multiple types of information and may represent      |
| 76 | systems-level behavior (Röttjers and Faust, 2018). The soil microbial network is           |
| 77 | viewed as a critical indicator of soil health and quality (Kuperman et al., 2014).         |
| 78 | Network analysis of taxon co-occurrence patterns provides new insight into the             |
| 79 | structure of complex microbial communities, insight that complements and expands           |
| 80 | on the information provided by the more standard suite of analytical approaches            |
| 81 | (Barberan et al., 2012). Previous work has shown that agricultural intensification can     |
| 82 | reduce microbial network complexity (Banerjee et al., 2019). Logging alters soil           |
| 83 | fungal network in tropical rainforests, i.e., a better-organized fungal community in the   |
| 84 | select cut stands when compared with the primary stands (Chen et al., 2019). Soil          |
| 85 | bacterial networks are less stable under drought than fungal networks (De Vries et al.,    |
| 86 | 2018). Soil networks become more connected as ecological restoration progresses            |
| 87 | (Morriën et al., 2017). So far, very few studies have assessed the impact of forest        |
| 88 | conversion on soil microbial networks and it is still unclear whether forest conversion    |
| 89 | as well as seasonal change influences the structure and complexity of microbial            |
| 90 | networks. Here we explored bacterial and fungal community network structure using          |
| 91 | Illumina sequencing based on samples collected from tropical rainforest and rubber         |
| 92 | plantations in Hainan Island, China. We aimed to test the hypothesis that (1) forest       |





- 93 conversion alters microbial networks by altering microbial community composition [3]
- 94 (Lan et al., 2020a) and that soil microbial activity is strongly influenced by plant
- 95 species (Galicia and García-Oliva 2004). (2) Soil microbial network structure in
- 96 rainforest sites is more complex and stable than in rubber plantations because natural
- 97 systems were more connected than artificial systems (Morriën et al., 2017). By
- evaluating these hypotheses, we want to clarify the drivers and mechanisms that link
- 99 forest conversion to differences in soil microbial network structure . This study will
- 100 provide critical information for understanding and managing microbial communities
- 101 in tropical forests of China and elsewhere.
- 102 **2. Methods**
- 103 **2.1 Study site**
- 104 This study was conducted on Hainan Island ( $18^{\circ}10'-20^{\circ}10'$ N and  $108^{\circ}37'-111^{\circ}03'$ E),
- south China. The total area of Hainan Island is about 34,000 km<sup>2</sup> (Lopez et al., 2009).
- 106 Hainan Island is the largest island within the Indo-Burma Biodiversity Hotspot in
- 107 tropical Asia (Francisco-Ortega et al., 2010) and has a tropical monsoon climate.
- 108 Hainan Island has a warm and humid climate all year round, with an average annual
- 109 temperature of 22-26°C. The rainy season occurs from May to October, with a total
- 110 precipitation of about 1500 mm, accounting for 70-90% of the total annual
- 111 precipitation. Only 10-30% of the total annual precipitation falls within the dry
- season, from November to April. Rainfall is abundant, ranging from 1,000 mm to
- 113 2,600 mm yearly, with an average annual precipitation of 1,639 mm. The central part
- 114 of Hainan Island is mountainous and contains old-growth tropical rainforests and





- 115 monsoon forests. Rubber plantations are found on the plateaus surrounding the
- 116 central mountainous zone.
- 117 2.2 Soil sampling
- 118 We selected five rainforests as our study sites: Bangwang mountain, Diaoluo
- 119 mountain, Wuzhi mountain, Yinge mountain and Jianfeng mountain. Five rubber
- 120 plantations were selected in Haikou, Danzhou, Qiongzhong, Wanning and Ledong
- 121 (Figure S1). More information on the study sites is provided in Table S1. For each
- site, thirteen soil samples were collected, thus there were a total of 130 samples
- 123 collected between the rubber plantations and tropical rainforest per sampling interval.
- 124 Soil sampling was performed twice in 2018, once in the rainy season (July) and once
- in the dry season (January). Thus, there were a total of 260 soil samples (130 per
- 126 forest type). Soil samples were divided into two parts: one was used to analyze soil
- 127 water contents, soil pH, total nitrogen, total phosphorus (TP), total potassium (TK),
- 128 nitrate nitrogen (NN), ammonium nitrogen (AN), available phosphorus (AP),
- 129 potassium (AK). The other was used for DNA extraction. Soil properties were
- 130 analyzed following the methods described in by Lan et al. (2020b). Soil properties of
- 131 the rubber plantation and rainforest sites are shown in Table S2.

# 132 2.3 DNA extraction and PCR amplification

- 133 Microbial DNA was extracted from 0.5 g of soil using the E.Z.N.A.® Soil DNA Kit
- 134 (Omega Bio-tek, Norcross, GA, U.S.) following the manufacturer's protocol. The
- 135 fungal ITS1 hypervariable region was amplified using the PCR primers ITS1F
- 136 (5'-CTTGGTCATTTAGAGGAAGTAA-3') and ITS2R





- 137 (5'-GCTGCGTTCTTCATCGATGC-3') (Adams et al., 2013). For bacteria and
- 138 archaea, the V4 hypervariable region of the bacterial 16S rRNA gene was amplified
- using the PCR primers 515FmodF (5'-GTGYCAGCMGCCGCGGTAA-3') and
- 140 806RmodR (5'-GGACTACNVGGGTWTCTAAT-3') (Walters et al., 2016; Sampson
- 141 et al., 2016). The PCR reactions were conducted using the following approach: an
- 142 initial 3 min denaturation at 95°C; followed by 27 cycles of 30s at 95°C, 30s of
- annealing at 55°C, and 45s of elongation at 72°C; and a 10 min final extension at
- 144 72°C.
- 145 2.4 Illumina MiSeq sequencing
- 146 Amplicons were extracted from 2% agarose gels, purified using the AxyPrep DNA
- 147 Gel Extraction Kit (Axygen Biosciences, Union City, CA, U.S.) and quantified using
- 148 a QuantiFluor<sup>TM</sup> -ST Fluorometer (Promega, U.S.). Purified amplicons were pooled in
- an equimolar solution and then sequenced (paired-end,  $2 \times 250$  bp) on an Illumina
- 150 MiSeq platform according to standard protocols.
- 151 Metagenomic shotgun sequencing libraries were prepared and then sequenced by
- 152 Majorbio, Inc. (Shanghai, China) using the Illumina HiSeq 2000 platform. The NR
- 153 gene catalog was aligned against the Kyoto Encyclopedia of Genes and Genomes
- 154 (KEGG) database using BLAST (Version 2.2.28+) and then functionally annotated
- using KOBAAS 2.0 according to previously described methods (Qin et al., 2010)

# 156 2.5 Bioinformatics and data analysis

- 157 Raw fastq files were demultiplexed and quality-filtered using QIIME (Caporaso et al.,
- 158 2010) (version 1.17). Operational Taxonomic Units (OTUs) were clustered with a





| 159 | 97% similarity cut-off using UPARSE (Edgar, 2013), and chimeric sequences were      |
|-----|---|
| 160 | identified and removed using UCHIME. Using the RDP Classifier, the phylogenetic     |
| 161 | affiliation of each 16S rRNA gene and ITS gene sequence was determined using a      |
| 162 | confidence threshold of 70% with the SILVA 16S rRNA database and UNITE              |
| 163 | database, respectively (Amato et al., 2013). The relative abundance was determined  |
| 164 | for each taxon (Good, 1953), and the Shannon and Simpson diversity indices were     |
| 165 | calculated based on re-sampled sequence data using MOTHUR (Schloss et al.,          |
| 166 | 2009). For each site, the relative abundance of different taxa (Good, 1953) and the |
| 167 | Shannon diversity index were calculated based on re-sampled sequence data using     |
| 168 | MOTHUR (Schloss et al., 2009).  |
| 169 | 2.6 Statistical analysis  |
| 170 | For the co-occurrence network analyses, we only focus on the top 300 most abundant  |
|     |   |

171 OTUs of the two forest types. The networks of each habitat during each sampling

172 period (tropical rainforest and rubber plantations in dry season and rainy season) were

- 173 constructed with 65 samples each. Interactions consisted of Spearman's rank
- 174 correlations and co-occurrence networks were constructed using only significant
- 175 correlations of  $\rho > 0.6$  (P < 0.01) (Barberan et al. 2012), because this cutoff includes
- a range of interactions strengths (De Vries et al., 2018). The networks were then
- 177 visualized in R using the *igraph* package. To reveal the distribution pattern of
- correlation coefficients, the frequency of the coefficients of  $\rho > 0.3$  (P < 0.01) were
- 179 plotted. The Network Analyzer tool in Cytoscape (version 3.4.0) was used to calculate
- 180 network topology parameters including number of nodes, edges, degree, betweenness,





| 202 | 3.1 Bacterial and fungal networks  |
|-----|--|
| 201 | 3. Results   |
| 200 | GIANT package of Cytoscape. The $Z_i$ - $P_i$ plot was created with ggplot2 in R.                      |
| 199 | and network hubs ( $Z_i > 2.5$ and $P_i > 0.62$ ). The $Z_i$ and $P_i$ values were calculated using    |
| 198 | peripherals ( $Z_i < 2.5$ and $P_i < 0.62$ ), connectors ( $P_i > 0.62$ ), module hubs ( $Z_i > 2.5$ ) |
| 197 | into four groups according to the simplified criteria (Olesen et al., 2007), namely                    |
| 196 | connected with other modules (Roger and Amaral, 2005). All species can be divided                      |
| 195 | the nodes connectivity within a module, and $P_i$ measures the degree of a node                        |
| 194 | of each node in a network was assessed by the $Z_i$ and $P_i$ values, where $Z_i$ represents           |
| 193 | networks were also constructed on the Major bio cloud platform. The topological role                   |
| 192 | relationship between microbial taxon and environment variables, two-way correlation                    |
| 191 | on the Major bio cloud platform (https://cloud.majorbio.com/). To reveal the                           |
| 190 | (pathway level 3). Then the species and function correlation network was constructed                   |
| 189 | data. Here we used the top 50 most abundant species and top 50 KEGG functions                          |
| 188 | coefficients between species and functions were calculated based on metagenomics                       |
| 187 | degrees of each phylum were calculated and bar plots were created. Correlation                         |
| 186 | proportional influence of each phylum on bacterial and fungal network structure, node                  |
| 185 | betweenness centrality as defined by Berry and Widder (2014). To evaluate the                          |
| 184 | selected on the basis of high network degree, high closeness centrality, and low                       |
| 183 | edges and unique edges which were calculated using igraph. Keystone OTUs were                          |
| 182 | and rubber forest sites, Venn diagrams were plotted to reveal the number of shared                     |
| 181 | closeness. In order to evaluate the network differences between tropical rain forest                   |





Our results showed most nodes of bacterial networks (Figure 1) and fungal networks 203 204 (Figure 2) varied with forest type in both the dry season and rainy season. For 205 bacterial networks, there were 2559 and 2501 edges in tropical rainforest and rubber 206 plantation in dry season respectively, but these two networks only shared 262 edges, accounting only about 10% the total edges (Figure 3A-B). Similarly, these networks 207 only shared 519 edges during the rainy season. For fungal networks, there were only 4 208 and 5 shared edges (i.e., no more than 5% of the total edges) in dry season and rainy 209 season, respectively (Figure 3C-D). 210

211 The number of edges of bacterial and fungal networks were almost equivalent during the dry season. However, in rainy season, there were more edges in the 212 213 bacterial network in tropical rainforest than in the rubber plantation (Table 1 & Figure 214 **3B**). For the network structure of the fungal community, more edges were observed in 215 rubber plantations in rainy season (Table 1 & Figure 3D). Similarly, there were no significant differences in both bacterial and fungal network degree between tropical 216 217 rainforest sites and rubber plantations in the dry season (Figure 4A, C). In the rainy season, rainforest sites had higher bacterial network degree, while rubber plantations 218 219 showed higher fungal network degree (Figure 4B, D). For bacterial networks, more 220 nodes (OTUs) with high degree (rubber plantation had 2 nodes with degree greater 221 than 75, rainforest had 8 such nodes) of rainforest were observed during the rainy 222 season (Figure S2B). For fungal networks, 15 nodes of higher degree (degree greater than 25) were observed in rubber plantations, however, only 7 such nodes existed for 223 224 rainforest sites (Figure S2D). These results indicate rubber plantation fungal network





- structure was more complex than tropical rainforest during the rainy season, but that
- the reverse was true for bacteria.
- When considering the ratio of positive to negative correlation coefficients, more correlations (greater than 0.3, P < 0.05) were calculated, and the results showed that the negative correlations between bacterial and fungal OTUs of rubber plantations were consistently stronger than for tropical rainforest in both dry season and rainy season (Figure 5).
- For both the bacterial and fungal communities, neither tropical rainforest nor
- rubber plantation networks possessed module hubs and network hubs (Figure S3-S4).
- 234 For bacterial network, the majority of nodes in both the rubber plantation and tropical
- rainforest networks were connectors (Figure S3). However, for fungal networks, the
- 236 majority of the nodes in both rubber plantation and tropical rainforest networks were
- 237 peripherals and connectors (Figure S4). The ratio of peripherals and connectors of
- 238 these two forest types was not different indicating the network structures of rainforest
- and rubber plantation were very similar as visualized in Figure 1 and Figure 2.
- 240 However, the bacterial networks had more connectors than fungal networks, which
- 241 suggests bacterial networks may contain more generalists than fungal networks do.
- 242 This indicated bacterial network were more complex than fungal network, which can
- further confirmed by visualization of the network (Figure 1 and Figure 2).
- For bacterial groups, members of the phyla Acidobateria, Planctomycetes and Verrucomicrobia showed higher degree in the tropical rainforest sites than in rubber plantations, suggesting that these taxa are strongly associated with the other members





of the community in tropical rainforest (Figure S5A). Members of the phyla Actinobacteria showed higher degree in rubber plantations. Seasonal change also had effects on network degree for soil bacterial networks. For instance, Chloroflex had higher degree in rubber plantation in the dry season, but the opposite is true in the rainy season. For fungal networks, members of Basidiomycota showed higher degree in tropical rainforest sites duringin the dry season, however, Ascomycota showed higher degree in rubber plantations (Figure S5C-D) during the rainy season.

We used total degree of each phylum to reveal the influence of each phylum on 254 network structure (Figure 6). For bacteria, Proteobacteria, Actinobacteria and 255 Acidobacteria had a large influence on network structure (Figure 6). Acidobacteria 256 257 and Planctomycetes contributed more to rainforest networks than rubber plantation 258 networks. However, Actinobacteria and Chloroflex showed the opposite. For fungi, 259 Ascomycota and Basidiomycota had large influence of network structure. Both Ascomycota and Basidiomycota had stronger influence on rainforest networks than 260 261 rubber plantation networks. The influence of Ascomycota was stronger during the 262 rainy season than in the dry season, indicating seasonal change also had impact on 263 fungal community networks.

264 3.2 Keystone taxa

Keystone OTUs of the bacterial and fungal communities were selected on the basis of
high degree, high closeness centrality, and low betweenness centrality. The results
showed that forest conversion altered the keystone taxa of bacteria and fungi. The
keystone taxa of bacteria were very different between rubber plantations and tropical





| 269 | rainforest sites in | both the dry seasor | n and rainy season. | For bacteria, | there were more |
|-----|---------------------|---------------------|---------------------|---------------|-----------------|
|     |                     |                     |                     |               |                 |

- 270 keystone taxa in tropical rainforest sites than in rubber plantations in both the dry
- 271 season and rainy season indicating that the tropical rainforest networks had higher
- 272 complexity. We found that some groups of Acidobacteria are keystone taxa in tropical
- 273 rainforest sites but disappeared after forest conversion. There were more
- 274 Actinobacteria bacteria in rubber plantations than in tropical rainforest sites (Table
- 275 <mark>S</mark>3).
- 276 For fungi, more keystone taxa were observed in rubber plantations than in tropical
- 277 rainforest sites during both the dry season and rainy season, indicating the rubber
- 278 plantation networks were more complex. Most keystone taxa belong to Ascomycota
- 279 suggesting member of this group are very import for network structure. In addition to
- 280 forest conversion, seasonal changes also affect the keystone taxa of the fungal
- 281 community network. There were more Basidiomycota OTUs in the dry season, but
- 282 more Ascomycota in rainy season (Table S4).

### 283 **3.3 Two-ways correlation networks**

Two-way network analysis of the 50 most abundant species (metageomic data, the 50 most abundant species all belong to bacteria groups) and the 50 most abundant KEGG functions revealed that soil microbial community structure in at rainforests sites was more complex than rubber plantations. (Figure 7). Both rubber plantations and rainforest networks were more complex in the rainy season than in dry season. We also found that metabolism was the most important function in soil microbial network. Surprisingly, species of Actinobacteria negatively correlated with other species and





291 function in rubber plantations (Figure 7).

292 Two-ways correlation network analysis revealed the interaction between microbial 293 composition and environmental variables. This analysis includes different 294 environmental factors as nodes in the network, and the number of connections these nodes have indicates the number of OTUs that are impacted by that environmental 295 factor (Figure 8). For bacteria, elevation had the highest network degree at 106, and 296 was followed by AK (104), soil pH (86) and TK (9). In other words, elevations, AK, 297 soil pH are all drivers of bacterial community composition. Soil pH negatively 298 correlated with most bacterial Acidobcteria OTUs. For fungi, elevation had the 299 highest network degree (61), followed by AK (51), longitude (15), and NN (11). AK 300 301 positively correlated with most OTUs of Basidiomycota. Relationship between OTU 302 abundance and soil pH revealed the soil pH negatively correlated with members of 303 Acidobacteria, but positively correlated with members of Chloroflexi and members of Ascomycota (Figure 9). AK positively correlated with members of Planctomycetes 304 305 Verrucomicrobia and Basidiomycota, however negatively correlated with Chloroflexi 306 and Ascomycota.

307

# 308 4. Discussion

## **309 4.1 Forest conversion reduces soil bacterial network complex**

310 Land-use changes increasingly threaten biodiversity, particularly in tropical forests

- 311 (Gibson et al., 2011). However, we still have little understanding of how soil
- 312 networks response to forest conversion, such when rainforests are converted to rubber





| 313 | plantations. Our results showed that forest conversion had large effects on both soil    |
|-----|--|
| 314 | bacterial and fungal networks. More edges (Table 1) and higher degree (Figure 4) of      |
| 315 | tropical rainforest bacterial networks were observed, especially during the rainy        |
| 316 | season, which indicates that the rainforest bacterial network was more complex than      |
| 317 | the rubber plantation network. This consistent with previous observations that soil      |
| 318 | bacterial networks were more complex in natural systems than in crop soil (Karimi et     |
| 319 | al., 2019). Further study showed that soil networks become more connected as nature      |
| 320 | restoration progresses (Morriën et al., 2017). The observed decrease in network          |
| 321 | complexity and cohesion supports the hypothesis that cropping may enhance the            |
| 322 | isolation of bacterial taxa (Karimi et al., 2019), which results in lower connection of  |
| 323 | the network. In addition, at the microscale, the structure of tilled soils is more       |
| 324 | homogeneous, and the soil pores are less connected than in soils under without tillage   |
| 325 | (Pagliai et al., 2004), such as rainforest soil. In nature, soil ecosystems are highly   |
| 326 | heterogeneous since soil microbial biodiversity hot spots can form spatial and           |
| 327 | temporally within soil aggregates (Bach et al., 2018). This spatial heterogeneity likely |
| 328 | plays an important role for the interactions among microbes and the mechanisms by        |
| 329 | which more complex and diverse communities drive various nutrient cycling                |
| 330 | processes on small spatial scales (Wagg et al., 2018).                                   |
| 331 | A large number of studies employing microbial network analysis have enriched             |
| 332 | our understanding of microbial co-occurrence patterns in various soil ecosystems,        |
| 333 | however, very little is known of whether differences in the structure of microbial       |
| 334 | networks have consequences for microbiome functioning (Wagg et al., 2018). Our           |





| 335 | results demonstrated that more species related with metabolism in natural system than   |
|-----|---|
| 336 | in the agricultural system, especially in the rainy season. This is in line with a      |
| 337 | previous study conducted in Sumatra, Indonesia, which found that the transformation     |
| 338 | of forest to rubber results in a 10-16% decrease in community metabolism (Barnes et     |
| 339 | al., 2014). Fewer interactions between microbial species (most of them are bacteria)    |
| 340 | and functions in rubber plantations demonstrated that forest conversion reduced soil    |
| 341 | bacterial network complex.  |
| 342 | 4.2 Forest conversion does not reduce soil fungal network complexity                    |
| 343 | Surprisingly, rainforest bacterial networks were characterized by fewer edges (Table 1) |
| 344 | and lower degree (Figure 4), which means that rubber plantation bacterial networks      |
| 345 | were more complex than the native forest. Although, our results were consistent with    |
| 346 | previous observations which found that fungal community networks were better            |
| 347 | organized disturbed forest compared to primary forest (Chen et al., 2019). Banerjee et  |
| 348 | al. (2019)'s observation showed that organic agricultural fields harbored much more     |
| 349 | complex fungal networks with many more keystone taxa than conventional managed          |
| 350 | fields. Forest conversion resulted in shifts in fungal composition from Basidiomycota   |
| 351 | to Ascomycota (Figure S7), as seen in previous investigations (Lan et al., 2020a; Lan   |
| 352 | et al., 2020b). Previous work showed that Basidiomycota species show higher drought     |
| 353 | sensitivity than Ascomycota species (Taniguchi et al., 2018), this would result in a    |
| 354 | shift in richness and abundance of Basidiomycota species (Figure S6). Many              |
| 355 | Basidiomycota species are capable of long-distance dispersal (Egidi et al. 2019, Geml   |
| 356 | et al., 2012), which may result in a decrease in fungal network. This possibly          |





- 357 explained why Ascomycota OTUs contribute more to the network structure than
- 358 Basidiomycota (Figure 6). Overall, reduction in abundance and richness of
- 359 Basidiomycota species led to an increase in fungal links in rubber plantations.

#### **360 4.3 Forest conversion enhanced the stability of soil network**

The positive to negative ratio of network links indicates the balance between 361 facilitative and inhibitive relationships within a network (Karimi et al., 2017). 362 Theoretical studies, for example, predict that ecological networks that consist of weak 363 interactions are more stable than those with strong interactions (Neutel et al., 2002, 364 Coyte et al., 2015), and that compartmentalization and presence of negative 365 interactions increase the stability of networks under disturbances (Coyte et al., 2015, 366 367 Rooney et al., 2006, Stouffer & Bascompte 2011). In our study, more negative 368 correlations were detected in rubber plantation, indicating the network structure of 369 rubber plantation soils was more stable than rainforest soils (De Vries et al., 2018).

# **370 4.4 Driver of the network structure**

371 Forest conversion results in the loss of plant diversity, plant biomass and increasing

- soil pH (Lan et al., 2017a, 2017b). Rubber plantations had a significantly higher pH,
- 373 which explains the relative decrease in the abundance of Acidobacteria (Lan et al.,
- 374 2017a). Our results demonstrate that keystone taxa of soil microbes change after
- 375 forest conversion (Table 1). We found that many OTUs of Acidobacteria fit our
- 376 criteria as keystone species for rainforest sites, which is consistent with previous
- findings (Banerjee et al., 2018). Unexpectedly, OTU11388 and OTU11373, both
- 378 Acidobacteria, were observed in rainforest soils in both the dry and rainy seasons,





| 379 | indicating Acidobacteria were very important for rainforest soil bacterial networks    |
|-----|--|
| 380 | (Figure 6 and table S3). Higher AK concentration resulted in a higher abundance and    |
| 381 | more taxa of of Actinobacteria (Figure 9), which suggests that Actinobacteria          |
| 382 | contributed more in rubber plantation than in rainforest (Figure 6). Indeed, forest    |
| 383 | conversion reduced the abundance of Actinobacteria OTUs (Figure S7) Due to the         |
| 384 | human disturbance in rubber plantations, the soil will inevitably be slightly polluted |
| 385 | with herbicides and domestic garbage. Previous study showed member of                  |
| 386 | Actinobacteria were observed in contaminated soil (Jiao et al., 2016).                 |
| 387 | Forest conversion also increases land use intensity (Brinkmann et al., 2019),          |
| 388 | including the application of fertilizer and herbicide. Herbicide application also      |
| 389 | caused significant decreases in root colonization and spore biomass of arbuscular      |
| 390 | mycorrhizal fungi in tropical agriculture (Zaller et al., 2014). Soil nutrient         |
| 391 | concentration shows a decline around the roots of rubber plantations compared to       |
| 392 | those from rainforests (Sahner et al., 2015). Our observation is no exception, for     |
| 393 | instance, AK and TN concentration was significant lower in rubber plantation than in   |
| 394 | samples from rainforest sites (Table S2). Higher concentration of AK reasonably        |
| 395 | explained the higher contribution of Basidiomycota on the network structure (Figure    |
| 396 | 8B) due to AK positive association with Basidiomycota.                                 |
| 397 | Spatiotemporal heterogeneity can be a major driver of the abundance and                |
| 398 | distribution of keystone taxa in soil which is a highly heterogeneous and multifaceted |
| 399 | environment (Mills et al., 1993, Power et al., 1996; Mouquet et al., 2013). Seasonal   |
| 400 | variability determines the structural and compositional properties of microbiomes in   |





| 401 | an environment, and as such, a keystone species might be present only in a specific     |
|-----|---|
| 402 | season or time period (Banerjee et al., 2018). It was interesting that more bacterial   |
| 403 | OTUs were identified as connectors during the rainy season than in the dry season.      |
| 404 | Connectors have been characterized as generalists (Olesen et al., 2007), and            |
| 405 | generalists drive covariation among communities in a network (Chen et al., 2019).       |
| 406 | Previous observation demonstrated that some keystone taxa that were found in the dry    |
| 407 | season disappeared during the rainy season (Lan et al., 2018) . Seasonal changes        |
| 408 | possibly explained the keystone taxa was observed in rainy season but not in dry        |
| 409 | season.   |
| 410 | 5. Summary  |
| 411 | Our knowledge about land-use impacts on soil ecosystems is mostly limited to            |
| 412 | biodiversity and ecosystem functions, leaving uncertainty about how soil networks       |
| 413 | change after forest conversion. This study is the most comprehensive report on          |
| 414 | changes in network structure that occur when tropical rainforests are converted into    |
| 415 | rubber forest. Our study showed that forest conversion altered both bacterial and       |
| 416 | fungal soil networks, reduced bacterial network complexity and enhanced fungal          |
| 417 | network complexity, especially during the rainy season. One possible reason maybe       |
| 418 | that forest conversion changed soil pH and other soil properties, which altered         |
| 419 | bacterial composition and subsequent network structure. Our study demonstrates the      |
| 420 | impact of forest conversion for soil network structure, which has important             |
| 421 | implications for ecosystem functions and health of soil ecosystems in tropical regions. |
| 422 |   |





# 423

# 424 Availability of data and material

- 425 The raw reads were deposited into the NCBI Sequence Read Archive (SRA) database
- 426 (Accession Number: SRP108394, SRP278296, SRP278319).

# 427 Code availability

428 Not applicable

# 429 Authors' contributions

- 430 Guoyu Lan: Conceptualization, Methodology, Writing- Reviewing and Editing;
- 431 Chuan Yang and Zhixiang Wu: Investigation

# 432 Competing interests

433 The authors declared that they have no conflicts of interest to this study.

### 434 Disclaimer

- 435 Publisher's note: Copernicus Publications remains neutral with regard to jurisdictional
- 436 claims in published maps and institutional affiliations.

### 437 Acknowledgements

- 438 We thank Dr. Tim Treuer for his assistance with English language and grammatical
- 439 editing.

# 440 Financial support

- 441 This work was supported by Finance Science and Technology Project of Hainan
- 442 Province (ZDYF2019145); National Natural Science Foundation of China
- 443 (31770661); High level Talents Project of Hainan Natural Science Foundation
- 444 (320RC733); the Earmarked Fund for China Agriculture Research System





445 (CARS-33-ZP3)

446

447 **References** 

- 448 Adams RI, Miletto M, Taylor JW, Bruns T D. Dispersal in microbes: fungi in indoor
- 449 air are dominated by outdoor air and show dispersal limitation at short distances,
- 450 ISME J., 7(7):1262-1273, 2013
- 451 Amato, K.R., Yeoman, C.J., Kent, A., Carbonero, F., Righini, N., Estrada, A.E.,
- 452 Gaskins, H.R., Stumpf, R.M., Yildirim, S., Torralba, M., Gillis, M., Wilson,
- 453 B.A., Nelson, K.E., White, B.A., Leigh, S.R., Habitat degradation impacts

454 primate gastrointestinal microbiomes, ISME J., 7: 1344-1353, 2013

- 455 Bach, E. M., Williams, R. J., Hargreaves, S. K., Yang, F. & Hofmockel, K. S.
- 456 Greatest soil microbial diversity found in micro-habitats, Soil Biol. Biochem.,

457 118: 217-226, 2018

461

458 Banerjee, S., Schlaeppi, K., van der Heijden, M.G.A. Keystone taxa as drivers of

459 microbiome structure and functioning, Nat. Rev. Microbiol., 16: 567-576, 2018;

- 460 Banerjee, S., Walder, F., Büchi, L., Meyer, M., Held, A.Y., Gattinger, A., Keller, T.,

Charles, R., van der Heijden, M.G.A. Agricultural intensification reduces

462 microbial network complexity and the abundance of keystone taxa in roots,

463 ISME J. 2019; 13:1722-1736.

Barberan, A., Bates, S.T., Casamayor, E.O., Fierer, N. Using network analysis to
explore co-occurrence patterns in soil microbial communities, ISME J.,
6:343-351,2012





| 467 | Barnes, A. D. Jochum, M., Mumme, S., Haneda, N. F., Farajallah, A., Widarto, T. H.       |
|-----|--|
| 468 | & Brose, U. Consequences of tropical land use for multitrophic biodiversity and          |
| 469 | ecosystem functioning, Nat Commun., 5: 5351, 2014  |
| 470 | Berry D, Widder S. Deciphering microbial interactions and detecting keystone species     |
| 471 | with co-occurrence networks, Front Microbiol., 5:1-14, 2014;.                            |
| 472 | Brinkmann N., Schneider D., Sahner J., Ballauff J., Edy N., Barus H., Irawan B., Budi    |
| 473 | S.W., Qaim M., Danie R., Polle, A Intensive tropical land use massively shifts           |
| 474 | soil fungal communities, Sci. Rep., 9:3403, 2019   |
| 475 | Cai, Z.Q., Zhang, Y. H, Yang, C., Wang, S. Land-use type strongly shapes community       |
| 476 | composition, but not always diversity of soil microbes in tropical China, Catena         |
| 477 | 165: 369-380, 2018;  |
| 478 | Caporaso, J.G., Kuczynski, J., Stombaugh, J., Bittinger, K., Bushman, F.D., Costello,    |
| 479 | E.K., Fierer, N., Peña, A.G., Goodrich, J.K., Gordon, J.I., QIIME allows analysis        |
| 480 | of high-throughput community sequencing data, Nat. Methods, 7: 335-336, 2010             |
| 481 | Chen, J., Xu, H., He, D., Li, Y.D., Luo, T.S., Yang, H.G., Lin, M.X., 2019. Historical   |
| 482 | logging alters soil fungal community composition and network in a tropical               |
| 483 | rainforest, For. Ecol. Manage., 433 (5): 228-239, 2019                                   |
| 484 | Coyte, K. Z., Schluter, J. Foster, K. R. The ecology of the microbiome: networks,        |
| 485 | competition, and stability, Science, 350: 663-666,2015;                                  |
| 486 | De Vries, V. F. T., Griffiths, R. I., Mark, B., Hayley, C., Mariangela, G., Soon, G. H., |
| 487 | et al. Soil bacterial networks are less stable under drought than fungal networks,       |

Nature Commun. 9(1): 3033, 2018

488





- 489 Dirzo, R., Raven, P.H. Global state of biodiversity and loss, Annu. Rev. Environ,
- 490 Resour. 28:137-167, 2003
- 491 Edgar, R.C. UPARSE: Highly accurate OTU sequences from microbial amplicon
- 492 reads, Nat. Methods 10:996-998, 2013
- 493 Egidi, E., Delgado-Baquerizo, M., Plett, J.M., Wang, J. Eldridge, D. J., Bardgett, R.D.,
- 494 Maestre, F.T., Singh, B.K. A few Ascomycota taxa dominate soil fungal
- 495 communities worldwide. Nat. Commun. 10: 2369,2019;
- 496 Francisco-Ortega, J., Wang, Z.S., Wang, F.G., Xing, F.W., Liu, H., Xu, H., Xu, W.X.,
- 497 Luo, Y.B., Song, X.Q., Gale, S., Boufford, D.E., Maunder, M., An, S.Q., Seed
- 498 plant endemism on Hainan Island: a framework for conservation actions, Bot.
- 499 Rev. 76: 346-376,2010
- 500 Galicia, L., García-Oliva, F. The effects of C, N and P additions on soil microbial
- 501 activity under two remnant tree species in a tropical seasonal pasture, Appl. Soil
- 502 Ecol. 2004; 26:(1) 31-39
- 503 Geml, J., Timling, I., Robinson, C. H., Lennon, N., Nusbaum, H.C., Brochmann,
- 504 C.; Noordeloos, M.E.; Taylor, D. L. An arctic community of symbiotic fungi
- 505 assembled by dispersers: phylogenetic diversity of ectomycorrhizal
- 506 basidiomycetes in Svalbard based on soil and sporocarp DNA, J. Biogeogr., 39:
- 507 74-88,2012;
- 508 Gibson L., Lee, T. M., Koh, L. P., Brook, B.W., Gardner, T.A., Barlow, J., Peres,
- 509 C.A., Bradshaw, C. J. A., Laurance, W.F., Lovejoy T.E., Sodhi, N.S. Primary
- 510 forests are irreplaceable for sustaining tropical biodiversity, Nature, 478,





- 511 378-381, 2011
- 512 Good, I. L. 1953. The population frequencies of species and the estimation of
- 513 population parameters, Biometrika, 40: 237-264,1953
- 514 Jiao, S., Liu, Z.S., Lin, Y.B., Yang, J., Chen, W.M., Wei, G.H. Bacterial communities
- in oil contaminated soils: biogeography and co- occurrence patterns. Soil Biol.
- 516 Biochem. 98:64-73, 2016
- 517 Karimi, B., Dequiedt, S., Terrat, Sébastien, Jolivet, C., Arrouays, D., & Wincker, P.,
- 518 Cruaud C., Bispo A., Prévost-Bouré N. C., Ranjard L. Biogeography of soil
- 519 bacterial networks along a gradient of cropping intensity, Sci Rep., 9(1): 3812,
- 520 2019
- 521 Karimi, B., Maron, P.A., Chemidlin-Prevost Boure, N., Bernard, N., Gilbert, D.,
- 522 Ranjard, L. Microbial diversity and ecological networks as indicators of

environmental quality, Environ. Chem. Lett., 15: 265-281, 2017

- 524 Kerfahi, D., Tripathi, B.M., Dong, K., Go, R., Adams, J.M. Rainforest conversion to
- rubber plantation may not result in lower soil diversity of bacteria fungi and
  nematodes, Microb. Ecol., 72: 359-371, 2016
- 527 Kuperman, R.G., Siciliano, S.D., Römbke, J., Oorts, K., Deriving site-specific soil
- 528 clean-up values for metals and metalloids: rationale for including protection of
- soil microbial processes, Integr. Environ. Assess. Manage., 10 (3):388-400, 2014
- 530 Lan, G.Y., Li, Y.W., Wu, Z.X., Xie GS. Soil bacterial diversity impacted by
- 531 conversion of secondary forest to rubber or eucalyptus plantations—a case study
- of Hainan Island, south China, For Sci., 63:87-93, 2017a





| 533 | Lan, G.Y., Li, Y.W., Jatoi, M.T., Tan, Z.H., Wu, Z.X., Xie, G.S. Change in soil       |
|-----|---|
| 534 | microbial community compositions and diversity following the conversion of            |
| 535 | tropical forest to rubber plantations in Xishuangbanan southwest China, Trop.         |
| 536 | Conserv. Sci., 10: 1-14, 2017b  |
| 537 | Lan, G.Y., Li, Y., Wu, Z.X., Xie, G.S. Impact of tropical forest conversion on soil   |
| 538 | bacterial diversity in tropical region of China, Eur. J. Soil Biol., 83: 91-97, 2017c |
| 539 | Lan, G.Y., Li,Y.W., Lesueur, D., Wu, Z.X., Xie, G.S. Seasonal changes impact soil     |
| 540 | bacterial communities in a rubber plantation on Hainan Island China, Sci. Total       |
| 541 | Environ., 626c: 826-834, 2018   |
| 542 | Lan, G.Y., Wu, Z.X., Sun, R., Yang, C., Chen, B.Q., Zhang, X.C. Forest conversion     |
| 543 | changed the structure and functional process of tropical forest soil microbiome,      |
| 544 | Land. Degrad. Dev. 2020a. DOI: 10.1002/ldr.3757                                       |
| 545 | Lan, G.Y., Wu, Z.X., Sun, R., Yang, C., Chen, B.Q., Zhang, X. Tropical rainforest     |
| 546 | conversion into rubber plantations results in changes in soil fungal composition,     |
| 547 | but underling mechanisms of community assembly remain unchanged,                      |
| 548 | Geoderma, 375: 114505, 2020b  |
| 549 | Lan, G.Y., Wu, Z.X., Li, Y.W., Chen, B.Q. The drivers of soil bacterial communities   |
| 550 | in rubber plantation at local and geographic scales, Arch. Agron. Soil. Sci., 66(3):  |
| 551 | 358-369, 2020c  |

- Li., H.M., Aide, T., Ma, Y.X., Liu, W.J., Cao, M. Demand for rubber is causing the 552 loss of high diversity rain forest in SW China, Biodiver. Conserv., 16 (6): 553
- 1731-1745, 2007 554





- 555 Lopez, S., Rousset, F., Shaw, F.H., Ruth, G., Shaw, R.G., Ophélie, R. Joint effects of
- 556 in? breeding and local adaptation on the evolution of genetic load after
- 557 fragmentation, Conserv. Biol., 23: 1618-1627, 2009
- 558 Mills, L.S., Soulé, M.E., Doak, D.F. The keystone- species concept in ecology and
- conservation, BioScience, 43: 219-224, 1993
- 560 Morriën, E., Hannula, S., Snoek, L. et al. Soil networks become more connected and
- take up more carbon as nature restoration progresses, Nat. Commun., 8: 14349,
- 562 2017
- 563 Mouquet, N., Gravel, D., Massol, F., Calcagno, V. Extending the concept of keystone
- species to communities and ecosystems, Ecol. Lett., 16:1-8, 2013
- 565 Neutel, A. M., Heesterbeek, J. A. P., de Ruiter, P. C. Stability in real food webs: weak
- 566 links in long loops, Science, 296:1120-1123, 2002
- 567 Olesen, J.M., Bascompte, J., Dupont, Y.L., Jordano, P. The modularity of pollination
- 568 networks, Proc. Nat. Acad. Sci. USA, 104: 19891-19896, 2007
- 569 Pagliai, M., Vignozzi, N., Pellegrini, S. Soil structure and the effect of management
- 570 practices. Soil Tillage Res. 79: 131-143, 2004
- 571 Philippot, L., Spor, A., He'nault, C., Bru, D., Bizouard, F., Jones, C. M., Sarr, A.,
- 572 Maron, P.A. Loss in microbial diversity affects nitrogen cycling in soil Laurent,
- 573 ISME J., 7:1609-1619, 2013
- 574 Power M, Tilman D, Estes J, Menge B, Bond W, Mills S, Daily G, Castilla J,
- 575 Lubchenco J, Paine R, Power M, Tilman D, Estes J, Menge B, Bond W, Mills L,
- 576 Daily G, Castilla J, Lubchenco J, Paine R. Challenges in the quest for keystones,





- 577 Bioscience. 46: 609-620,1996
- 578 Qin, J., Li, R., Raes, J., Arumugam, M., Burgdorf, K.S., Manichanh, C., Nielsen, T.,
- 579 Pons, N., Levenez, F., Yamada T. A human gut microbial gene catalogue
- established by metagenomic sequencing, Nature, 464: 59-65, 2010
- 581 Roger, G., Amaral, L.A.N., Functional cartography of complex metabolic networks,
- 582 Nature, 433: 895,2005
- 583 Rooney, N., McCann, K., Gellner, G., Moore, J. C. Structural asymmetry and the
- stability of diverse food webs, Nature, 442, 265-269, 2006
- 585 Röttjers, L, Faust, K. From hairballs to hypotheses-biological insights from microbial
- networks, FEMS Microbiol. Rev., 10:1093, 2018
- 587 Sahner, J., Budi, S. W., Barus, H., Edy, N., Meyer, M., Corré, MD, Polle, A.
- 588 Degradation of root community traits as indicator for transformation of tropical
- lowland rain forests into oil palm and rubber plantations, PloS One, 10(9):
- e0138077, 2015
- 591 Sampson, T.R., Debelius J W, Thron T, Janssen, S., Shastri, G.G., Ilhan, Z.E.,
- 592 Challis, C., Schretter, C.E., Rocha, S., Gradinaru, V., Chesselet, M.F.,
- 593 Keshavarzian, A., Shannon, K. M., Krajmalnik-Brown, R., Wittung-Stafshede,
- 594 P., Knight, R., Mazmanian, S.K. Gut microbiota regulate motor deficits and
- neuroinflammation in a model of Parkinson's disease. Cell, 167(6): 1469-1480,
- 596 2016
- 597 Schloss, P.D., Westcott, S.L., Ryabin, T., Hall, J.R., Hartmann, M., Hollister, E.B.,
- 598 Lesniewski, R.A, Oakley, B.B., Parks, D.H., Robinson, C.J. Introducing mothur:





| 599 | open-source, platform-independent, community-supported software for                   |  |  |  |  |  |  |  |  |
|-----|---|--|--|--|--|--|--|--|--|
| 600 | describing and comparing microbial communities. Appl Environ Microbiol. 75:           |  |  |  |  |  |  |  |  |
| 601 | 7537-7541, 2009   |  |  |  |  |  |  |  |  |
| 602 | Schneider, D., Engelhaupt, M., Allen, K., Kurniawan, S., Krashevska, V., Heinemann,   |  |  |  |  |  |  |  |  |
| 603 | M., Scheu, S. Impact of lowland rainforest transformation on diversity and            |  |  |  |  |  |  |  |  |
| 604 | composition of soil prokaryotic communities in Sumatra Indonesia, Front.              |  |  |  |  |  |  |  |  |
| 605 | Microbiol., 6: 296,2015;  |  |  |  |  |  |  |  |  |
| 606 | Stouffer, D. B., Bascompte, J. Compartmentalization increases food-web persistence,   |  |  |  |  |  |  |  |  |
| 607 | Proc. Natl. Acad. Sci. USA. 108:3648-3652, 2011                                       |  |  |  |  |  |  |  |  |
| 608 | Taniguchi, T., Kitajima, K., Douhan, G.W., Yamanaka, N., Allen, M.F. A pulse of       |  |  |  |  |  |  |  |  |
| 609 | summer precipitation after the dry season triggers changes in ectomycorrhizal         |  |  |  |  |  |  |  |  |
| 610 | formation, diversity, and community composition in a Mediterranean forest in          |  |  |  |  |  |  |  |  |
| 611 | California, USA. Mycorrhiza. 28(7): 665-677, 2018                                     |  |  |  |  |  |  |  |  |
| 612 | Tripathi, B. M., Kim, M., Singh, D., Lee-Cruz, L., Lai-Hoe, A., Ainuddin, A. N.,      |  |  |  |  |  |  |  |  |
| 613 | Adams, J. M. Tropical soil bacterial communities in Malaysia: pH dominates in         |  |  |  |  |  |  |  |  |
| 614 | the equatorial tropics too, Microb Ecol., 64: 474-484,2012                            |  |  |  |  |  |  |  |  |
| 615 | Wagg, C., Schlaeppi, K., Banerjee, S., Kuramae, E.E., van der Heijden, M.G.A.         |  |  |  |  |  |  |  |  |
| 616 | Fungal-bacterial diversity and microbiome complexity predict ecosystem                |  |  |  |  |  |  |  |  |
| 617 | functioning, Nat Commun., 10:4841, 2019   |  |  |  |  |  |  |  |  |
| 618 | Walters, W., Hyde, E. R., Berg-Lyons, D., Ackermann, G., Humphrey, G., Parada, A.,    |  |  |  |  |  |  |  |  |
| 619 | Gilbert, J. A., Jansson, J. K., Caporaso, J. G., Fuhrman, J. A., Apprill, A., Knight, |  |  |  |  |  |  |  |  |
|     |   |  |  |  |  |  |  |  |  |

620 R. Improved bacterial 16S rRNA gene (V4 and V4-5) and fungal internal





| 621 | transcribed spacer marker gene primers for microbial community surveys,                   |  |  |  |  |  |  |  |
|-----|---|--|--|--|--|--|--|--|
| 622 | mSystems, 1(1): e00009-15, 2016   |  |  |  |  |  |  |  |
| 623 | Zaller, J. G., Heigl, F., Ruess, L., Grabmaier, A. Glyphosate herbicide affects           |  |  |  |  |  |  |  |
| 624 | belowground interactions between earthworms and symbiotic mycorrhizal fungi               |  |  |  |  |  |  |  |
| 625 | in a model ecosystem, Sci Rep., 4:5634,2014   |  |  |  |  |  |  |  |
| 626 |   |  |  |  |  |  |  |  |
| 627 | Figure legend   |  |  |  |  |  |  |  |
| 628 | Figure 1 Soil bacterial network structure of rubber plantations and tropical rainforest   |  |  |  |  |  |  |  |
| 629 | in dry and rainy seasons. Red lines indicate positive correlation between OTUs, and       |  |  |  |  |  |  |  |
| 630 | green indicate negative correlation. Absolute value of correlation coefficient > 0.6, $P$ |  |  |  |  |  |  |  |
| 631 | < 0.05  |  |  |  |  |  |  |  |
| 632 | Figure 2 Soil fungal network structure of rubber plantations and tropical rainforest in   |  |  |  |  |  |  |  |
| 633 | dry and rainy season. Red lines indicate positive correlation between OTUs, and           |  |  |  |  |  |  |  |
| 634 | green indicate negative correlation. Absolute value of correlation coefficient > 0.6, $P$ |  |  |  |  |  |  |  |
| 635 | < 0.05  |  |  |  |  |  |  |  |
| 636 | Figure 3 Soil microbial (bacterial and fungal) network of tropical rainforest and         |  |  |  |  |  |  |  |
| 637 | rubber plantations in dry season and rainy season.  |  |  |  |  |  |  |  |
| 638 | Figure 4 Network degree of soil bacterial and fungal community of rubber plantations      |  |  |  |  |  |  |  |
| 639 | (blue) and tropical rainforest (red) in dry season and rainy season.                      |  |  |  |  |  |  |  |
| 640 | Figure 5 Frequency distributions of correlations in bacterial (a: dry season, b: rainy    |  |  |  |  |  |  |  |
| 641 | season) and fungal (c: dry season, d: rainy season) networks of rubber plantations and    |  |  |  |  |  |  |  |
| 642 | tropical rainforest in the dry season and rainy season. (Absolute correlation coefficient |  |  |  |  |  |  |  |





- greater than 0.3, P < 0.05) Correlations in rainforest networks are red, correlations in
- 644 rubber plantation networks are blue.
- 645 Figure 6 Proportional influence of different phylum on bacterial and fungal network
- 646 structure in both dry season and rainy. The influence was the number of degrees of
- 647 nodes belonging to a particular phylum. (a: bacteria in dry season, b: bacteria in rainy
- 648 season, c: fungal in dry season, d: fungal in rainy season).

**Figure 7** Network of the top 50 most abundant species (based on metagenomics data) and top 50 most frequent KEGG functions (pathway level 3) of rubber plantations and tropical rainforest sites in dry season and rainy season. (A: rubber in dry season; B: rainforest in dry season; C: rubber in rainy season; D: rainforest in rainy season) The size of the node indicates the species/function abundance. A red line indicates positive correlation between species/functions, and green indicates negative correlation. Absolute value of correlation coefficient > 0.6, p < 0.05

**Figure 8** Two ways correlation network of top 500 most abundant bacterial (A) and fungal (B) OTUs and environmental factors. The size of the node indicates the OTU abundance. A red line indicates positive correlation between species/functions, and green indicates negative correlation. Absolute value of correlation coefficient > 0.5, p < 0.05.

Figure 9 Relationship between abundance of phylum (bacteria: A-E, I-M; fungi: F-H,
N-P) and soil properties (Soil pH: A-H; AK (available potassium) concentration: I-P)





664 Table 1 Topological properties of soil microbial (bacterial and fungi) network

665 structure in rubber plantation and tropical rain forest in dry season and rainy season

|                                 | Bacteria | 1          |        |           | Fungi    |           |          |            |
|---------------------------------|----------|------------|--------|-----------|----------|-----------|----------|------------|
|                                 | Rubber   | Rainforest | Rubber | Rainfores | t Rubber | Rainfores | t Rubber | Rainforest |
|                                 | Dry      | Dry        | Rainy  | Rainy     | Dry      | Dry       | Rainy    | Rainy      |
| No. of nodes                    | 291      | 287        | 296    | 296       | 220      | 235       | 243      | 244        |
| No. of edges                    | 2448     | 2559       | 4248   | 5019      | 791      | 769       | 1250     | 905        |
| No. of positive edges           | 2052     | 2508.00    | 3385   | 4901      | 760      | 764       | 1195     | 897        |
| No. of negative edges           | 396      | 51         | 863    | 118.00    | 31       | 5         | 55       | 8          |
| Connectance                     | 0.06     | 0.06       | 0.09   | 0.11      | 0.02     | 0.02      | 0.03     | 0.02       |
| Average degree                  | 16.82    | 17.83      | 16.67  | 33.91     | 16.67    | 6.54      | 10.28    | 7.41       |
| Average betweenness             | 100.61   | 93.52      | 31.23  | 42.57     | 208.90   | 185.03    | 170.99   | 266.31     |
| Average of shortest path length | n 2.92   | 2.92       | 2.56   | 2.45      | 4.00     | 3.626     | 3.81     | 5.27       |
| Diameter                        | 6.00     | 7.00       | 7.00   | 6.00      | 10.00    | 12.00     | 9.00     | 15.00      |
| Cluster of coefficient          | 0.46     | 0.451      | 0.51   | 0.54      | 0.59     | 0.457     | 0.49     | 0.51       |
| No of clusters                  | 11.00    | 14.00      | 5.00   | 5.00      | 94.00    | 75.00     | 62.00    | 67.00      |
| Degree centralization           | 0.12     | 0.11       | 0.17   | 0.18      | 0.08     | 0.08      | 0.10     | 0.10       |
| Betweenness centralization.     | 0.0066   | 0.0067     | 0.0053 | 0.0049    | 0.02     | 0.03      | 0.121    | 0.04       |
| Closeness centralization.       | 0.35     | 0.35       | 0.40   | 0.41      | 0.31     | 0.351     | 0.291    | 0.26       |
| Neighborhood Connectivity       | 21.12    | 21.79      | 34.69  | 40.68     | 10.45    | 9.28      | 13.78    | 9.44       |
| Topological coefficient         | 0.26     | 0.26       | 0.26   | 0.27      | 0.33     | 0.36      | 0.37     | 0.50       |







666

667Figure 1 Soil bacterial network structure of rubber plantations and tropical rainforest in dry and668rainy season. Red line indicates positive correlation between OTUs, and green indicates negative669correlation. Absolute value of correlation coefficient > 0.6, p < 0.05







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**Figure 2** Soil fungal network structure of rubber plantations and tropical rainforest in dry and

rainy season. Red line indicates positive correlation between OTUs, and green indicates negative correlation. Absolute value of correlation coefficient > 0.6, p < 0.05







680 in dry season and rainy season.







682 683

Figure 4 Network degree betweenness of soil bacterial and fungal community of rubberplantations (blue) and tropical rainforest (red) in dry season and rainy season.





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**Figure 5** Frequency distributions of correlations in bacterial (a: dry season, b: rainy season) and fungal (c: dry season, d: rainy season) networks of rubber plantations and tropical rainforest in dry season and rainy season. (Absolute correlation coefficient greater than 0.3, p < 0.05) Correlations in rainforest networks are red, correlations in rubber plantation networks are blue.





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Figure 6 Proportional influence of different phylum on bacterial and fungal network structure in
both dry season and rainy. The influence was the number of degrees of nodes belonging to a
particular phylum. (a: bacteria in dry season, b: bacteria in rainy season, c: fungal in dry season, d:
fungal in rainy season).

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**Figure 7** Network of the top 50 abundant species (based on metagenomics data) and top 50 abundant KEGG function (pathway level 3) of rubber plantations and tropical rainforest in dry season and rainy season. (A: rubber in dry season; B: rainforest in dry season; C: rubber in rainy season; D: rainforest in rainy season) The size of the node indicates the species/function abundance. Red line indicates positive correlation between species/functions, and green indicates negative correlation. Absolute value of correlation coefficient > 0.6, p < 0.05







711

712 Figure 8 Two ways correlation network of top 500 abundant bacterial (A) and fungal (B) OTU 713 and environmental factors. The size of the node indicates the OTU abundance. Red line indicates

714 positive correlation between species/functions, and green indicates negative correlation. Absolute

715 value of correlation coefficient > 0.5, p < 0.05.







717 718

719 Figure 9 Relationship between abundance of phylum (bacteria: A-E, I-M; fungi: F-H,

N-P) and soil properties (Soil pH: A-H; AK (available potassium) concentration: I-P)