# 1 Network complexity of rubber plantations is lower than tropical forests for soil

# 2 bacteria but not fungi

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# 17 Abstract:

Soil microbial communities play a crucial role in ecosystem functioning. Past 18 research has examined the effects of forest conversion on soil microbial composition 19 and diversity, but it remains unknown how networks within these communities 20 21 respond to forest conversion, including when tropical rainforests are replaced with rubber plantations. Microbial networks are viewed as critical indicators of soil health 22 23 and quality, They consist of two parts: nodes and edges. In this study, we used data 24 from Illumina sequencing and metagenome shotgun sequencing to analyze bacterial and fungal community network structure in a large number of soil samples from 25 tropical rainforests and rubber plantation sites in Hainan Island, China. Our results 26 27 showed only 5-10% shared network edges (i.e., links between species A and B existing in both rubber plantations and rainforests) were observed in both bacterial 28 29 and fungal communities, which indicates that forest conversion altered soil microbial 30 network structure. The identity of keystone OTUs differed entirely between rubber plantation and rainforest sites, further underscoring the altered network structure, 31 More edges and more negative correlations within the soil bacterial-fungal networks 32 were observed in rubber plantation sites (dry season: 4,284 total edges, 844 negative; 33 rainy season: 7,257 total, 1,744 negative) than in rainforest sites (dry season: 3,650 34 total, 149 negative; rainy season: 6,018 total, 489 negative), demonstrating that soil 35 36 bacterial-fungal network structure was more complex and stable in rubber plantations than in rainforests. For bacteria, a larger number of network edges were observed 37 among bacterial networks in samples from tropical rainforest than in samples from 38 rubber plantations, indicating that rainforest bacterial networks were more complex 39 40 than those from rubber plantations. However, soil fungal networks from rubber plantations showed more links, suggesting that forest conversion increased fungal 41 42 network complexity. More edges of network and more links between species and functions were observed in the rainy season than in the dry season indicating that 43 44 seasonal changes had a strong effect on network structure and function. Further

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- 67 analysis shows soil pH, potassium (AK), and total nitrogen (TN) had more links with
- 68 species of some phyla. In conclusion, forest conversion results in an increase in soil
- 69 pH, and a decrease in AK and TN, and these changes as well as seasonal variations
- 70 had a great impact on soil microbial composition, network structure and function.
- 71 Copyright statements: The copyright statement will be included by Copernicus, if72 applicable.
- 73 Keyword: Tropical rainforest, Rubber plantations, Networks, Soil microbial
- 74 community, Forest conversion
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# 76 1. Introduction

77 'Soil microbiome' refers to the enormous variety of microorganisms inhabiting the complex soil environment. It is highly diverse and comprises up to one quarter of 78 Earth's species (Wagg et al., 2019). Soil microbes play a critical role in the 79 80 maintenance of soil quality and function, and they represent more than half of the 81 biodiversity of terrestrial ecosystems (Philippot et al., 2013). A number of studies 82 have investigated how the conversion of tropical forests to rubber plantations impacts soil microbial composition and diversity (Schneider et al., 2015; Kerfahi et al., 2016, 83 Lan et al., 2017a; 2017b; 2017c; Lan et al., 2020a; 2020b; 2020c). Studies conducted 84 in Indonesia (Schneider et al. 2015), Malaysia (Kerfahi et al. 2016) and South China 85 86 (Lan et al. 2017a; Lan et al. 2017b; Lan et al. 2017c) have found significant differences between rubber plantations and tropical forests, specifically that the 87 88 diversity of soil bacteria was higher in rubber plantations than in rainforest. Compared 89 to primary forests, agricultural systems tend to have higher bacterial richness but lower fungal richness (Lan et al., 2017a; Cai et al., 2018; Tripathi et al., 2012; Kerfahi 90 et al., 2016). Compared to Eucalyptus plantations, rubber plantations have been found 91 to possess higher diversity of both bacteria and fungi (Ma et al., 2019). Song et al. 92 (2019) reported that tropical forest conversion to rubber plantation results in reduced 93 fungal microbial community network complexity. However, there are few studies on 94 95 the impact of forest conversion on soil bacterial community network structure, nor on, 96 the drivers leading to changes in network structure. 97 The soil microbial network is viewed as a critical indicator of soil health and quality (Kuperman et al., 2014). Microbial networks consist of two parts: nodes and 98

- 99 edges. Nodes indicate microbe types, generally OTUs or species, but can also indicate
- 100 environmental variables <u>of</u> interest, such as soil pH. Edges (or links) indicate
- significant correlations between nodes. The number of links <u>connected</u> to a node is
- 102 <u>known as</u> the node's degree (Röttjers and Faust, 2018). Usually, the more links <u>a</u>
- 103 network has, the more complex the community is. Similarly, the higher the degree of

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a node (OTU or species), the more important the species is to the network structure 125 (Berry and Widder, 2014). In recent years, microbial network analysis has been used 126 127 to identify keystone taxa (Banerjee et al., 2018) and provide comprehensive insight into the microbial community structure and assembly (Fuhrman, 2019). The 128 129 complexity of network structure is understood by calculating the number of edges, 130 while keystone species of the community are found by estimating the degree of each 131 species. The stability of the microbial community is determined by calculating the 132 ratio of positive and negative correlation of the network because a large proportion of 133 positive correlations in a microbial community are thought to be unstable (Coyte et al., 134 2015). 135 Previous work has shown that agricultural intensification can reduce microbial network complexity (Banerjee et al., 2019). Logging alters soil fungal network in 136 137 tropical rainforests, i.e., a better-organized fungal community in cut stands when 138 compared with primary stands (Chen et al., 2019). Soil bacterial networks are less 139 stable under drought than fungal networks (De Vries et al., 2018). While the exact 140 drivers of microbial network structure still remain unknown. Previous work has 141 shown that soil nutrients, such as soil phosphorus content, and soil pH, are the main drivers for the network structure for microbial communities (Banerjee et al., 2019). 142 Seasonal variation also affects the network structure by changing the identity of 143 keystone species in the community, given that a keystone species might be present 144 only in a specific season or time period (Banerjee et al., 2018). 145 Hainan Island is home to a large area of tropical rainforest rich in biodiversity. It is 146 147 a part of the Indian-Malay rainforest system at the northern edge of the world's rainforest distribution. However, rubber plantations now account for almost a quarter 148

of the total extent of vegetated areas on Hainan Island (Lan et al., 2020a). Previous
 work on the island has shown that forest conversion from rainforest to rubber
 plantations can result in shifts in bacterial composition from the Proteobacteria to

152 Chloroflexi, and fungal composition from Basidiomycota to Ascomycota (Lan et al.,

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167	2020a). Other research has shown microbial (including bacteria and fungi) diversity
168	was higher in rainforest soils than in rubber plantation (Lan et al., 2020a). Therefore,
169	we hypothesize the following; (1) the conversion of rainforest to rubber plantations
170	results in changes in the dominant phyla of soil microbes, with concomitant changes
171	in network structure and identity of keystone species. (2) Soil bacterial-fungal
172	network structure is less complex but more stable in rubber plantation than in
173	rainforest because high species diversity leads to complex network structure and
174	unstable microbiome communities (Coyte et al., 2015). (3) Soil bacterial networks in
175	rubber plantations are less complex than in rainforests because intensive cropping
176	reduces the complexity of bacterial network, even as richness increases (Karimi et al.,
177	2019). (4) Soil fungal network structure in rubber plantation sites is more complex
178	than in the rainforest because organic farming showed a much more complex fungal
179	network than conventional or no-tillage farm systems resulting from organic fertilizer
180	and the return of litter to the field (Banerjee et al., 2019). By testing these hypotheses,
181	we want to clarify the drivers and mechanisms of microbial community assembly that
182	link forest conversion to differences in soil microbial network structure. This study
183	will provide critical information for understanding and managing microbial
184	communities in tropical forests of China and elsewhere.

185 2. Methods

# 186 **2.1 Study site**

This study was conducted on Hainan Island (18°10′–20°10′N and 108°37′–111°03′E), south China. The total area of Hainan Island is about 34,000 km<sup>2</sup> (Lopez et al., 2009). Hainan Island is the largest island within the Indo-Burma Biodiversity Hotspot of, tropical Asia (Francisco-Ortega et al., 2010) and has a tropical monsoon climate. Hainan Island has a warm and humid climate all year round, with an average annual temperature of 22-26°C. Rainfall is abundant, ranging from 1,000 mm to 2,600 mm yearly, with an average annual precipitation of 1,639 mm. The rainy season occurs

194 from May to October, with a total precipitation of about 1500 mm, accounting for

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- 214 70-90% of the total annual precipitation. Only 10-30% of the total annual
- 215 precipitation falls within the dry season, from November to April. The central part of
- 216 Hainan Island is mountainous and contains old-growth tropical rainforests and
- 217 monsoon forests. Rubber plantations are found on the plateaus surrounding the
- 218 central mountainous zone.

# 219 2.2 Soil sampling

The tropical rainforests of Hainan are mainly distributed in Bangwangling, Diaoloshan, 220 Wuzhishan, Yinggeling and Jianfengling. Therefore, we selected a tropical rainforest in 221 each of these sites for our study objectives. There are about 530,000 hectares of rubber 222 plantations on Hainan Island. Five rubber plantations were selected, one each in 223 224 Wanning, Ledong, Danzhou, Haikou, Qiongzhong, These five rubber plantations are located in the east, south, west, north and middle of Hainan, respectively. (Figure S1). 225 We selected 25-30 years old rubber trees (i.e., mature rubber plantations) for our study 226 objectives. Latex harvest and fertilizer application are common management practices 227 in rubber plantations. Usually, compound fertilizer (1-1.5 kg per tree) and organic 228 fertilizers (20-25 kg per tree) were applied once or twice a year. More information on 229 230 the study sites is provided in Table S1. Before soil sampling, we sterilized the soil drill 231 with 75% alcohol. After the removal of the litter layer using a 5-cm diameter steel drill, top soil (0 to 20 cm) was collected, then homogenized and passed through a 2-mm 232 233 mesh sieve. For each site, thirteen soil samples were collected, thus there were a total of 130 samples collected between the rubber plantation, and tropical rainforest sites. 234 Soil sampling was performed twice in 2018, once in January (dry season) and once in 235 July (rainy season). There were thus a total of 260 soil samples (130 per forest type). 236 Soil samples were divided into two parts. One was used to analyze soil water content, 237 238 soil pH, total nitrogen, total phosphorus (TP), total potassium (TK), nitrate nitrogen (NN), ammonium nitrogen (AN), available phosphorus (AP), potassium (AK). Soil 239 water content (%) was measured gravimetrically. Soil pH was measured in a soil/water 240 suspension (1: 2.5, w/w) using a pH meter. TN was determined using a micro-Kjeldahl 241

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- 258 digestion followed by steam distillation. TP and TK were measured following
- 259 digestion with NaOH. NN and AN were determined by steam distillation and
- 260 indophenol-blue colorimetry, respectively. Soil samples were extracted with NaHCO3
- 261 and the extract was then used to measure available soil phosphorus AP via
- 262 molybdate-blue colorimetry. To measure soil AK, soil was extracted with
- 263 ammoniumacetate and then the extract was loaded onto an atomic absorption
- spectrometer with ascorbic acid as a reductant (Chen et al., 2019). Soil properties of the
- rubber plantation and rainforest sites are shown in Table S2. The other part of each soil
- 266 <u>sample</u> was stored in ultra-low temperature (- 80 °C) refrigerator for later DNA
- 267 extraction.

#### 268 2.3 DNA extraction and PCR amplification

- 269 Microbial DNA was extracted from 0.5 g of soil using E.Z.N.A.® Soil DNA Kits
- 270 (Omega Bio-tek, Norcross, GA, U.S.) following the manufacturer's protocol. The
- 271 fungal ITS1 hypervariable region was amplified using the PCR primers ITS1F
- 272 (5'-CTTGGTCATTTAGAGGAAGTAA-3') and ITS2R
- 273 (5'-GCTGCGTTCTTCATCGATGC-3') (Adams et al., 2013). For bacteria and
- archaea, the V4 hypervariable region of the bacterial 16S rRNA gene was amplified
- using the PCR primers 515FmodF (5'-GTGYCAGCMGCCGCGGTAA-3') and
- 276 806RmodR (5'-GGACTACNVGGGTWTCTAAT-3') (Walters et al., 2016; Sampson
- et al., 2016). The PCR reactions were conducted using the following approach: an
- initial 3 min denaturation at 95°C, followed by 27 cycles of 30s at 95°C, 30s of
- annealing at 55°C, and 45s of elongation at 72°C, and a 10 min final extension at
- 280 72°C.

### 281 2.4 Illumina MiSeq sequencing

- Amplicons were extracted from 2% agarose gels, purified using the AxyPrep DNA
- 283 Gel Extraction Kit (Axygen Biosciences, Union City, CA, U.S.) and quantified using
- a QuantiFluor<sup>TM</sup> -ST Fluorometer (Promega, U.S.). Purified amplicons were pooled in

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- an equimolar solution and then sequenced (paired-end,  $2 \times 250$  bp) on an Illumina
- 289 MiSeq platform according to standard protocols.
- 290 Metagenomic shotgun sequencing libraries were prepared and then sequenced by
- 291 Majorbio, Inc. (Shanghai, China) using the Illumina HiSeq 2000 platform. The NR
- 292 gene catalog was aligned against the Kyoto Encyclopedia of Genes and Genomes
- 293 (KEGG) database using BLAST (Version 2.2.28+) and then functionally annotated
- using KOBAAS 2.0 according to previously described methods (Qin et al., 2010)

## 295 2.5 Bioinformatics and data analysis

- 296 Raw fastq files were demultiplexed and quality-filtered using QIIME (Caporaso et al.,
- 2010) (version 1.17). Operational Taxonomic Units (OTUs) were clustered with a
- 298 97% similarity cut-off using UPARSE (Edgar, 2013), and chimeric sequences were
- 299 identified and removed using UCHIME. Using the RDP Classifier, the phylogenetic
- affiliation of each 16S rRNA gene and ITS gene sequence was determined using a
- 301 confidence threshold of 70% with the SILVA 16S rRNA database and UNITE
- database, respectively (Amato et al., 2013). For each site, the relative abundance of
- 303 different taxa (Good, 1953) and the Shannon diversity index were calculated based on
- 304 re-sampled sequence data using MOTHUR (Schloss et al., 2009). The raw reads were
- 305 deposited into the NCBI Sequence Read Archive (SRA) database (Accession Number:
- 306 SRP108394, SRP278296, SRP278319).

# 307 2.6 Statistical analysis

- 308 <u>To make the co-occurrence network analyses simple, we only used</u> the top 300 most
- 309 abundant bacterial and fungal OTUs to analyze the network structure of the two forest
- 310 types. These OTUs are roughly match the OTUs with a relative abundance greater
- than 0.05% (Jiao et al., 2016). Four networks, representing each habitat during each
- sampling period (rainforest and rubber plantations in <u>the</u> dry and rainy seasons), were
- 313 constructed with 65 samples each. We also performed the bacterial-fungal community
- network analysis to investigate soil microbial network complexity of <u>each of the four</u>.

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324 Interactions consisted of Spearman's rank correlations, Co-occurrence networks were constructed using only significant correlations of  $\rho > 0.6$  (P < 0.01) (Barberan et al. 325 2012), because this cutoff includes a range of interactions strengths (De Vries et al., 326 2018). The networks were then visualized in R using the igraph package. To reveal 327 the distribution pattern of correlation coefficients, the frequency of the coefficients of 328  $\rho > 0.3$  (P < 0.01) were plotted. The Network Analyzer tool in Cytoscape (version 329 330 3.4.0) was used to calculate network topology parameters including number of nodes, 331 edges, degree, betweenness, closeness. The number of shared edges and unique edges 332 as well as keystone OTUs were calculated to evaluate whether the network structure 333 changed. We defined a shared network edge as the link (edge) between species A and species B not only existing in rubber plantation network, but also in rainforest 334 335 network. Similarly, a unique edge was a link that only existing in either rubber plantations or rainforests. The number of shared edges and unique edges were 336 calculated using igraph and visualized using Venn diagrams. Keystone OTUs are 337 known to be important for ecosystem structure and function and were selected on the 338 339 basis of high network degree, high closeness centrality, and low betweenness centrality as defined by Berry and Widder (2014). To evaluate the proportional 340 341 influence of each phylum on bacterial and fungal network structure, node degrees of each phylum were calculated and bar plots were created. Correlation coefficients 342 between species and functions were calculated based on metagenomics data. Here we 343 used the top 50 most abundant species and top 50 KEGG functions (pathway level 3). 344 345 Then the species and function correlation network was constructed on the Major bio cloud platform (https://cloud.majorbio.com/). To reveal the relationship between 346 347 microbial taxa and environment variables, two-way correlation networks were also constructed on the Major bio cloud platform. The topological role of each node in a 348 network was assessed by the  $Z_i$  and  $P_i$  values, where  $Z_i$  represents the connectivity 349 350 of node *i*, within a module, and  $P_i$  measures the degree of node *j* connected with other modules (Roger and Amaral, 2005). All species can be divided into four groups 351 according to the simplified criteria (Olesen et al., 2007), namely peripherals ( $Z_i < 2.5$ 352

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and  $P_i < 0.62$ ), connectors ( $P_i > 0.62$ ), module hubs ( $Z_i > 2.5$ ) and network hubs ( $Z_i >$ 

369 2.5 and  $P_i > 0.62$ ). The  $Z_i$  and  $P_i$  values were calculated using GIANT package of

- 370 Cytoscape. The  $Z_i$ - $P_i$  plot was created with ggplot2 in R.
- 371 **3. Results**

# 372 **3.1 Bacterial and fungal networks**

- 373 Our results showed that the bacterial-fungal community network were more complex
- in rubber plantations (4,284 edges in <u>the</u> dry season, 7,257 in <u>the</u> rainy season) than in
- 375 primary forests (3,650 in the dry season, 6,507 in the rainy season), and more
- complex <u>during the rainy season than the dry season (Figure 1)</u>. Most nodes of
- 377 bacterial networks (Figure 2) and fungal networks (Figure 3) varied with forest type
- in both the dry and rainy seasons. For bacterial networks, there were 2,559 and 2,501
- arges in the forests and rubber plantations, resepectively, in the dry season, but these
- two networks only shared 262 edges, accounting <u>for</u> only about 10% the total edges
- 381 (Figure 4A-B). Similarly, these networks only shared 519 edges during the rainy
- season. For fungal networks, there were only 4 and 5 shared edges (i.e., no more than
- 5% of the total edges) in <u>the dry and rainy seasons</u>, respectively (Figure 4C-D).

The number of edges of bacterial and fungal networks were almost equivalent 384 385 during the dry season. However, in the rainy season, there were more edges in the 386 rainforest bacterial network than in the rubber plantation network (Table 1 & Figure 4B). For the network structure of the fungal community, more edges were observed in 387 388 rubber plantations in the rainy season (Table 1 & Figure 4D). Similarly, there were no significant differences in either the degree of the bacterial and fungal networks 389 between the tropical rainforest sites and rubber plantations in the dry season (Figure 390 5A, C). In the rainy season, rainforest sites had a higher bacterial network degree, 391 392 while rubber plantations showed higher fungal network degree (Figure 5B, D). For bacterial networks, more nodes (OTUs) with high degree (rubber plantation had 2 393 394 nodes with degree greater than 75, rainforest had 8 such nodes) were observed in the

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408 rainforest during the rainy season (Figure S2B). For fungal networks, 15 nodes of 409 higher degree (degree greater than 25) were observed in rubber plantations, however, 410 only 7 such nodes existed for rainforest sites (Figure S2D). These results indicate that 411 the fungal network structure in the rubber plantations was more complex than for 412 tropical rainforests during the rainy season, but that the reverse was true for bacteria. 413 When considering the ratio of positive to negative correlation coefficients, the 414 bacterial-fungal community network revealed that rubber plantations (844 edges in dry season, 1744 edges in rainy season) have more negative links than rainforest (149 415 416 edges in dry season, 489 edges in rainy season) indicating that the network of rubber 417 plantation was more stable than that of the rainforests (Figure 1). More correlations 418 (greater than 0.3, P < 0.01) were also calculated to investigate the ratio of positive to negative correlation and the results showed that negative correlations between 419 420 bacterial and fungal OTUs in the rubber plantations were consistently stronger than in 421 the tropical rainforests in both dry season and rainy season (Figure 6). 422 3.2 Keystone taxa 423 Module hubs and network hubs have been commonly identified as keystone taxa in 424 networks. However, for both the bacterial and fungal communities, neither tropical

- 425 rainforest nor rubber plantation networks possessed module hubs and network hubs
- 426 (Figure S3-S4). The keystone taxa of bacteria were very different between rubber
- 427 plantations and tropical rainforest sites in both the dry and rainy seasons. For bacteria,
- 428 there were more keystone taxa in tropical rainforest sites than in the rubber
- 429 plantations in both the dry and rainy seasons, indicating that the tropical rainforest
- 430 networks had higher complexity. We found that some groups of Acidobacteria are
- 431 keystone taxa in tropical rainforest sites but disappeared after forest conversion. There
- 432 were more Actinobacteria in rubber plantations than in tropical rainforest sites (Table
- 433 S3).
- 434 For fungi, more keystone taxa were observed in rubber plantations than in tropical

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rainforest sites during both the dry and rainy seasons, indicating that the rubber
plantation networks were more complex. Most keystone taxa belong to Ascomycota,
suggesting that member of this group are very important for network structure. In
addition to forest conversion, seasonal changes also affect the keystone taxa of the
fungal community network. There were more Basidiomycota OTUs in the dry season,
but more Ascomycota in rainy season (Table S4).

448 **3.3 <u>Drivers</u> of microbial composition and network structure** 

Two-way correlation network analysis of the 300 most abundant OTUs and 449 environmental variables revealed elevation, (106 degree), AK (104), and soil pH (86) 450 were the main drivers of bacterial community composition (Figure S5A). For fungi, 451 452 elevation had the highest degree (61), followed by AK (51) (Figure S5B). The 453 relationship between OTU abundance and soil pH revealed that the soil pH negatively correlated with some species of Acidobacteria, but positively correlated with some 454 species of Chloroflexi and some species of Ascomycota (Figure S6). AK positively 455 correlated with species of Planctomycetes, Verrucomicrobia and Basidiomycota, but 456 negatively correlated with Chloroflexi and Ascomycota. In words, AK and soil pH 457 458 were important driver for microbial composition.

459 The total degree of each phylum were used to reveal the influence of each phylum 460 on network structure (Figure 7). For bacteria, Proteobacteria, Actinobacteria and Acidobacteria had a large influence on network structure. Acidobacteria and 461 462 Planctomycetes contributed more to rainforest networks than rubber plantation 463 networks. However, Actinobacteria and Chloroflexi showed the opposite pattern. For 464 fungi, Ascomycota and Basidiomycota had large influences on network structure. 465 Both Ascomycota and Basidiomycota had stronger influences on rainforest networks than rubber plantation networks. The influence of Ascomycota was stronger during 466 the rainy season than in the dry season, indicating seasonal change also had impact on 467 fungal community networks. The degree of each phylum of bacterial and fungal 468 network also revealed seasonality had effects on network degree (Figure S7). 469

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### 480 **3.4 Interaction between species and functions**

- 481 Two-way network analysis of the 50 most abundant species (all belonging to bacteria
- 482 groups based on metagenomic data) and the 50 most abundant KEGG functions
- 483 revealed the interaction between microbial composition and functions. The results
- 484 showed that soil network structure at rainforests sites (460 edges in the dry season,
- 485 1,750 in the rainy season) was more complex than that of rubber plantations (223 in the
- 486 dry season, 451 in the rainy season) (Figure 8), indicating that more links between
- 487 microbial species and functions were observed in soils of rainforest sites. Both rubber
- 488 plantation, and rainforest networks were more complex in the rainy season than in the
- 489 dry season. We also found that metabolism was the most important function in soil
- 490 microbial networks. Surprisingly, species of Actinobacteria negatively correlated with
- 491 function in rubber plantations.

## 492 4. Discussion

493 4.1 Forest conversion reduces soil bacterial network complexity

494 Land-use changes increasingly threaten biodiversity, particularly in tropical forests

- 495 (Gibson et al., 2011). However, we still have little understanding of how soil
- 496 networks respond to forest conversion, such when rainforests are converted to rubber
- 497 plantations. Our results showed that forest conversion had large effects on both soil
- 498 bacterial and fungal networks. <u>Tropical rainforest bacterial networks had more edges</u>
- (Table 1) and <u>a higher degree (Figure 4) than did rubber plantation networks</u>,
- 500 especially during the rainy season, indicating that the rainforest bacterial network was
- 501 more complex than the rubber plantation network. This <u>finding is</u> consistent with
- 502 previous observations that soil bacterial networks were more complex in natural
- 503 systems than in crop soil (Karimi et al., 2019). Further study showed that soil
- 504 networks become more connected <u>through the process of ecological</u> restoration
- 505 (Morriën et al., 2017). The observed decrease in network complexity and cohesion
- supports the hypothesis that cropping may enhance the isolation of bacterial taxa

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519 (Karimi et al., 2019), which results in lower numbers of connections in the network. 520 In addition, at the microscale, the structure of tilled soils is more homogeneous, and 521 the soil pores are less connected than in soils that are not tilled (Pagliai et al., 2004), 522 such as rainforest soil. In nature, soil ecosystems are highly heterogeneous since hot 523 spots of soil microbial biodiversity can form spatial and temporally within soil aggregates (Bach et al., 2018). This spatial heterogeneity likely plays an important 524 525 role for the interactions among microbes and the mechanisms by which more complex and diverse communities drive various nutrient cycling processes on small spatial 526 527 scales (Wagg et al., 2018).

528 A large number of studies employing microbial network analysis have enriched

- 529 our understanding of microbial co-occurrence patterns in various soil ecosystems.
- 530 However, very little is known of whether differences in the structure of microbial
- networks have consequences for microbiome functioning (Wagg et al., 2018). Our
- results demonstrated that more species <u>cor</u>related with metabolism in natural system
- than in the agricultural system, especially during the rainy season. This is in line with
- a previous study conducted in Sumatra, Indonesia, which found that the
- transformation of forest to rubber results in a 10-16% decrease in community
- 536 metabolism (Barnes et al., 2014). Fewer interactions between microbial species (most
- 537 of them are bacteria) and functions in rubber plantations demonstrated that forest
- 538 conversion reduced soil bacterial network complexity.
- 539 4.2 Forest conversion increased soil fungal network complexity
- 540 Surprisingly, rainforest bacterial networks were characterized by fewer edges (Table 1)
- 541 and <u>a lower degree (Figure 4) than the rainforest bacterial networks, meaning</u> that
- rubber plantation bacterial networks were more complex than those of native forests.
- 543 Our results were not consistent with a study conducted in Xishuangbanna (Song et al.,
- 544 2019) which showed that tropical forest conversion reduced fungal network complex,
- 545 but <u>is consistent with previous observations which found that fungal community</u>
- 546 networks were better organized in disturbed forest compared to primary forest (Chen

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555 et al., 2019). Banerjee et al. (2019)'s observation showed that organic agricultural 556 fields harbored much more complex fungal networks with many more keystone taxa 557 than conventionally managed fields. Forest conversion resulted in shifts in fungal composition from Basidiomycota to Ascomycota (Figure S8), as seen in previous 558 559 investigations (Lan et al., 2020a; Lan et al., 2020b). Previous work showed that Basidiomycota species show higher drought sensitivity than Ascomycota species 560 561 (Taniguchi et al., 2018), this would result in a shift in richness and abundance of Basidiomycota species (Figure S8). Many Basidiomycota species are capable of 562 563 long-distance dispersal (Egidi et al. 2019, Geml et al., 2012), which may result in a 564 decrease in fungal network complexity. This possibly explained why Ascomycota 565 OTUs contribute more to the network structure than Basidiomycota (Figure 6). 566 Overall, reductions in the abundance and richness of Basidiomycota species led to an increase in fungal links in rubber plantations. 567

568 4.3 Forest conversion enhanced the stability of soil network

The ability of a microbial community to resist and recover from disturbances is 569 570 defined as the stability of the community, (De Vries and Ashley, 2013). The ratio of 571 positive to negative network links constitutes the balance between facilitative and inhibitive relationships within a network (Karimi et al., 2017). Theoretical studies, for 572 example, predict that ecological networks that consist of weak interactions are more 573 stable than those with strong interactions (Neutel et al., 2002, Coyte et al., 2015), and 574 that compartmentalization and presence of negative interactions increase the stability 575 of networks under disturbances (Coyte et al., 2015, Rooney et al., 2006, Stouffer & 576 Bascompte 2011). In our study, more negative correlations were detected in rubber 577 plantation, indicating the network structure of rubber plantation soils was more stable 578 579 than that of rainforest soils (De Vries et al., 2018). Forest conversion from rainforest to rubber plantations resulted in shifts in bacterial composition from Proteobacteria to 580 Chloroflexi, and fungal composition from Basidiomycota to Ascomycota (Lan et al., 581

582 2020a). Proteobacteria and Basidiomycota <u>are</u> copiotrophic microorganisms, while

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596	4.4 Main drivers for soil microbial composition and network structure	
595	response of microorganisms to disturbance are very complex.	
594	is not the only mechanism that could maintain community stability because the	l
593	quick-growing and less resistant but more resilient to disturbances. However, growth	
592	more k-strategists to more r-strategists (De Vries et al., 2013). R-strategists are	[
591	This indicated forest conversion resulted in shifts in soil microbial composition from	1
590	Chloroflexi and Ascomycota are oligotrophic, (Yang et al., 2019; Lan et al., 2020a).	1

Forest conversion results in the loss of plant diversity, plant biomass and an increase 597

- in soil pH (Lan et al., 2017a, 2017b). Rubber plantations had a significantly higher pH, 598
- which explains the relative decrease in the abundance of Acidobacteria (Lan et al., 599
- 600 2017a). Our results demonstrate that forest conversion drives a shift in which soil
- 601 microbes act as keystone species (Table 1). We found that many OTUs of
- Acidobacteria fit our criteria as keystone species for rainforest sites, which is 602
- consistent with previous findings (Banerjee et al., 2018). Unexpectedly, OTU11388 603
- 604 and OTU11373, both Acidobacteria, were observed in rainforest soils in both the dry
- and rainy seasons, indicating Acidobacteria were very important for rainforest soil 605
- 606 bacterial networks (Figure 6 and table S3). Higher AK concentrations resulted in a
- 607 higher abundance and more taxa of Actinobacteria (Figure 9), which suggests that
- Actinobacteria contributed more in rubber plantation than in rainforest (Figure 6). 608
- 609 Indeed, forest conversion reduced the abundance of Actinobacteria OTUs (Figure S9).
- Due to the human disturbance in rubber plantations, the soil will inevitably be slightly 610
- polluted with herbicides and domestic garbage. Previous study showed member of 611
- Actinobacteria were observed in contaminated soil (Jiao et al., 2016). 612
- 613 Forest conversion also increases land use intensity (Brinkmann et al., 2019),
- 614 including the application of fertilizer and herbicide. The application of herbicides has
- also been linked to significant decreases in root colonization and spore biomass of 615
- arbuscular mycorrhizal fungi in tropical agriculture (Zaller et al., 2014). Soil nutrient 616
- 617 concentrations show, a decline around the roots of rubber plantations compared to

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630 those from rainforests (Sahner et al., 2015). Our study was no exception. For instance, 631 AK and TN concentration were significant lower in rubber plantations than in 632 samples from rainforest sites (Table S2). Higher concentrations of AK reasonably explained the higher contribution of Basidiomycota to the network structure (Figure 633 634 8B) due to the positive association of AK with Basidiomycota.

635 Spatiotemporal heterogeneity can be a major driver of the abundance and 636 distribution of keystone taxa in soil, which is a highly heterogeneous and multifaceted environment (Mills et al., 1993, Power et al., 1996; Mouquet et al., 2013). Seasonal 637 variability determines the structural and compositional properties of microbiomes in 638 an environment, and as such, a keystone species might be present only in a specific 639 640 season or time period (Banerjee et al., 2018). It was interesting that more bacterial 641 OTUs were identified as connectors during the rainy season than the dry season. 642 Connectors have been characterized as generalists (Olesen et al., 2007), and 643 generalists drive covariation among communities in a network (Chen et al., 2019). Previous observations have demonstrated that some keystone taxa that were found in 644 the dry season disappeared during the rainy season (Lan et al., 2018). Seasonal 645 changes possibly explain why some keystone taxa were observed in the rainy season 646 647 but not in the dry season. 4.5 Possible impact of forest conversion on microbial community function 648 Forest conversion results in a decrease in the abundance of Proteobacteria and an 649 650 increase of Actinobacteria. Most species of Proteobacteria were positively correlated with metabolic function, while most of Actinobacteria, were negatively associated with 651 652 metabolic function (Figure 8). Therefore, the changes in the abundance of these two phyla results in a reduction of microbial community function after forest conversion. 653

- Because the metabolic function of a specific species is often affected by environmental 654
- conditions (Louca et al., 2018), some species are not correlated with any function in the 655
- dry season, but correlated with metabolic function in the rainy season, indicating there 656
- some microorganisms do not participate in the metabolic process during the dry season, 657

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672 especially for the rubber plantations. In conclusion, forest conversion as well as

673 seasonal variation had a great impact on soil microbial community functions.

## 674 5. Summary

- 675 Our knowledge about land-use impacts on soil ecosystems is mostly limited to
- biodiversity and ecosystem functions, leaving uncertainty about how soil networks
- 677 change after forest conversion. This study is the most comprehensive report on
- 678 changes in network structure that occur when tropical rainforests are converted into
- 679 rubber <u>plantations</u>. Our study showed that forest conversion altered both bacterial and
- 680 fungal soil networks, reduced bacterial network complexity and enhanced fungal
- network complexity, especially during the rainy season. One possible reason maybe
- 682 that forest conversion <u>altered</u> soil pH and other soil properties, which altered bacterial
- 683 composition, subsequent network structure as well as <u>certain soil ecosystem</u> functions.
- 684 Our study demonstrates the impact of forest conversion on soil network structure,
- 685 which has important implications for ecosystem functioning, such as metabolic
- 686 <u>activity</u>, and health of soil ecosystems in tropical regions.
- 687

# 688 Code availability

- 689 Not applicable
- 690 Authors' contributions
- 691 Guoyu Lan: Conceptualization, Methodology, Writing, Reviewing and Editing;
- 692 Chuan Yang, Rui Sun, Bangqian Chen, Xicai Zhang and Zhixiang Wu: Investigation

# 693 Competing interests

- 694 The authors declared that they have no conflicts of interest to this study.
- 695 Disclaimer
- 696 Publisher's note: Copernicus Publications remains neutral with regard to jurisdictional

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703 claims in published maps and institutional affiliations.

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## 911 Figure legend

- 912 Figure 1 The soil bacterial-fungal community network of rubber plantations and
- 913 rainforests in the dry and rainy seasons. Red solid circles indicate fungi, and green
- solid circles indicate bacteria. Red lines indicate positive correlation between OTUs,
- and green indicate negative correlation. Absolute value of correlation coefficient  $\rho$  >
- 916 0.6, P < 0.01
- 917 **Figure 2** Soil bacterial network structure of rubber plantations and tropical rainforests
- 918 in dry and rainy seasons. Red lines indicate positive correlation between OTUs, and
- green indicate negative correlation. Absolute value of correlation coefficient  $\rho > 0.6$ ,
- 920 P < 0.01
- 921 Figure 3 Soil fungal network structure of rubber plantations and tropical rainforests
- 922 in <u>the dry and rainy seasons</u>. Red lines indicate positive correlation between OTUs,

923 and green indicate negative correlation. Absolute value of correlation coefficient  $\rho >$ 924 0.6, P < 0.01

- **Figure 4** Number of shared and unique edges of soil bacterial and fungal networks in
- 926 rubber plantations and tropical rainforests in the dry and rainy seasons. The number
- 927 where the two circles cross is number of shared edges. Numbers in <u>the</u> red circle
- presents the unique edges in rainforests, while in blue circle present the unique edgesin rubber plantations.
- Figure 5 Network degree of soil bacterial and fungal communities of rubber
  plantations (blue) and tropical rainforests (red) in the dry and rainy seasons.
- **Figure 6** Frequency distributions of correlation\_strength in bacterial (a: dry season, single season) and fungal (c: dry season, d: rainy season) networks of rubb
- plantations and tropical rainforests in the dry and rainy seasons. (absolute correlation coefficient greater than 0.3, P < 0.01) Correlations in rainforest networks are red correlations in rubber plantation networks are blue.
- Figure 7 Proportional influence of different <u>phyla on bacterial and fungal netwo</u>
  structure in both dry and rainy seasons. The influence was the number of degrees
- nodes belonging to a particular phylum. (a: bacteria in dry season, b: bacteria in rseason, c: fungal in dry season, d: fungal in rainy season).
- 941 Figure 8 Network of the top 50 most abundant species (based on metagenomics data)
- and top 50 most frequent KEGG functions (pathway level 3) of rubber plantations and
- 943 tropical rainforest sites in the dry and rainy seasons. (A: rubber in dry season; I
- 944 rainforest in dry season; C: rubber in rainy season; D: rainforest in rainy season) The
- size of the node indicates the species/function abundance. A red line indicates positive
- 946 correlation between species/functions, and green indicates negative correlation.
- 947 Absolute value of correlation coefficient > 0.6, P < 0.01

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# Table 1 Topological properties of soil microbial (bacterial and fungi) network

structure in rubber plantation and tropical rain forest in dry and rainy seasons

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	Bacteria				Fungi			
	Rubber	Rainforest	Rubber	Rainfores	t Rubber	Rainfores	t Rubber	Rainforest
	Dry	Dry	Rainy	Rainy	Dry	Dry	Rainy	Rainy
No. of nodes	291	287	296	296	220	235	243	244
No. of edges	2448	2559	4248	5019	791	769	1250	905
No. of positive edges	2052	2508.00	3385	4901	760	764	1195	897
No. of negative edges	396	51	863	118.00	31	5	55	8
Connectance	0.06	0.06	0.09	0.11	0.02	0.02	0.03	0.02
Average degree	16.82	17.83	16.67	33.91	16.67	6.54	10.28	7.41
Average betweenness	100.61	93.52	31.23	42.57	208.90	185.03	170.99	266.31
Average of shortest path length	2.92	2.92	2.56	2.45	4.00	3.626	3.81	5.27
Diameter	6.00	7.00	7.00	6.00	10.00	12.00	9.00	15.00
Cluster of coefficient	0.46	0.451	0.51	0.54	0.59	0.457	0.49	0.51
No of clusters	11.00	14.00	5.00	5.00	94.00	75.00	62.00	67.00
Degree centralization	0.12	0.11	0.17	0.18	0.08	0.08	0.10	0.10
Betweenness centralization.	0.0066	0.0067	0.0053	0.0049	0.02	0.03	0.121	0.04
Closeness centralization.	0.35	0.35	0.40	0.41	0.31	0.351	0.291	0.26
Neighborhood Connectivity	21.12	21.79	34.69	40.68	10.45	9.28	13.78	9.44
Topological coefficient	0.26	0.26	0.26	0.27	0.33	0.36	0.37	0.50

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Figure 1 The soil bacterial-fungal community network of rubber plantations and rainforests in the
 dry and rainy seasons. Red solid circles indicate fungi, and green solid circles indicate bacteria. Red

967 lines indicate positive correlation between OTUs, and green indicate negative correlation. Absolute

968 value of correlation coefficient  $\rho > 0.6, P < 0.01$ .

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972 Figure 2 Soil bacterial network structure of rubber plantations and tropical rainforest in the dry

- 973 and rainy seasons. Red line indicates positive correlation between OTUs, and green indicates
- 974 negative correlation. Absolute value of correlation coefficient  $\rho > 0.6, P < 0.01$



**Figure 3** Soil fungal network structure of rubber plantations and tropical rainforests in the dry and

978 rainy season<u>s</u>. Red line indicates positive correlation between OTUs, and green indicates negative

- 979 correlation. Absolute value of correlation coefficient  $\rho > 0.6, P < 0.01$



982 Figure 4 Number of shared and unique edges of soil bacterial and fungal networks in rubber

plantations and tropical rainforests in the dry and rainy seasons. The number where the two circles

cross is number of shared edges. Numbers in <u>the</u> red circle presents the unique edges in rainforest<u>s</u>,
while in blue circle present the unique edges in rubber plantations.







Figure 6 Frequency distributions of correlation strength in bacterial (a: dry season, b: rainy season) and fungal (c: dry season, d: rainy season) networks of rubber plantations and tropical rainforests in the dry and rainy seasons (absolute correlation coefficient greater than 0.3, P < 0.01). 999 Correlations in rainforest networks are red, correlations in rubber plantation networks are blue. 1000

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- 1007 Figure 7 Proportional influence of different phyla on bacterial and fungal network structure in
- both dry season and rainy season. The influence was the number of degrees of nodes belonging to
- 1009 a particular phylum. (a: bacteria in dry season, b: bacteria in rainy season, c: fungal in dry season,
- 1010 d: fungal in rainy season).
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- 1015 Figure 8 Network of the top 50 most\_abundant species (based on metagenomics data) and the top
- 1016 50 abundant KEGG functions (pathway level 3) of rubber plantations and tropical rainforest in the

1017 dry and rainy seasons (A: rubber in dry season; B: rainforest in dry season; C: rubber in rainy 1018

season; D: rainforest in rainy season). The size of the node indicates the species/function

- 1019 abundance. Red line indicates positive correlation between species/functions, and green indicates
- 1020 negative correlation. Absolute value of correlation coefficient > 0.6, P < 0.01

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