



1 **Stronger microbial nutrient limitations in subsoil along the precipitation gradient of**
2 **agroecosystem: Insights from soil enzyme activity and stoichiometry**

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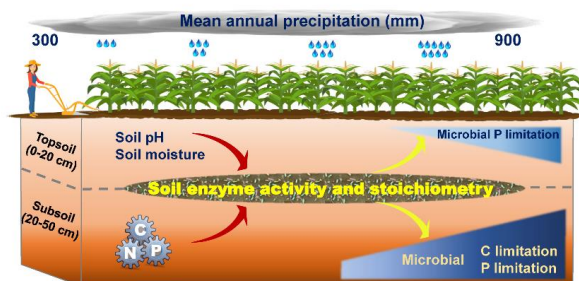
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21 **Abstract:** Soil extracellular enzymes are central in terrestrial ecosystem responses to climate
22 change, and their research can be crucial for assessing microbial nutrient demand. However,
23 the effects of climate-induced precipitation patterns on soil microbial nutrient demand in
24 different soil profiles of agroecosystems are rarely studied. Here, we present how the
25 precipitation gradient affects soil enzymes related to carbon (C), nitrogen (N) and phosphorus
26 (P) cycling and identified microbial nutrient limitation determinants at five depth intervals (0-
27 10, 10-20, 20-30, 30-40 and 40-50 cm) in seven agroecosystems. We found that N- and P-
28 acquiring enzymes have a tendency to increase or decrease, but C- acquiring enzymes did not
29 change along the precipitation gradient throughout soil profiles. Soil pH and moisture were the
30 most important factors affecting the enzyme activity in 0-50 cm. Our results also revealed a
31 crucial soil boundary (at 20 cm) that differentiated responses of microbial nutrient limitation
32 to precipitation changes. In the topsoil (0-20 cm), the stoichiometry of soil nutrients did not
33 vary with precipitation. Microbial P limitation was exacerbated with increased precipitation,
34 which was controlled by soil pH and moisture in the topsoil. In contrast, in the subsoil (20-50
35 cm), soil nutrient stoichiometry decreased with increasing precipitation, and microbial C and
36 P limitation displayed a positive correlation with precipitation. Furthermore, microbial P
37 limitation tended to be stronger in the subsoil than in the topsoil along the precipitation gradient.
38 Microbial C and P limitation was regulated by the soil nutrients and their stoichiometry in the
39 subsoil. Our study is an essential step in soil enzyme activity and stoichiometry response to
40 precipitation in agroecosystems and provides novel insights into understanding microbial
41 nutrient limitation mechanisms in soil profiles along the precipitation gradient.
42 **Keywords:** cropland, ecological stoichiometry, extracellular enzyme, nutrient limitation,



43 Northeast China Transect



44

45 **Graphical abstract** Soil enzyme activity and stoichiometry along a precipitation gradient
46 (300-900 mm) among soil profiles in agricultural ecosystems. Microbial C limitation,
47 microbial carbon limitation; Microbial P limitation, microbial phosphorus limitation.

48 **1. Introduction**

49 Changes in precipitation can lead to shifts in belowground ecological processes that are
50 interlinked with primary production, damaging regional crop production in agroecosystems
51 (Franco et al., 2019; Lesk et al., 2016; Sünnemann et al., 2021). Under rapid precipitation
52 changes, carbon (C), nitrogen (N), and phosphorus (P) cycles may become uncoupled because
53 of the different degrees of control exerted on the supply of these elements by biological and
54 geochemical processes (Delgado-Baquerizo et al., 2013). To meet the food demand, farmers
55 are continuously raising N and P inputs, altering C availability, and leading to a further
56 imbalance in the stoichiometry of microbial substrates in soils (Peñuelas et al., 2013; Schleuss
57 et al., 2021). A combination of precipitation and agricultural activities more drastically alters
58 the patterns, magnitude, and extent of nutrient limitation in soil (Marklein and Houlton, 2012).
59 Such limitation is directly relevant to the fate of soil microbial processes (Camenzind et al.,
60 2018), but the extent and nature of microbial nutrient limitation has never been systematically



61 assessed in agroecosystems along a precipitation gradient. Thus, the underlying ecological
62 mechanisms governing microbial nutrient limitation in agroecosystems are required to support
63 future predictions of precipitation changes.

64 Extracellular enzyme activities (EEA) play a central role in regulating imbalances
65 between resources and growth demands in microbe self-organizational processes (Sinsabaugh
66 and Follstad Shah, 2011; Sinsabaugh et al., 2015). Extracellular enzymes are mediators of
67 decomposition and mineralization of N and P (Moorhead and Sinsabaugh, 2006; Schimel and
68 Weintraub, 2003), and are also indicators of microbial nutrient demand (Moorhead et al., 2016).
69 According to resource allocation theory, when a resource is scarce, microorganisms
70 preferentially invest metabolic resources into extracellular enzyme production to mine the
71 limited element (Allison and Vitousek, 2005; Johnson et al., 2010; Mooshammer et al., 2014).
72 Due to the context dependence of enzyme activity, EEA has gradually received greater
73 attention and become a popular method for examining the direct effects of precipitation on the
74 functional potential of the soil microbial community in field experiments (Henry, 2013;
75 Weedon et al., 2011).

76 The theory of enzyme stoichiometry further reflects the balance between microbial
77 nutrient demand and soil nutrient resource supplies, highlighting the limiting factors of
78 biogeochemical cycling in terrestrial ecosystems (Sinsabaugh et al., 2009; Xu et al., 2017;
79 Zechmeister-Boltenstern et al., 2015). Sinsabaugh et al. (2008) developed an approach to
80 visualize the relative C, N, and P controls on soil microbial communities by plotting the ratios
81 of activities for enzymes associated with C, N and P acquisition. Moorhead et al. (2016)
82 proposed that vector lengths and angles can be used to improve the quantification and



83 visualization of these controls. A global meta-analysis exhibited that enzymatic C:P and N:P
84 ratios were inversely related to precipitation, while the ratio of C:N acquisition increased with
85 precipitation (Sinsabaugh et al., 2008). In forest ecosystems, the C:N, C:P, and N:P acquisition
86 ratios were negatively related to precipitation from a depth of 0-10 cm soil (Xu et al., 2017).
87 However, Cui et al. (2019a) found a shift from P to N limitation of microorganisms with an
88 increase in precipitation from 300 mm to 900 mm in grassland ecosystems from the 0-20 cm
89 soil layer. Therefore, generalizing the effects of precipitation on soil EEA can be challenging
90 and elusive, and it is necessary to strengthen our research on EEA patterns at regional scales.
91 To the best of our knowledge, there remains a lack of studies on the impact of precipitation
92 gradients on soil enzyme in agroecosystems, especially in deeper soil profiles.

93 The enzyme distribution in the soil matrix reflects the dynamic interactions among the
94 soil structure, environmental conditions, and microbial nutrient limitation (Allison et al., 2007).
95 Microbial nutrient limitation has been studied in soil profiles (Brockett et al., 2012; Jing et al.,
96 2017; Peng and Wang, 2016; Zuo et al., 2018). Changes in abiotic environments due to long-
97 term climate drivers, such as alterations in soil moisture (Morugán-Coronado et al., 2019), soil
98 pH (Kivlin and Treseder, 2014; Rousk et al., 2010), soil C:N ratio (Zhou et al., 2020) and soil
99 type (Wakelin et al., 2008), can affect microbial nutrient limitation in complex and
100 unpredictable ways (Burns et al., 2013; Li et al., 2018; Xu et al., 2020a). These changes often
101 occur in strong gradients with depth (Dove et al., 2020; Jia et al., 2019; Stone et al., 2014).
102 Especially in agroecosystems, cropland soils have separate soil layers due to different
103 intensities of human disturbance: topsoil (0-20 cm) and subsoil (below 20 cm) (Yan et al.,
104 2017). It is therefore not surprising that most soil properties (e.g., moisture and nutrients) and



105 microbial communities generally vary with soil depth (Kanal and Kölli, 1996; Li et al., 2014).
106 More importantly, emerging evidence shows that subsoil may exhibit a stronger response to
107 climate change in microbial activities and functioning (Fontaine et al., 2007) in association
108 with microbial nutrient limitation. However, it remains unclear whether the relationships
109 between microbial nutrient limitation and precipitation in topsoil and subsoil differ at the
110 regional scale. Identifying the mechanisms that govern these divergent responses of microbial
111 nutrient limitation to precipitation in soil profiles is critical to minimize uncertainty in future
112 soil health and function.

113 In this study, we focused on soil enzyme activity in five soil layers of seven maize fields
114 along a steep precipitation gradient. We examined the effect of precipitation on microbial
115 nutrient limitation concerning the drivers in agroecosystem soil profiles. We developed the
116 following hypotheses: 1) microbial nutrient limitation may exhibit more significant responses
117 to precipitation in the subsoil than in the topsoil; and 2) the driving factors change with an
118 increase in depth. This study determines the microbial nutrient limitation patterns of semi-arid
119 to humid regions and identifies drivers of variations in these patterns across soil profiles in
120 agroecosystems.

121 **2. Materials and methods**

122 *2.1. Study site and experiment design*

123 The study was conducted in seven agroecosystems in the Northeast China Transect
124 (NECT), a climatic zone with significant precipitation gradients and a critical component of
125 the International Geosphere-Biosphere Programme terrestrial transects (Ni and Zhang, 2000;
126 Zhang et al., 1997). At latitudes ranging from 42° N to 44° N and longitudes from 123° E to



127 128° E within the NECT, precipitation and aridity demonstrate steep gradients (Ni and Zhang,
128 2000). From west to east, the MAP increases from 381 mm to 840 mm, and the mean annual
129 temperature (MAT) declines from 6.2 °C to 4.1 °C. We standardized our site selection by
130 focusing on maize planting areas. These agricultural sites, which have existed for over 50 years,
131 are located in flat areas and are managed conventionally under the same government and
132 according to the same policies. Conventional tillage included fall plowing, spring cultivation,
133 planting. The depth of agricultural tillage (commonly referred to as plough sole) is about 20
134 cm (Yan et al., 2017). Fertilizer for maize was applied with 200 kg N ha⁻¹, 70 kg P ha⁻¹ and
135 90 kg K ha⁻¹ as the base fertilizer at planting. The weeds were controlled using broad-spectrum
136 herbicides and supplemented as required with manual hoeing (Guo et al., 2020). The soil types
137 are classified as salt-alkali, light chernozem, chernozem, black soil, and dark brown soil, based
138 on the Chinese soil classification system (Xiong and Li, 1987). The basic properties of the
139 study sites are listed in Table S1.

140 2.2. Soil sampling and characterization

141 The seven sites, Baihe (BH), Longwan (LW), Dongliao (DL), Changchun (CC), Nongan
142 (NOA), Sijianfang (SJF) and Changling (CL), are positioned along a precipitation gradient.
143 Three plots (approximately 50 × 50 m) situated 50-100 m from each other were randomly
144 selected at each site. After removing the litter and rocks, the soil was sampled by horizon to a
145 maximum depth of 50 cm. Samples were taken discretely at five depth intervals (0-10 cm, 10-
146 20 cm, 20-30 cm, 30-40 cm and 40-50 cm). Three cores of soil, each with a diameter of 5.5 cm,
147 were randomly collected and gently mixed to form one composite sample for each layer of
148 each plot. In total, 105 soil samples (7 sites × 5 layers × 3 plots) were collected during the



149 maize harvest period in September 2018 to minimize the effects of agricultural management.

150 Each sample was separated into two portions. The first fresh soil samples were
151 immediately placed inside coolers containing ice packets for transport, returned to the
152 laboratory, and stored at 4 °C. SM was determined gravimetrically using 10 g of fresh soil
153 samples dried at 105 °C for 48 h to a constant weight. The second portion was air-dried for
154 edaphic property analyses. The soil pH was measured in a soil-water suspension (1:5 soil to
155 water ratio) using a glass electrode with a pH meter (Thermo Fisher Scientific Inc., USA).
156 Total C (TC) and N (TN) contents were determined using a Vario MICRO cube Elemental
157 Analyzer. Total P was first digested using H₂SO₄-HClO₄, and then the total P (TP)
158 concentration was measured using a continuous flow analyzer.

159 2.3. Soil enzyme activity

160 The potential activities of five hydrolytic enzymes, including two C-acquiring enzymes,
161 β-glucosidase (BG) and β-1,4-cellobiosidase (CB), two N-acquiring enzymes, β-N-
162 acetylglucosaminidase (NAG) and Leucine aminopeptidase (LAP), one P-acquiring enzyme,
163 acid phosphatase (AP), were measured at a temperature of 25 °C based on a modified version
164 of the methods described by Saiya-Cork et al. (2002) and German et al. (2011). The mass of
165 1.5 g of fresh soil was accurately recorded and 125 ml of 50 mM acetate buffer was added and
166 the samples were homogenized to prepare a suspension. The methodology of enzymes is
167 described in detail Fig. S1. The pre-experimental results exhibited that the activity was the
168 highest in the dark after 4 h of incubation at 25 °C. Fluorescence plates were read at 365 nm
169 excitation and 450 nm emission using a microplate reader (BioTek Synergy HTX, Winooski,
170 Vermont, USA). Eight analytical replicates were conducted per sample. Enzymatic activity



171 was enumerated as nanomoles of substrate released per gram of dry soil and incubation time
172 (nmol h⁻¹ g⁻¹).

173 2.4. *Quantification of microbial nutrient limitation*

174 Resource limitation was measured by vector analysis of enzymatic stoichiometry (length
175 and angle) (Moorhead et al., 2013) according to the relative activities of enzymes that acquire
176 C, N and P. The rationale for using the relative activities of C-, N-, and P-acquiring enzymes
177 to interpret the relative limitation of C, N, and P for microorganisms is based on ecosystem
178 stoichiometry and metabolic theory (Allison et al., 2010). Relatively longer vector lengths
179 express a greater C limitation, while vector angles < 45° and > 45° indicate the relative degrees
180 of N and P limitation, respectively (Moorhead et al., 2016). The vector length (unitless) and
181 angle (°) are calculated as follows:

$$182 \quad \text{Vector length} = \text{SQRT}((\ln \text{BG} / \ln(\text{NAG} + \text{LAP}))^2 + (\ln \text{BG} / \ln \text{AP})^2)$$

$$183 \quad \text{Vector angle} = \text{Degrees} (\text{ATAN2}((\ln \text{BG} / \ln \text{AP}), (\ln \text{BG} / \ln (\text{NAG} + \text{LAP}))))$$

184 2.5. *Statistical analysis*

185 The Shapiro-Wilk normality test was performed using the R version 3.5.3 ‘stats’ package
186 (R Development Core Team, 2019). We examined the relationships between precipitation and
187 different soil properties, enzyme activities, and C:N, C:P, and N:P ratios using linear
188 regressions. C-acquiring enzymes were calculated as BG + CB, N-acquiring enzymes were
189 calculated as NAG + LAP, and P-acquiring enzyme was represented by AP. Redundancy
190 analysis (RDA) was performed to analyze the relationships between soil enzyme activities and
191 selected soil properties for 105 samples by the packages “vegan” (Oksanen et al., 2019),
192 “adespatial” (Dray et al., 2018). We used the variation inflation factor (VIF) (“vif.cca” function)



193 to assess multicollinearity among soil properties ($VIF < 10$). The significance of RDA
194 correlations was tested Monte Carlo permutation test.

195 Similarly, we explored the relationships between precipitation and vector length and angle
196 for each soil profile. The analysis of covariance (ANCOVA) was used to detect the differences
197 in the slope of regression lines between the angle and length of vector and precipitation among
198 soil profiles by SPSS software version 26.0 (IBM Corporation, Armonk, New York, USA).
199 The angle and length of vector were dependent variables, precipitation was covariates, and soil
200 profiles were a grouping factor with five levels. A significant soil profiles \times covariate
201 interaction would indicate that the slope of the precipitation–vector relationship differs among
202 soil profiles. When the interaction was significant, the homogeneity of slopes assumption of
203 ANCOVA was rejected (Poorter et al., 2008). Differences at the $P < 0.05$ level were considered
204 statistically significant. To determine the relative influences of environmental variables on the
205 ecoenzymatic vectors in the five soil layers, we used the `rdacca.hp` function to calculate the
206 independent adjusted R-squared (R^2) and the percentage of each explanatory variable in the
207 RDA, running 999 permutations in the ‘`rdacca.hp`’ R package (Lai and Peres-neto, 2020). This
208 function performs variation partitioning and hierarchical partitioning to calculate the unique,
209 shared, and independent contributions of each predictor to the explained variation (R^2 and
210 adjusted R^2) in canonical analysis (RDA) by applying the hierarchy algorithm of Chevan and
211 Sutherland (1991).

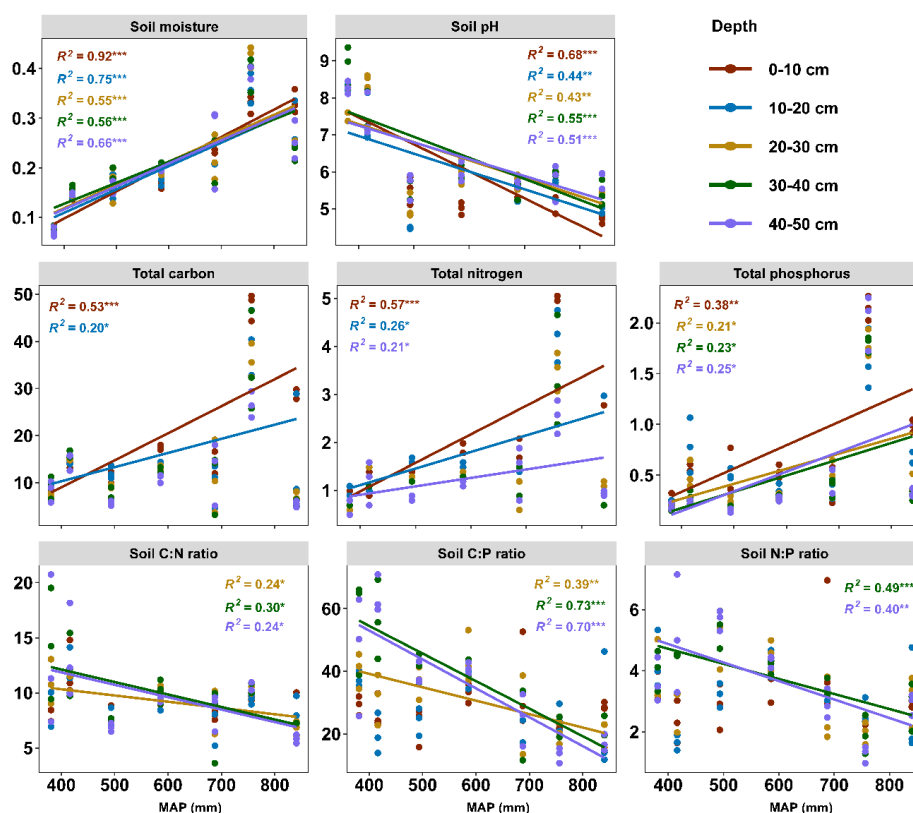
212 **3. Results**

213 *3.1. Soil properties and their stoichiometry*

214 The physicochemical properties and soil nutrient ratios displayed various patterns in



215 topsoil (0-20 cm) and subsoil (20-50 cm) along the precipitation gradient (Fig. 1). The soil pH
 216 decreased significantly from 9.36 to 4.46 with increasing precipitation. In contrast, soil
 217 moisture and nutrients (TC, TN, and TP) were positively correlated with precipitation, but this
 218 strong correlation was not observed in the subsoil for TC. Interestingly, for these components
 219 of soil nutrient stoichiometry components (TC:TN, TC:TP, and TN:TP), significant positive
 220 relationships with precipitation were observed in the subsoil but not in the topsoil.



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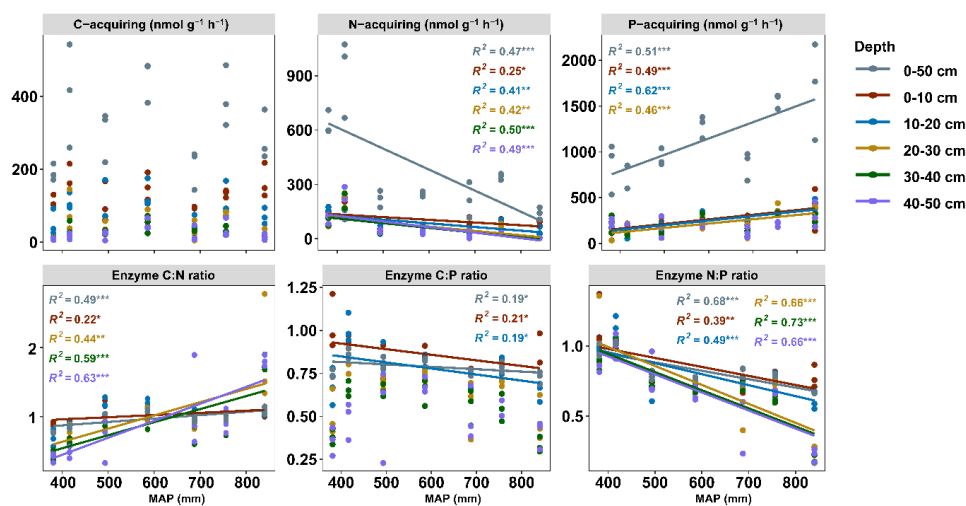
222 **Figure 1.** Linear regression between soil properties and MAP for each soil layer. Colored lines
 223 represent the regression of each soil layer (n=105). Statistics (R^2 and P values) for the
 224 regression are indicated using the corresponding colored text: *** $P < 0.001$, ** $P < 0.01$, and



225 * $P < 0.05$. Only significant differences are shown.

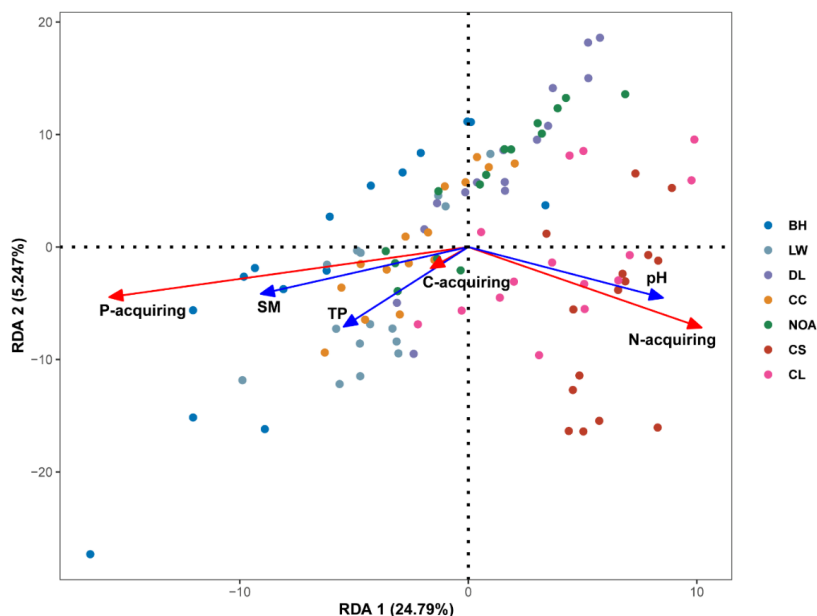
226 *3.2. Soil enzyme activity and stoichiometry across precipitation gradient*

227 The soil enzyme patterns were dissimilar to soil nutrient patterns along the precipitation
228 gradient (Fig. 2). C-acquiring enzyme activities were unaffected by precipitation. The N-
229 acquiring enzyme activities decreased across the precipitation gradient in all soil profiles, while
230 P-acquiring enzyme activities were positively correlated with precipitation at a depth of 0-30
231 cm but not at that of 30-50 cm. The C:P enzyme stoichiometry pattern was the opposite of that
232 of the P-acquiring enzyme. With increasing precipitation, the C:N and N:P enzyme nutrient
233 stoichiometries demonstrated increasing and decreasing trends, respectively; however, both
234 stoichiometries responded to precipitation more drastically in the subsoil than in the topsoil.
235 The major environmental variables shaping the soil enzyme pattern were determined through
236 RDA (Fig. 3). According to the RDA results, approximately 33.58% of the variance in total
237 soil enzyme activity can be explained by environmental variables. The first two RDA axes
238 explained 30.04% of the relationships between enzymes and the environment (24.79% and
239 5.25% for the first and second axes, respectively). According to the model selection results,
240 soil pH and moisture were the most important factors affecting the enzyme activity.



241

242 **Figure 2.** Patterns of enzyme activity and stoichiometry along the precipitation gradient for
243 each soil layer. Colored lines represent the regression of each soil layer (n=105). Statistics (R^2
244 and P values) for the regression are indicated using the corresponding colored text: *** $P <$
245 0.001, ** $P <$ 0.01, and * $P <$ 0.05. Only significant differences are shown. C, N, and P represent
246 carbon, nitrogen, and phosphorus.



247

248 **Figure 3.** Redundancy analysis of 105 soil samples based on soil enzyme activity and dominant

249 environmental factor ordination biplot. The overall model is significant ($P < 0.001$).

250 Environmental factors: SM, soil moisture; pH, soil acidity and alkalinity; TP, total phosphorus.

251 3.3. Soil microbe metabolic limitation and its driving factors

252 The nutrient limitation of soil microbes was calculated using enzyme stoichiometry and

253 quantified by vector angle and length (Fig. 4). Vector angles ranging from 36.1° to 80.5° were

254 positively related to the precipitation, and they displayed a more intense correlation with

255 precipitation in the subsoil than in the topsoil (Fig. 4a; Table S2). We also subdivided the

256 driving factors for nutrient limitation according to the soil layer (Fig. 5). We observed that pH

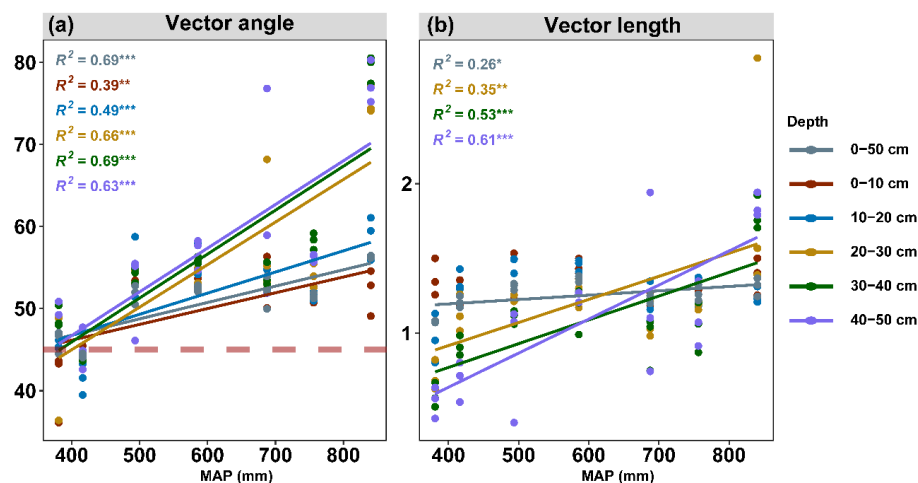
257 had the largest independent effects on vector angle in topsoil, explaining 69.41% and 44.78%

258 of the variance in the depth ranges of 0-10 cm and 10-20 cm, respectively (Fig. 5a). From the

259 topsoil to the subsoil, the soil physicochemical properties (soil pH and moisture) variance

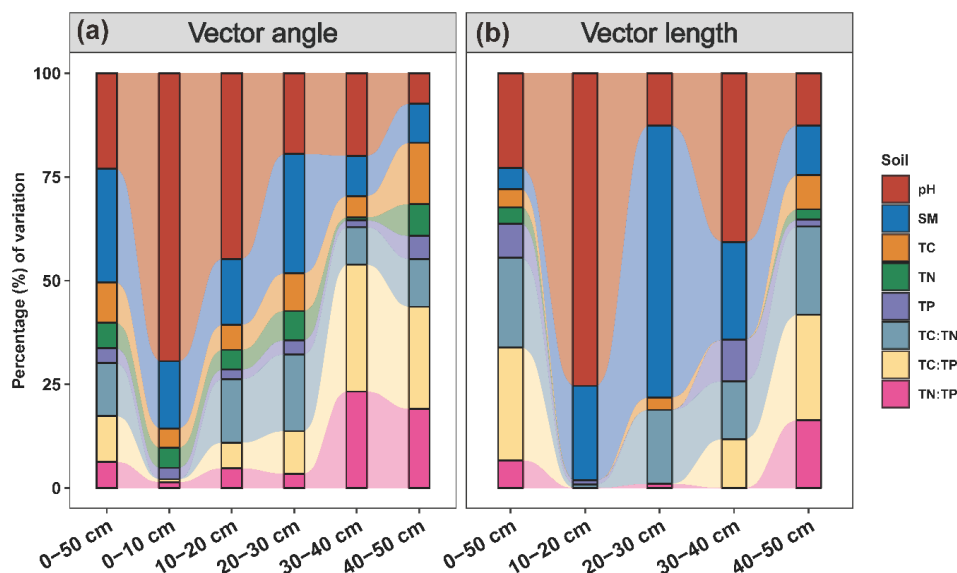


260 explanation ranged from 85.64% to 16.79% (Fig. 5a). In contrast, the interpretations of soil
261 nutrients and their stoichiometry ranged from 14.36% to 83.21% (Fig. 5a). Vector length was
262 positively correlated with precipitation only in the subsoil (Fig. 4b). The interpretation of soil
263 nutrient stoichiometry with respect to the vector length grew stronger with depth (Fig. 5b).



264

265 **Figure 4.** Effects of MAP (mm) on microbial N/P limitation (A) and microbial C limitation
266 (B). Soil microbial N/P limitation is represented by vector angles: $> 45^\circ$ represents P limitation,
267 and $< 45^\circ$ represents N limitation. Soil microbial C limitation is represented by vector lengths,
268 where microbial C limitation increases as the values increase.



269

270 **Figure 5.** Influences of soil variables on vector angles (A) and lengths (B). Percentage of
271 variation in vector angles and lengths explained by soil variables. SM, soil moisture; pH, soil
272 acidity and alkalinity; TC, total carbon; TN, total nitrogen; TP, total phosphorus.

273 4. Discussion

274 4.1. Patterns of microbial nutrient limitation across precipitation regions throughout soil 275 profiles

276 We did not observe a significant relationship between C-acquiring enzymes and
277 precipitation (Fig. 2), and surprisingly, the vector length increased with precipitation,
278 indicating that microbial C limitation was gradually exacerbated along the precipitation
279 gradient in 0-50 cm. Moreover, N-acquiring enzyme activity decreased and P-acquiring
280 enzyme activity increased with increasing precipitation (Fig. 2), which alters the stoichiometry
281 of enzymes leading to C limitation. We attributed this pattern of microbial C limitation to



282 changes in soil nutrients and their stoichiometry across precipitation regions (Fig. 5B).
283 Microbial resource limitation is relative (Chen et al., 2018); thus, microbial C limitation can
284 depend on the combined effect of other nutrients.

285 We detected a decrease in the enzymatic C:P and N:P acquisition ratios with increasing
286 precipitation (Fig. 2). Microorganisms can shift from secreting more C-acquiring enzymes and
287 N-acquiring enzymes to secreting more P-acquiring enzymes, which implies an increased
288 demand for microbial P. Moreover, the vector angle became greater than 45° with increasing
289 precipitation (Fig. 4A). Our results notably indicated that microbial P limitation was gradually
290 exacerbated along the precipitation gradient. P is predominately derived from mineral aerosol
291 deposition and weathering, and it is highly dependent on pH (Mahowald et al., 2008; Thingstad
292 et al., 2005). High soil moisture and low soil pH increase P-acquiring enzyme activity, which
293 facilitates P solubilization (Collavino et al., 2010; Xu et al., 2020b). Our study also identified
294 that the P-acquiring enzyme was influenced by soil pH and moisture, and soil pH and moisture
295 were the main factor determining the pattern of soil enzymes along the precipitation gradient
296 (Fig. 3). Therefore, soil pH and moisture play an essential role in microbial P limitation
297 throughout soil profiles.

298 Furthermore, soil nutrients and their stoichiometry and soil physicochemical properties
299 are equally important for microbial phosphorus limitation in 0-50 cm (Fig. 5A). More C and N
300 might lead to microbial P limitation (Cui et al., 2020). This observation may be explained for
301 two reasons. On the one hand, if the C and N accumulated in the soil dilute the soil P content,
302 the microorganisms may be limited by P due to the elemental stoichiometric balance of the
303 microorganisms (Chen et al., 2017; Sinsabaugh et al., 2009; Sinsabaugh et al., 2008). Thus,



304 despite the increase in soil P, microbial P limitation remains because of the slow growth rate
305 of P relative to that of C and N. On the other hand, phosphatase molecules are rich in C and N,
306 and microorganisms require sufficient C and N to construct extracellular phosphatases
307 (Houlton et al., 2008; Treseder and Vitousek, 2001). Phosphatase may be highly responsive to
308 changes in soil elements, a critical strategy for organism adaption to human modifications of
309 the global biogeochemical cycle that can delay the onset of P limitation (Güsewell and Freeman,
310 2005; Marklein and Houlton, 2012). These findings demonstrate that precipitation regulates
311 soil physicochemical properties and nutrient status (Ma et al., 2015; Smith, 2011), which
312 consequently greatly enhances microbial C and P limitation. Our current results take a new step
313 and are valuable for understanding microbial nutrient limitation along the precipitation gradient
314 in agroecosystems.

315 *4.2. Contrasting mechanisms governing microbial nutrient limitation in topsoil and subsoil*

316 Soil properties and enzyme activities can change with soil profiles (Li et al., 2020; Luan
317 et al., 2020; Stone et al., 2014), and we observed wide differences in soil and enzyme
318 stoichiometry between the topsoil (0-20 cm) and subsoil (20-50 cm). Our study found that
319 enzymatic C:P acquisition ratios were negatively correlated with precipitation only in the
320 topsoil, which is attributable to the phosphatase distribution pattern. Enzymatic N:P acquisition
321 ratios decreased, but enzymatic C:N acquisition ratio increased along the precipitation gradient,
322 and changes substantially in the subsoil. The vector angles visualized that microbial P
323 limitation increased and responded more strongly to precipitation in the subsoil than in the
324 topsoil. In topsoil, soil pH and moisture are the major drivers of microbial P limitation (Fig.
325 5A). Soil pH governs phosphatase spatial variations and is positively correlated with enzymatic



326 C:P and N:P acquisition ratios (Xu et al., 2017). It is interesting to note that, compared to the
327 topsoil, soil nutrient availability and their stoichiometry are important factors in determining
328 microbial P limitation in subsoil (Fig. 5A). Soil substrate availability greatly influences enzyme
329 activity patterns (Stock et al., 2019), and low resources (C and N) and O₂ concentrations can
330 limit microbial metabolism (Fierer, 2017; Gu et al., 2017). Zhou et al. (2020) also found that
331 the soil C:N ratio explains a large proportion of the variation in microbial nutrient acquisition.
332 Emerging evidence suggests that subsoils, exhibit more varied organic matter sources,
333 microbial communities, and substrate availability compared to those of the topsoil (Chen et al.,
334 2019; Rumpel et al., 2012). Differences in soil properties across soil depths can lead to greater
335 environmental filtering or competition for nutrients among microorganisms than in the topsoil
336 (Bahram et al., 2015; Schlatter et al., 2020). Especially in agroecosystems, this difference
337 between topsoil and subsoil may be attributed to the agricultural tillage depth, which is
338 typically 0-20 cm. Cropland activities contribute to the homogenization of microbial
339 communities (Rodrigues et al., 2013), and alter microbial resource limitations (Cui et al., 2021).
340 On the contrary, the original soil conditions were preserved in the subsoil, resulting in microbes
341 secreting soil enzymes that are less responsive to precipitation in the topsoil than that in the
342 subsoil and attenuate the reaction of microbe nutrients limitation to precipitation in the topsoil.

343 Ecosystems cannot be limited by carbon or nutrients alone (Soong et al., 2020), and most
344 microbial communities are co-limited by energy and key nutrients (Moorhead et al., 2016).
345 Another marked trend is that the vector length increases in the subsoil with precipitation. This
346 result indicates that microbial C limitation increased in the subsoil along the precipitation
347 gradient. The detected relationships might result from that soil C:N and C:P ratios



348 progressively decline across the precipitation region in the subsoil (Fig. 1), which can lead to
349 insufficient soil C (Chen et al., 2018). In contrast, there is no relationship between microbial C
350 limitation and precipitation in topsoil. This might occur because the effect of precipitation is
351 masked by cropland activity, causing soil nutrient stoichiometry in topsoil that exhibited no
352 response to precipitation. Moreover, soil nutrients and their stoichiometry increase in their
353 interpretation of C limitation with depth (Fig. 4). Therefore, soil C:N:P stoichiometry
354 tremendously affected microbial groups involved in major biogeochemical processes (Luo et
355 al., 2020). Additionally, balanced nutrient stoichiometry is essential for maintaining microbial
356 and elemental homeostasis (Cui et al., 2019b).

357 Further, P limitation changes rapidly in the subsoil, which contributes to C limitation.
358 Nutrients are interdependent and they have a complex coupling, so changing in one nutrient
359 cycle type can alter the availability of another nutrient (Marklein and Houlton, 2012). The role
360 of P limitation in the storage capacity of the C ecosystem has previously been substantially
361 underestimated (Peng and Thomas, 2010). P affects the C-storage capacity both directly and
362 indirectly by affecting the plant growth capacity and limiting biological activities for N fixation
363 (Peñuelas et al., 2012). A model predicted less efficient microbial growth in P-limited soils,
364 indicating that P availability may affect C cycling (Waring et al., 2014). These nutrient
365 interdependencies confirm that resource limitation coupling is complex, and excess
366 precipitation leads to depletion of P and C and induces imbalances in agroecosystems. Our
367 results highlight the contrasting mechanisms governing microbial nutrient limitation in topsoil
368 and subsoil. Microbial P limitation was controlled by the soil pH and moisture in the topsoil,
369 while microbial C and P limitation in the subsoil were regulated by the soil nutrients and their



370 stoichiometry. The findings suggest that ecosystem functions and microbial nutrient limitation
371 cannot be comprehensively assessed from topsoil observations alone.

372 **5. Conclusions**

373 Understanding the key processes by which precipitation threatens microbial nutrient
374 limitation allows the assessment of nutrient trade-offs in agroecosystems, which can help us
375 meet crop production goals under change in precipitation. We detected spatial variations in soil
376 enzyme activities and stoichiometry according to the agroecosystem soil profile and along a
377 precipitation gradient. We observed an increase in microbial P limitation in the topsoil and a
378 stronger limitation of microbial C and P in the subsoil compared to the topsoil and elucidated
379 the differentiation of mechanisms among soil profiles. Our study also provides insights into the
380 change mechanism of microbial resource limitation in the context of global precipitation
381 changes and protecting vital life-support systems.

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389 **Declaration of interests**

390 The authors declare that they have no known competing financial interests or personal
391 relationships that could have appeared to influence the work reported in this paper.



392 **CRedit authorship contribution statement**

393 **Jingjing Yang:** Conceptualization, Methodology, Software, Formal Analysis,
394 Investigation, Writing-Original Draft, **Pingting Guan:** Methodology, Formal Analysis, Data
395 Curation, Writing-Review & Editing, **Peng Zhang:** Validation, Investigation, Resources,
396 **Yunga Wu:** Investigation, Resources, Visualization, Writing-Review & Editing, **Donghui Wu:**
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398 Funding Acquisition.

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