Performance of three machine learning algorithms for predicting soil organic carbon in German agricultural soilSpatial prediction of organic carbon in German agricultural topsoil using machine learning algorithms

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10 Abstract

8

11 As the largest terrestrial carbon pool, Soil organic carbon (SOC), as the largest terrestrial carbon pool, has the 12 potential to influence and mitigate climate change and mitigation, and consequentlyhence the importance of SOC 13 monitoring is important in the frameworks of different various international treaties. There is therefore a need for 14 hHigh resolution SOC maps are therefore required. Machine learning (ML) offers new opportunities to do this 15 develop these due to its eapability for data mining of large datasets. The aim of this study, therefore, was to test 16 apply three commonly used algorithms commonly used in digital soil mapping – random forest (RF), boosted 17 regression trees (BRT) and support vector machine for regression (SVR) - on the first German Agricultural Soil 18 Inventory to model agricultural topsoil (0-30 cm) SOC content- and develop a two-model approach to address the 19 high variability of SOC in German agricultural soils. Model performance is often limited by the size and quality 20 of the soil dataset available for calibration and validation. Therefore, the impact of enlarging the training data was 21 tested by including data from the European Land Use/Land Cover Area Frame Survey for agricultural sites in 22 Germany. Nested cross-validation was implemented for model evaluation and parameter tuning. Moreover, Gerid 23 search and the differential evolution algorithm were also applied to ensure that each algorithm was appropriately 24 tuned and optimised suitably. The SOC content of the German Agricultural Soil Inventory was highly variable, ranging from 4 g kg⁻¹ to 480 g kg⁻¹. However, only 4% of all soils contained more than 87 g kg⁻¹ SOC and were 25 26 considered organic or degraded organic soils. The results showed that SVR provided produced the best 27 performance with an RMSE of 32 g kg⁻¹ when the algorithms were trained on the full dataset. However, the average 28 RMSE of all algorithms decreased by 34% when mineral and organic soils were modelled separately, with the best result from SVR with a RMSE of 21 g kg⁻¹. Model performance is often limited by the size and quality of the 29 30 available soil dataset for calibration and validation. Therefore, the impact of enlarging the training data was tested 31 by including 1223 data points from the European Land Use/Land Cover Area Frame Survey for agricultural sites 32 in Germany. The model performance was enhanced for maximum by up to 1% for mineral soils and by 2% for 33 organic soils. Despite the capability of machine learning algorithms in general, and particularly SVR in particular, 34 in to modelling SOC on a national scale, the study showed that the most important aspect to for improvinge the 35 model performance was to separate the modelling of mineral and organic soils.

36 **1** Introduction

- 37 Soil organic carbon (SOC) is the largest terrestrial carbon pool (Wang et al., 2020) and plays an essential role in 38 agriculture. Since SOC influences various physical, chemical and biological properties of soil (Reeves, 1997), 39 numerous studies recognise it as a crucial indicator of soil quality (Castaldi et al., 2019; Meersmans et al., 2012a; 40 Reeves, 1997) (Castaldi et al., 2019; Meersmans et al., 2012; Reeves, 1997) and therefore. Thus, its decline is 41 identified as a threat that leads to soil degradation (Castaldi et al., 2019; Poeplau et al., 2020). Moreover, when 42 considering carbon sequestration, the SOC pool provides the option for climate change mitigation (Meersmans et 43 al., 2012a; Ward et al., 2019) (Meersmans et al., 2012; Ward et al., 2019). Consequently, SOC monitoring is 44 therefore important in the frameworks of various international treaties such as the European Union Soil Thematic 45
- 46 al., 2020) (Meersmans et al., 2012; Poeplau et al., 2020), and There is, therefore growing interest in understanding

Strategy and the United Nations Framework Convention on Climate Change (Meersmans et al., 2012b; Poeplau et

- 47 the spatial distribution of SOC at different scales in response to an increasing demand for a better assessment of
- 48 SOC (Minasny et al., 2013). This is particularly important for agricultural land due to its potential for carbon
- 49 sequestration (Lal, 2004).

50 In digital soil mapping (DSM), a soil attribute is formulated asdescribed by an empirical quantitative function of 51 seven factors: soil properties, climate, organisms, topography, parent material, time, and spatial position 52 (McBratney et al., 2003). Therefore, tThis function, known as the SCORPAN model, can be applied to spatially 53 predict the soil attribute property of interest (Minasny et al., 2013). Within this framework, machine learning 54 algorithms aim to automatically extract the information from the data for predictive purposes (Behrens et al., 55 2005). This is of particularly intriguing interest in view of the recent expansion of databases at a different scale in 56 soil sciencesoil databases and the the complexity of the covariates vast amount of data to approximate the soil 57 forming factors in recent years (McBratney et al., 2003; Wadoux et al., 2020), thus making DSM cost-effective, 58 time-efficient and applicable over large areas with good results (Behrens and Scholten, 2006; Camera et al., 2017).

59 Despite the advantages of DSM, it is crucial to consider note that its application requires soil databases of an 60 adequate sample size for training and testing. Furthermore, consistent and quality-checked datasets are a 61 prerequisite for DSM. Several soil inventories and monitoring networks for SOC have been formed established on 62 a national scale in countries such as Sweden (Poeplau et al., 2015), France (Belon et al., 2012; Arrouays et al., 63 2002) (Belon et al., 2012; Meersmans et al., 2012) Denmark (Taghizadeh-Toosi et al., 2014) and Scotland 64 (Chapman et al., 2013). NonethelessHowever, in Germany the most critical shortcomings of soil inventories-in 65 Germany concern are the lack of a large-scale, high-quality SOC inventory monitoring (Wiesmeier et al., 2012) 66 with a-periodic and standardised sampling focused on agricultural soils (Prechtel et al., 2009). These issues have 67 now been solved-addressed in the first German Agricultural Soil Inventory (Poeplau et al., 2020). This inventory 68 was conducted carried out on a national scale with a sampling depth down to 100 cm considering a sampling depth 69 of 1 m at 3104 sampling sites covering agricultural land. Furthermore, on a European scale, the Land Use/Land 70 Cover Area Frame Survey (LUCAS) undertaken in 2009 is the first harmonised topsoil survey with physico-71 chemical analyses of georeferenced topsoil samples in-from 23 European states (Tóth et al., 2013). Therefore, by 72 taking advantage of DSM and of both the German Agricultural Soil Inventory and the LUCAS survey, it is possible 73 to regionalise from single-point measurements to obtain complete high-resolution cover soil data nationwide and 74 thus provide a baseline for both SOC monitoring as well as forand environmental and climatic modelling for 75 Germany.

76 Boosted regression trees (BRT), random forest (RF) and support vector machine for regression (SVR) are among 77 the most widely used algorithms in DSM (Padarian et al., 2020). For example, Martin et al. (2014) predicted topsoil 78 SOC on a national scale for France using the BRT algorithm and comparinged its results when the same algorithm 79 was coupled with a geostatistical approach. They concluded that due to the large distances between sampling sites, 80 spatial autocorrelation is unlikely since spatial autocorrelation is not feasible in most national inventories, and the 81 BRT algorithm alone is sufficient for this purpose. This algorithm has was also been used on a national scale in 82 China for data from the 1980s and 2010s in order to predict topsoil SOC and its spatial-temporal change, as well 83 as the main drivers of its variability (Wang et al., 2021). Moreover, RF has also become more popular in DSM due 84 to its relative simplicity and performance. For example, this algorithm was implemented to map topsoil SOC on a 85 national scale in Madagascar and obtain-identify its main drivers (Ramifehiarivo et al., 2017). Ramifehiarivo et al. 86 (2017) concluded that the uncertainty of the map generated by RF model traininguncertainty of the algorithm 87 was lower when compared with the maps that were formerly generated for the country. Moreover, this algorithm 88 was compared with the Cubist model-algorithm for mapping SOC at different resolutions on a regional scale in 89 China and could was found to outperformed it (Li et al., 2021). Fewer studies have used SVR than RF to predict 90 SOC-than RF. Studies have mainly implemented SVR on a regional scale with a limited number of samples 91 (Forkuor et al., 2017; Were et al., 2015) or on a national scale (Switzerland) with very few samples (150 samples 92 from the European LUCAS survey) (Zhou et al., 2021). However, in a study comparing different algorithms, 93 including SVR and RF, on a continental scale and within each country in Latin America, the results indicated that 94 the best-performing algorithm varied in different countries from country to country (Guevara et al., 2018). Theis 95 difference mainly depended on sample-data density, quality, dispersity and representativeness and also-country 96 size, which affects the heterogeneity of land use and environmental conditions.

97 Another important consideration when applying machine learning is the impact of the parameter-tuning strategy 98 in algorithm performance. This is particularly crucial when the objective of the study is the comparisons ofto 99 compare different machine learning algorithms. Although some algorithms are less sensitive to tuning, this step is 100 more important for others, particularly those with a higher number of parameters (Tziachris et al., 2020; Wadoux 101 et al., 2020). Furthermore, as algorithms differ by the type of their parameters, continuous or discrete, the chosen 102 strategy should be aligned in accordance with this difference (Ließ et al., 2021). This is particularly more important 103 for algorithms with continuous parameters. For example, it has been shown that that the performance of SVR and 104 BRT is has been shown to be better and more stable when optimised by a differential evolution (DE) algorithm 105 than tuned by grid search (Zhang et al., 2011; Gebauer et al., 2020). Despite this importance, in a review of studies 106 that have applied DSM, Wadoux et al. (2020) state that almost half of them implemented parameter tuning, with 107 grid search the most common strategy applied for this purpose. This finding indicates that the role of parameter 108 tuning and optimisation is unfortunately undermined in DSM. This is particularly evident when the application of 109 machine learning in this field is compared with other fields, where various studies have shown the impact of 110 parameter-tuning strategies on the performance of algorithms such as SVR and BRT (Liang et al., 2011; Santos et 111 al., 2021; Bhadra et al., 2012; Deng et al., 2019).

112 The aims of the present study werewas therefore: i) to address the above-mentioned parameter-tuning issue and

113 consequently provide a true comparison of the performance of BRT, RF, and SVR in modelling the SOC contents

approach to address the high variability of SOC in German agricultural soils and compare it with a single-model approach.

118 2 Materials and methods

119 2.1 Soil data

120 The models were built using SOC content data from two soil inventories. The first dataset was from the German 121 Agricultural Soil Inventory, which consists comprise of 3104 sites with a fixed collected along a grid of 8x8 km 122 throughout Germany (Poeplau et al., 2020). The sites were sampled and analysed for different soil properties, 123 including SOC content measured via dry combustion, for the upper 30 cm of the soil between 2012 and 2018. The 124 second dataset was the European LUCAS survey that provides SOC content, similarly also measured via dry 125 combustion, for all EU countries, with the sampling depth limited to 0-20 cm (Tóth et al., 2013). For Germany, 126 data collected on agricultural soils cover 1223 sites. Therefore, in order to harmonise the depths of both datasets, 127 theyse were subdivided into two classes: mineral and organic soils -classes according to a SOC threshold value of 128 87.0 g kg⁻¹. Accordingly, considering all soils above this threshold were considered as organic soils comprising 129 peat soils and disturbed and degraded peat soils (Poeplau et al., 2020). Linear regression functions were derived 130 for both mineral, Eq. 1, and organic, Eq. 2, soil classes on behalf of the data of the German Agricultural Soil 131 Inventory to relate the SOC content of 0-30 cm to that of 0-20 cm. Linear correlation functions between 0.30 cm 132 and 0-20 cm were derived for each soil class of the German Agricultural Soil Inventory separately. These functions 133 were then applied to the corresponding soil class from the LUCAS data in order to estimate 0-30 cm topsoil SOC. 134 With a slope of 0.881 for mineral soils and 1.02 for organic soils, they changed the mean of LUCAS data by less 135 than 6%. The depth extrapolated values of mineral and organic soils were then combined to form the complete 136 dataset. The 0-30 cm LUCAS data generated and the original 0-20 cm LUCAS data were then used by each 137 algorithm to check the effect of depth extrapolation.

- 138 y = 1.01 + 0.881x_____
- 139 y = 1.6 + 1.02x____

140 2.2 Covariates

Covariates from multiple sources were included to approximate the SCORPAN factors throughout Germany. In the case of multiple data products for one covariate, the one with the best quality (least-fewer_artefacts), and the highest spatial resolution was added. These were then resampled in ArcGIS (ESRI, 2013) using the INSPIRE standard grid at 100 m resolution (Eurostat grid generation tool for ArcGIS). The resampling method was either the nearest neighbour for categorical covariates or bilinear interpolation for continuous covariates. The same INSPIRE grid was also used to rasterise the vector covariates as well. Finally, they were stacked and overlaid on SOC databases in order to extract the values at the sampling points.

(1)

(2)

148 Following the SCORPAN framework, 24 covariates including x and y <u>coordinates for</u> spatial positions were

compiled. In order to capture-represent the climate factors (C factor), precipitation (DWD, 2018c), sunshine

duration (DWD, 2017), summer days (DWD, 2018b), and minimum temperature (DWD, 2018a) were used applied

according to the study of Schneider et al. (2021). Using principal component analysis, these four covariates were

152 indicated identified to be the most important among out of 34 available climate factors for SOC in the German

Agricultural Soil Inventory dataset. Moreover, type of agricultural land use is one of the main drivers of SOC

variability <u>at on</u> a national scale (Poeplau et al., 2020). <u>Thus, therefore</u> the land_-use map from the official topographic<u>cartographic</u> information system (BKG, 2019) with its corresponding classes according to the German Agricultural Soil Inventory was rasterised and included. This is a categorical covariate, representing the organism factor of SCORPAN (O factor), <u>that-which</u> distinguishes croplands from grasslands and captures their spatial distribution throughout Germany.

159 The European Digital Elevation Model (EUDEM) (European Union Copernicus Land Monitoring Service, 2016) 160 with original resolution of 25 m was resampled to 100 m. and sSix covariates derived from the is layer resampled 161 layer were also added to integrate the topography and relief parameters (R factor). Slope, plan curvature and profile 162 curvature, generated on-with SAGA (Conrad et al., 2015), were included to capture the slope's gradient, convexity-163 concavity and convergence-divergence. These factors influence the soil distribution throughout the landscape, e.g. 164 affecting flow over the surface, thus impacting SOC and its dynamic (Ritchie et al., 2007). Moreover, north south 165 and east-west aspects were slope exposition (aspect) was obtained calculated from the from EUDEM as these it 166 influences soil development and subsequently affects SOC (Carter and Ciolkosz, 1991). The circular variable was 167 then decomposed into northness and eastness. The Topographic Wetness Index (TWI), generated on SAGA 168 (Conrad et al., 2015), was also added since it captures the soil moisture distribution of the landscape and some 169 studies have shown its direct correlation has a direct correlation with SOC (Pei et al., 2010). A geomorphographic 170 map of Germany (Federal Institute for Geosciences and Natural Resources (BGR), 2007) -containing featuring 25 171 geomorphic categories was also used to distinguish between four different landscape areas of the country: North 172 German lowlands, highlands, Alpine foothills and the Alps.

173 Continuing with the framework, a large-scale soil landscape unit map ("Soil Scapes in 174 GermanyBodengrosslandschaft") (Federal Institute for Geosciences and Natural Resources (BGR), 2008) 175 comprising 38 classes was used. This covariate divides Germany by various geo-factors that can be compiled into 176 a map with 12 soil regions representing mainly the parent materials. Similarly, large scale the soil-climate region 177 map ("Bodenklima") (Roßberg et al., 2007) with 50 classes was added. Moreover, the Hydrogeological unit 178 according Hydrogeological map of GermanyGermany's hydrogeological unit map (BGR and SGDG, 2019). The 179 hydrogeological map provides information about -hydrogeologically relevant attributes including consolidation, 180 type of porosity, permeability, type of rock and geochemical classification. lithology and its hydrological 181 characteristics. These categorical maps were rasterised and applied to the model as the P factor of SCORPAN. 182 Moreover, the soil factor of the framework (S factor) was captured by eight covariates that represent different 183 aspects of its properties: the map of organic soils (Roßkopf et al., 2015) that distinguishes mineral soils from 184 organic ones and explains their spatial distribution throughout the country, as well as the maps of nitrogen 185 (Ballabio et al., 2019) and clay content (Ballabio et al., 2016) since they directly correlate with SOC. As nitrogen 186 is a crucial component of soil organic matter, regions with higher total nitrogen have higher SOC (Ballabio et al., 187 2019). Also for clay content, different studies have shown that coarser soil textures tend to have a lower 188 accumulation of SOC (Zhong et al., 2018; Hoyle et al., 2011). The M-map of pH from (Ballabio et al., (2019) was 189 included since soil pH directly impacts microbial activities that influence the turnover of soil organic matter, and 190 consequently negatively correlates with SOC (Malik et al., 2018). Furthermore, the map of available water capacity 191 (Ballabio et al., 2016) was used as this soil properties property is another interactive factor with SOC through plant 192 productivity and soil texture (Burke et al., 1989; Yu et al., 2021). Soil erosion is also a key factor in the SOC cycle 193 (Li et al., 2019), which was added through the soil-map of Europe's net soil erosion and deposition rates erosion

map of Europe (Borrelli et al., 2018). <u>Based on the WaTEM/SEDEM model, this map illustrates the potential</u>
 spatial displacement and transport of soil sediments due to water erosion (Borrelli et al., 2018). <u>Figure S1 provides</u>
 a more detailed view for better visualisation of the covariates that were used in this study.

197 2.3 Boosted Regression Trees

198 Developed by Friedman et al. (2000), BRT is a tree-based algorithm that applies boosting method-to improve 199 accuracy. Boosting method relies on combining several approximate prediction models rather than obtaining one 200 single-highly accurate one-model (Schapire, 2003). Thus, the decision trees are grown sequentially so that each 201 decision tree predicts the residual of the previous one and therefore. Consequently, the number of trees influences 202 the performance of the algorithm and requires tuning. However, to incorporate randomness into the model and 203 subsequently increase the robustness of performance, the trees are grown on a randomly selected data subset with 204 no replacement (Friedman, 2002). The size of this subset is controlled by a parameter known as a bag fraction. 205 Furthermore, the contribution of each new tree to the final model is regularised by learning rate, also known as 206 shrinkage(Friedman et al., 2009). Finally, the number of splits in each tree that divides the response variable into subsets is optimised by interaction depth. The BRT model was built in R using the "gbm" package (Greenwell et 207 208 al., 2019).

209 2.4 Random Forest

210 Similar to BRT, RF is another tree-based algorithm. RF uses bootstrap sampling of the dataset for growing a 211 decision tree. Subsequently, by aggregating the results of a large number of decision trees, the bias and variance 212 of the final model can be reduced (Breiman, 1999). The method of bootstrapping in conjunction with aggregating, 213 known as bagging, increases the robustness and stability of RF. However, the trees from different bootstraps may 214 form a similar structure if all covariates participate in a split of each node. Thus, the variance cannot be reduced 215 optimally through the bagging process (Kuhn and Johnson, 2013). In order to avoid this tree correlation, a random 216 subset of <u>covariates</u>, i.e. predictors, is selected at each split. The parameter m_{try} defines the number of predictors 217 included in this subset and should be tuned (Kuhn and Johnson, 2013). The RF algorithm was implemented by 218 setting the number of trees to 1000 and using the "Ranger" package (Wright and Ziegler, 2017) in R.

219 2.5 Support Vector Regression

220 SVR is a form of support vector machine adopted for regression. From all possible solutions, i.e. estimation 221 function, for the problem, SVR tries to obtain an estimation function with that has at most the maximum ε error 222 deviation from the response values of the training data while minimising model complexity (Smola and Schölkopf, 223 2004). Thus, a symmetrical tolerance threshold, $\boldsymbol{\varepsilon}$ -insensitivity zone, is created around the estimation function 224 within which the vectors are not penalised (Awad and Khanna, 2015). However, tThe data vectors of the samples 225 that lie on the boundary of the ε -insensitivity zone are called support vectors. The vector lying within the 226 insensitivity zone are not penalized. Therefore, ϵ is an optimisable parameter that controls the width of ϵ -227 insensitivity, alters the model complexity and inversely impacts the number of support vectors inversely 228 (Cherkassky and Ma, 2004). Moreover, the trade-off between model complexity and tolerance of $\boldsymbol{\varepsilon}$ deviation is 229 controlled by a parameter named C (Smola and Schölkopf, 2004; Cherkassky and Ma, 2004). Optimising the C 230 parameter has a crucial impact on SVR performance since a high C can lead to overfitting, while a low C can cause 231 under fitting (Kuhn and Johnson, 2013). The use of kernel functions makes SVR a powerful tool for nonlinear 232 problems. By implementing these functions, SVR can map the data space from its original dimension to a higher dimensional space where a nonlinear problem can be solved linearly. In this study, the Radial Basis Function
(RBF) kernel was used with gamma as its tuneable parameter. This parameter affects the generalisation
performance of SVR by <u>inversely</u> controlling the influence of support vectors <u>inversely</u> (Battineni et al., 2019).
SVR was implemented from the package e1071 in R (Hornik et al., 2021).

237 2.6 Performance evaluation

238 When training a predictive model, it is important to evaluate its generalisation performance on unseen data of the 239 same type (Hawkins et al., 2003). However, as the number of available samples is usually a limiting factor, the 240 evaluation process is often done by k-fold cross validation (CV). Therefore, the dataset is divided into k folds and 241 k-1 folds are used for training the model and one fold for testing. This process is repeated k times so each fold 242 participate in train and test. However, as the number of available samples is usually a limiting factor, the evaluation 243 process is often done by randomly splitting the available dataset into training and testing sets multiple times, i.e. 244 eross-validation (CV). Although this process is effective, it is not entirely immune from biased estimation of error. 245 However, to ensure that the estimated error in model evaluation is as unbiased as possible. However, to ensure the 246 robustness of the model, every each model training step should be performed within the CV. This includes finding 247 the best parameter sets for the chosen algorithm (Varma and Simon, 2006). Thus, the algorithms in this study were 248 applied on a stratified nested CV.

249 First, to ensure that the SOC distribution was represented in the CV scheme, Germany was divided into 50 strata 250 using a 100x100 km INSPIRE grid-into 50 strata. Random samples from each stratum were then taken and 251 compiled into a fold. This procedure was continued to create five folds and was repeated five times, forming the 252 outer loop of CV used for model evaluation. Large-A long distance between neighboring samples, 8120 m on 253 average, prevents train and test data from being spatially autocorrelated. Since the aim was to tune the algorithms' 254 parameters of the algorithms, the training set of the outer loop of CV was nested, creating five folds as the inner 255 loop on which the parameter tuning was performed. To evaluate the performance of algorithms, root-mean-squared 256 error (RMSE), Eq. 34, mean absolute error (MAE), Eq. 42, and mean absolute percentage error (MAPE), Eq. 53, 257 were used-, Furthermore, AIC, Eq. 6, BIC, Eq. 7, and % Bias, Eq. 8, are also included in Table S2 for more detailed 258 comparison.

259
$$RMSE = \sqrt{\frac{1}{n}\sum_{i=1}^{n}(P_i - O_i)^2}$$
 (34)

260
$$MAE = \frac{1}{n} \sum_{i=1}^{n} (P_i - O_i)$$
 (42)

261
$$MAPE = \frac{1}{n} \sum_{i=1}^{n} \left| \frac{P_i - O_i}{O_i} \right| \times 100$$
 (53)

262
$$AIC = -2ln(L) + 2k$$
 (6)
263 $BIC = -2ln(L) + log(n) \frac{\log \log k}{(7)}$ (7)
264 $\%BIAS = \frac{1}{n} \sum_{i=1}^{n} \frac{(P_i - O_i)}{O_i} \times 100$ (8)

265 <u>w</u>Where *n* is the number of samples, *L* is likelihood, *k* is the number of parameters, and P_i and O_i are the predicted

and observed values, respectively. were

267 2.6.1 Parameter tuning

268 As mentioned in Sect. I previously, choosing a suitable strategy for parameter tuning is a crucial step in machine 269 learning particularly for when comparing the performance of the algorithms. Therefore, two strategies were applied 270 depending on the algorithm: 1) a grid search for RF and 2) optimisation with the DE algorithm for BRT and SVR. 271 The first strategy was an exhaustive search over a defined space consisting of lower bound, upper bound and n272 steps in between for the target parameter<u>One major problem with applying the grid search strategy for algorithms</u> 273 that comprise continuous parameters such as BRT and SVR is that it is impossible to consider the whole continuous 274 parameter space. Thus, the parameter combination for testing should be determined. However, this is not 275 problematic for tuning RF in the present case since m_{trv} is a parameter with discrete values. The DE algorithm 276 however, is an stochastic approach to solve an optimisation problem that can be applied to both continuous and 277 discrete parameters. Therefore, the target parameters in this strategy should be discrete or discretised beforehand 278 if they are continuous (Probst et al., 2019). This strategy was applied to RF since the tuning parameter is discrete. 279 However, the second strategy is a stochastic approach of searching over a continuous space in order to solve an 280 optimisation problem (Qin et al., 2009) and. This method is described in more detail by Storn & Price (1997). 281 Therefore, SVR and BRT were are optimised by this strategy as the former algorithm has continuous parameters 282 and the latter one has both continuous and discrete parametersthey have continuous parameters. For the optimisation task in the present study, the R package "DEoptim" was applied (Peterson et al., 2021). Table S1 283 284 shows the parameters and their tuning range for each algorithm.

285 2.6.2 Variable importance

286 Variable importance was assessed by permutation (Ließ et al., 2021). Therefore, tThe values of each a particular 287 covariate in the test set was were shuffled 10 times and on each occasion the prior to applying the respective model 288 to eliminate any predictor-response relationship present with regards to that predictor-trained model corresponding 289 to that test set was applied. The variable importance corresponds to the relative increase in the test set RMSE. This 290 processprocedure was repeated 10 times for each covariate. The resulting values were averaged and T_the 291 population of RMSE was averaged and its to be used for calculation of relative change to the RMSE of the original 292 test set was calculated. Thus, the variable importance of each covariate in terms of percentage relative change in 293 RMSE was obtained.

294 2.7 Modelling approaches

295 Three approaches were designed We followed a two-by-two strategy resulting in four modelling approaches to test 296 the performance of the algorithms (Table 1). The models were built based on nested CV, while the train and test 297 sets remained identical for the three algorithms to make the results comparable.

298Table 1: Modelling approaches

	Dataset 1: German Agricultural Soil Inventory	<u>Dataset 2:</u> <u>German Agricultural Soil Inventory +</u> <u>LUCAS</u>
One-Model-Approach	<u>AP1</u>	AP1L
Two-Model-Approach	<u>AP2</u>	AP2L

On the one hand, we The first approach (AP1) only used the SOC content data from the German Agricultural Soil
 Inventory and corresponding values from the covariates were used to build train the models (AP1). Thus, the
 dataset was cross validated and used by BRT, RF and SVR to predict the SOC content of German agricultural
 soils. The results of this approach served as a baseline on which the model improvement for each algorithm in the
 other two approaches was assessed.

305 Due to the high variability of SOC in the agricultural soils of Germany, we then trained two separate models for 306 organic and mineral soils (AP2) was developed and tested to identify whether it this could improve model 307 performance. Accordingly, the German Agricultural Soil Inventory was subdivided by the threshold 87 g kg⁻¹ into 308 mineral and organic soils, and then two were used to train separate models were trained. This approach was named 309 AP2. The same nested CV procedure was applied for both data subsets. The results of BRT, RF and SVR were 310 compared to identify which one had better performance under mineral and organic soils separately. Finally, each 311 algorithm's predicted SOC values from the two separate models was were combined, and the error metrics were 312 calculated for the full data set to identify the impact of AP2 on model performance. The CV folds for this procedure 313 match the one from the AP1 models.

The impact of enlarging the training set on model performance was <u>then_examined</u> for both, AP1 and AP2 approaches. Thus, 1223 depth-extrapolated samples of the LUCAS data were added to the training sets of AP1. <u>The corresponding modelling approach was and-named AP1L</u>. Moreover, the same threshold (87 g kg⁻¹) was used to subdivide this dataset and each soil class was included to the training set of the corresponding soil class of AP2.

318 <u>This modelling approach was then</u> and named AP2L.

The test sets for the model performance evaluation of the CV procedure remained the same for all four approaches.
 <u>The models were built based on nested CV</u>, while the train and test sets remained identical for the three algorithms
 to make the results comparable. Thus, the dataset was cross-validated and used by BRT, RF and SVR to predict
 the SOC content of German agricultural soils. The results of this the AP1 approach served as a baseline on which

323 the model improvement for each algorithm in the other two approaches werewas assessed.

324 3 Results and Discussion

325 3.1 Comparison of algorithms on the data from the German Agricultural Soil Inventory (AP1)

326 The range of the topsoil SOC content of topsoil for the German Agricultural Soil Inventory dataset was 4 g kg⁻¹ to 327 480 g kg⁻¹, with a mean of 27 g kg⁻¹ and a median of 16 g kg⁻¹. Figure 1 shows the spatial distribution of the 328 implemented data. For the first approach (AP1), BRT, RF, and SVR were applied to model SOC using data from 329 German Agricultural Soil Inventory. The RMSE and MAPE indicated that SVR had a better general performance 330 than the two-other two algorithms (Fig.5 2). In this respect, the RMSE of SVR was 5% lower than that from RF 331 and 4% lower than that from BRT. Furthermore, its MAPE was 3% and 7% lower than that from RF and BRT 332 respectively. However, despite the difference in overall performance, the spatial distribution of relative residuals 333 indicated that all three algorithms were less accurate in the northern of Germany compared with the centre and 334 south of the country (Fig. 3A). This can be explained by the characteristics of this region and its higher SOC 335 variability. The northern part of Germany is-a lowland dominated by a sandy soil texture from pleistoceen 336 sedimentation with geomorphological structures such as ground moraines, terminal moraines and aprons (Roßkopf 337 et al., 2015). Despite general geomorphological and pedological similarities throughout the region, 1) organic soils 338 in under agricultural use Germany are mainly located in the north and 2) mineral soils with the lowest and the

- 339 highest SOC contents are also located in the northeast and northwest respectively. Therefore, this region has the
- 340 <u>highest-widest</u> SOC range-on agricultural soils.



Figure 1: Soil organic carbon content in <u>the</u> topsoil of two soil inventories: A) German Agricultural Soil Inventory (0-30 cm), B) LUCAS at its original sampling depth (0-20 cm) <u>and</u>; C) LUCAS after depth extrapolation (0-30 cm)





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351 Consequently, the variable importance (Fig. 4A) indicated that the map of organic soils contains the highest 352 available information among all was the most important covariates for the algorithms. The value for of the variable 353 importance for this covariate was 65% in SVR, 72% in RF and 84% in BRT. These values firstly show the crucial 354 role of the map of organic soils for the algorithms in explaining the variability of SOC and, secondly, the 355 comparatively greater importance of this predictor and the lower variable importance of other predictors in the 356 BRT model compared with the SVR model.how BRT mainly relies on the map of organic soils to predict SOC 357 compared with SVR. Despite the importance of the organic soil map, the scatterplots (Fig. 5A) show that all three 358 algorithms underpredicted the SOC of the organic soils and had similar heteroscedasticity patterns in their 359 residuals. Thus, while most residuals from mineral soils followed the 1:1 line, they became more scattered in soils 360 with a higher SOC content. The underprediction of SOC in organic soils can be explained by their low-small 361 sample size, resulting in a dataset with a high-wide SOC range and a unimodal distribution that leaves these soils 362 in the tail. Consequently, the organic soils were underrepresented and the results were systematically pulled 363 towards mineral soils, regardless-irrespective of the choice of algorithm. Different studies have shown that

Figure 2: Performance indicators of the three algorithms. One-model approach (<u>w</u>Without LUCAS data AP1 and with LUCAS data AP1L) versus the two-model approach (AP2 and AP2L) for A) RMSE (g kg⁻¹), B) MAE (g kg⁻¹) and C) MAPE (%). The whiskers of boxplots show 1.5 times the interquartile range. Please note that the y-axis is shortened for better visibility and does not display a zero. <u>BRT = boosted regression trees</u>, RF = random forest, and SVR = support vector regression.

- predicting soil properties with mineral and organic soils combined can lead to underprediction or overprediction
 of one soil class, depending on the distribution of the dataset (Brogniez et al., 2015; Guio Blanco et al., 2018;
- **366** Mulder et al., 2016).

367 Although the map of organic soils was able to distinguish between the two soil classes, i.e. between mineral and 368 organic soil, it could not separate the mineral soils with a low SOC content in the northeast from those with a high

SOC content in the northwest. The spatial distribution of the residuals (Fig. 6A) showeds that SVR and BRT

generally underpredicted the mineral soils in the northwest part of Germany, while RF overpredicted them.

- 371 Furthermore, unlike RF and SVR, BRT distinctively appreciably overpredicted SOC of the north-east Germany's
- mineral soils with the lowest SOC content ($<10 \text{ g kg}^{-1}$). This result indicates that the algorithms differed in their
- performance in mineral soils. This difference was mainly due to the information they obtained from the land use
 map. As the second most important covariate for all three algorithms (Fig. 4 A), the value for variable importance
- for this covariate was 22% in SVR, but just 11% in RF and 9% in BRT. Thus, SVR exploits more information
- 376 from this covariate than RF and particularly BRT. Land use is one of the main drivers of SOC variability on a
- 377 national scale due to the higher SOC content in grasslands than in croplands (Poeplau et al., 2020). Therefore, this
- 378 covariate was able to differentiate between the soils of the northeast, which are under cropland, and those in the
- 379 northwest as they are more under grassland. Consequently, the reliance of BRT on the map of organic soils at the
- 380 <u>cost expense</u> of land use could explain why this algorithm overpredicted SOC in croplands in the northeast.



381

Soil class Mineral Organic Relative residuals (%) More than 100 • 50 to 100 • 5 to 50 • Less than 5 •

Figure 3: Spatial distribution of relative residuals. A) AP1 approach, B) AP1L approach, C) AP2 approach and D)
 AP2L approach. <u>BRT = boosted regression trees, RF = random forest, and SVR = support vector regression.</u>



Figure 4: Variable importance in terms of average relative change (%) in RMSE. A) AP1, B) mineral soil subset of
 AP2 and C) organic soil subset of AP2. The full name for each abbreviation is presented in Table S43. BRT = boosted
 regression trees, RF = random forest, and SVR = support vector regression.

388 3.2 Enlarging the dataset with additional soil inventories (AP1L)

A larger soil dataset may provide additional information and consequently improve model performance. This possibility was explored in the AP1L approach with-by adding the-LUCAS data. The SOC content of LUCAS data at its original depth ranged from 4 g kg⁻¹ to 500 g kg⁻¹-, with a mean of 30 g kg⁻¹ and a median of 18 g kg⁻¹. After extrapolating the depth to 30 cm, the new range was from 5 g kg⁻¹ to 512 g kg⁻¹-, with a mean of 28 g kg⁻¹ and a median of 17 g kg⁻¹. The spatial distribution of LUCAS data at their original and extrapolated depth is shown in Figure 1.

395 A statistical test was performed on the residuals of models built on LUCAS data with the original and extrapolated 396 depths. That was done to identify whether extrapolating the depth of LUCAS data to that of the German 397 Agricultural Soil Inventory would significantly affect model performance after their inclusion in the training set. 398 With the Shapiro-Wilk test rejecting the normality assumption of residuals of all corresponding algorithms at 20 399 cm and 30 cm, the non-parametric Kruskal-Wallis test showed no significant difference between the residuals at 400 both either depths. Thus, the extrapolation of the soil depth had no significant impact on the data quality to 401 regionalisze SOC. As a result, any further change in the performance of the algorithms after adding LUCAS data 402 was due to enlargement of the training set being enlarged. The result of the algorithms at both depths can be found 403 in the supplementary information (Fig. $S_{\underline{34}}$).

404 After enlarging the training set from 2278 to 3501 sampling points, BRT obtained the lowest RMSE (Fig. 2A1) 405 and MAE among the algorithms (Fig. 2B1). A comparison of the error metrics of corresponding algorithms from 406 the AP1 approach with those from the AP1L approach showed that BRT had the highest error reduction at 7% in 407 the MAPE and 5% in the RMSE and MAE. Furthermore, although the error metrics of RF did not improve as 408 much as those of BRT, additional training points were still beneficial for this algorithm. However, SVR did not 409 follow any systematic change under the AP1L. Despite a 2% decrease in MAPE, the RMSE increased by 3% and 410 MAE remained unchanged. To explore the potential explanation for this behaviour by SVR, the residuals of 411 mineral soils were separated from those of organic soils. Additional samples reduced the RMSE in mineral soils for all algorithms by between 9% and 13%. However, this error increased by 9% in the organic subset for SVR, 412 413 while it increased by just 1% for RF and even decreased by 1% for BRT. This indicated that enlarging the training 414 set by data with similar characteristics had a greater influence on systematic error of the underrepresented soil 415 class in SVR. This influence is understandable when considering the higher optimised $\boldsymbol{\varepsilon}$ in the AP1L approach 416 compared with that of the AP1 approach. The higher value of $\boldsymbol{\varepsilon}$ means that the hyperplane for the training set is 417 less complex (Cherkassky and Ma, 2004) and more suitable for predicting most soil samples, i.e. mineral soils. 418 Thus, when this hyperplane was fitted to the test set identical to the AP1, the generalisation performance was 419 hindered because it could not capture the variability of samples with higher SOC values, i.e. organic soils.

420 Further evaluation revealed that regardless of the change in error metrics, the relative residuals of the three
421 algorithms had a similar spatial pattern to their counterpart from the AP1. Thus, they all showed lower accuracy

422 in the northern region of Germany for similar reasons (Fig. 3B). Moreover, the scatterplots had a similar pattern
423 with underpredicted organic soils (Fig. 5B). This confirmeds that when organic soils are modelled with mineral

- 424 soils, enlarging the training set does not provide enough information for BRT or RF to capture the high variability
- 425 of SOC, particularly in the north of Germany.



Figure 5: Scatterplot of residuals. A) AP1 approach and mineral and organic soils of AP2 and B) AP1L approach and mineral and organic soils of AP2L. BRT = boosted regression trees, RF = random forest, and SVR = support vector

428 mineral and 429 <u>regression.</u>





Soil class Mineral ■Organic▲ Residuals (g kg-1) More than 50 ● 0 to 5 ● -5 to 0 ● Less than -50 ● 5 to 0 ● Less than -50 ●

431 Figure 6: Spatial distribution of residuals. A) AP1 approach, B) AP1L approach, C) AP2 approach and D) AP2L
432 approach. <u>BRT = boosted regression trees</u>, <u>RF = random forest</u>, <u>and SVR = support vector regression</u>.

433 3.3 Subdividing soil inventories into mineral and organic subsets (AP2 and AP2L)

434 As presented-outlined in the sections above, the modelling of SOC content when mineral and organic soils were 435 combined led to a systematic underprediction of soils with higher SOC values by all three algorithms, regardless 436 irrespective of the number of training samples. Therefore, by implementing the AP2 approach with two models 437 one for mineral soils and one for organic soils, a noticeable improvement in the performance of all algorithms was 438 observed, observed (Table S3B), with SVR showing the best error metrics (Fig. 2A6, Fig. 2B6, Fig. 2C6). This 439 meant 34% lower RMSE, 30% lower MAE, and 32% lower MAPE than when this algorithm was trained under 440 the AP1 approach with one model for all soils. As the high variability of SOC was initially hard to capture, the 441 subdivision of the dataset provided a range that better represented each soil class. This was particularly beneficial 442 for mineral soils (ranging from 4 g kg⁻¹ to 85 g kg⁻¹) since the number of samples did not reduce drastically (only 443 by 99 samples). Thus, the algorithms could better capture the relationship between SOC and covariates. 444 Consequently, the overall performance improved when the underrepresented soil class was modelled separately. 445 This is in line with the study of Rawlins et al. (2009) which that recommends the separate modelling of mineral 446 and organic soils.

447 Nonetheless, following the AP2L approach with additional data, the RMSE and MAPE of the algorithms improved 448 by less than 2% compared with AP2 (Table S3E). However, the greatest change was observed in the MAE of SVR 449 with a 2% improvement. Therefore, additional training samples did not considerably greatly influence the 450 performance since the majority of these samples were in mineral soils, while the limiting factor was the high 451 variability of organic soils combined with its low number of samples. NeverthelessHowever, an improvement was 452 noted in relation to the all error metrics of SVR in the AP2L approach. This was in contrasted to with when the 453 training set was enlarged without subdividing the data, i.e. AP1L. Therefore, it further confirmed that it is more 454 important for SVR than for BRT and RF to model the soil classes separately when its the training set is enlarged 455 by datasets with similar characteristics.

- 456 Furthermore, the improvement of the algorithms in AP2 and AP2L was particularly noticeable in their relative 457 residuals. By comparing these results with those from AP1 and AP1L, it was evident that the greatest improvement 458 was observed in the northern region and the spatial distribution of relative residuals was more homogenous 459 throughout the country for all algorithms, but particularly for RF and SVR (Fig. 3 C and D). This is understandable 460 since by subdividing the data, the algorithms can no longer exploit any information from the map of organic soil 461 for spatial variability of SOC in mineral soils. Thus, they obtain information from other covariates for this soil 462 class (Fig. 4 B). Although land use and total nitrogen were still among the most important variables for the 463 algorithms in mineral soils, the importance of the predictors representing the SCORPAN C and P factors increased
- in the absence of a soil organic map. This <u>could_was to</u> be expected because <u>the</u>-north-east <u>of</u>-Germany, for
 example, has <u>a</u> continental climate (Roßkopf et al., 2015) and young moraine landscapes, while the north-west has
- a more oceanic climate (Roßkopf et al., 2015) with old moraine landscapes.
- 467 It is unsurprising that all the algorithms still relied on the map of organic soil to explain SOC in organic soil class.
 468 However, while SVR and RF still-obtained information from other covariates, the value for variable importance
- 469 of this map alone is-was 93% in BRT (Fig. 4 C). That which makes this algorithm prone to greater errors, as can
- 470 be seen in its error metrics (Table S2). Similar to mineral soils, the order of covariates was different between the
- algorithms in organic soils. In other words, in AP1 the three algorithms obtained almost all <u>the</u> information from
- 472 the map of organic soil, land_-use and total nitrogen with-in that similar-order_of importance. In contrast, after

subdividing the data, the algorithms differe<u>d</u>ntiated from each other by the order of covariates in their variable
importance (Fig<u>ure</u> 4).

A comparison of the error metrics of each soil class in AP2 with its counterpart in AP2L revealed that the additional 1177 samples had a minor influence on the performance (from zero to a maximum of 2%) of the algorithms in mineral soils (Table S2). These results indicated that the German Agricultural Soil Inventory offers a good representation of the spatial variability of SOC in mineral soil under agricultural use throughout the country and that the inclusion of including-more sample points do did not provide additional information about SOC variability in this soil class.

481 However, 46 additional organic soil samples from the LUCAS dataset improved the MAPE and MAE by 12% and 482 6% for SVR, by 10%, and 4% for RF, and by 7% and 2% for BRT, respectively, but the RMSE of the three algorithms was improved by less than 2%. Thus, additional organic samples mainly influenced the average 483 484 magnitude of the error. This could be explained by organic soils having a wide range of SOC and the number of 485 samples was being limited. Thus, the addition of LUCAS data to the training set offered gave the algorithms more 486 information about spatial variability of SOC in this soil class. Despite this limitation, SVR had the best overall 487 performance among the algorithms in AP2 and AP2L. It should be noted that training samples must span the 488 complexity of the parameter space in order for the model to be able to effectively match the training data effectively 489 and to-generalisze unseen data. A ssmall sample size can therefore negatively influence the predictive power of 490 the algorithms. This complexity can be addressed by structural risk minimisation (SRM) (Al-Anazi and Gates, 491 2012). Implementation of SRM makes SVR capable of performing well in such datasets. Other studies have 492 compared the performance of algorithms on different sample sizes for in predicting soil properties and shown that 493 SVR is one of the best choices, if not the best, when the number of samples is a limiting factor (Al-Anazi and 494 Gates, 2012; Khaledian and Miller, 2020). In contrast, in a study by Zhou et al. (2021), 150 samples with different 495 sets of covariates at different resolutions were used to compare RF, BRT and SVR to predict SOC content in 496 Switzerland. Their results showed that the best-performing algorithm varied depending on the resolution and 497 covariates. However, the best performance throughout all scenarios was obtained by BRT. The discrepancy 498 between their results and the results of the present study may be due to the parameter-tuning method of the 499 algorithms, as they only used grid search, or other factors, including the spatial distribution of samples or the 500 chosen set of covariates.

Anneach	Mean RMSE	Mean MAE	Mean MAPE
Approach	<u>(g kg⁻¹)</u>	<u>(g kg⁻¹)</u>	<u>(%)</u>
<u>AP1</u>	<u>32.6</u>	<u>12.3</u>	<u>49.0</u>
<u>AP1L</u>	<u>32.1</u>	<u>12.1</u>	<u>46.9</u>
<u>AP2</u>	<u>21.6</u>	<u>8.8</u>	<u>34.4</u>
<u>AP2L</u>	<u>21.3</u>	<u>8.7</u>	<u>34.3</u>

501 <u>Table 2: Mean of error metrics of the three models for each approach.</u>

502 Overall, the change in performance across different sample sizes, different algorithms and different approaches
 503 (Table S3) indicated that the most important aspect of modeling SOC content of German agricultural topsoil is a

504 two-model approach. Although combining soil inventories for more training samples can possibly improve model

505 performance, the effect was not noticeable compared to when each soil class was predicted by its dedicated model

506 (Table S3B and Table S3D). The advantage of two-model approach can also be seen in the average error metrics
507 of the three models (Table 2). While the average RMSE of the models reduces by less than 1 g kg⁻¹ after enlarging
508 the training set, the same error metrics reduces by more than 10 g kg⁻¹ in AP2 and AP2L (Table 2). Therefore, it
509 is also recommended to consider the two-model approach in soil-landscape settings similar to Germany or
510 situations where one-model approach cannot have good predictive performance.

- 511 The map of organic soil was used to spatially distinguish each soil class to map the SOC content of the class by its 512 corresponding model. Figure S5 shows the spatial distribution of SOC content using the AP2L approach for the 513 three algorithms. Although SVR captured a wider range of SOC, 2 g kg⁻¹ to 371.5 g kg⁻¹, than BRT, 8 g kg⁻¹ to 514 341.1 g kg⁻¹, and RF, 7.7 g kg⁻¹ to 354.6 g kg⁻¹, all three algorithms showed a relatively similar distribution of SOC 515 content across the country particularly in mineral soils. As shown in Figure S5, organic soils are mainly distributed 516 in the north. These soils are mostly bogs in the northwest and fens in the northeast (Roßkopf et al., 2015). There is also a small distribution of organic soil in the foothills of Alps in the south. In mineral soils, a higher SOC 517 518 content is mainly found in northwest and south of the country. As explained in the previous sections, one of the
- 519 main reasons for this distribution is land use since these regions are mainly under grassland while low SOC content
- 520 <u>regions are found under cropland.</u>

521 4 Conclusions

522 The three most commonly used algorithms most commonly used in DSM were implemented applied to predict the 523 SOC content of German agricultural soils under different approaches. Suitable tuning strategies for each algorithm 524 ensured optimum parameter tuning and made their performance truly comparable. Machine learning algorithms 525 was shown to be powerful in at modelling SOC on a national scale. However, the study showed that separate 526 modelling of mineral and organic soils was a better approach for modelling SOC compared to with using just one 527 model. Thus, this approach has takes priority to over the choice of algorithm and number of training samples. We 528 recommendFurther testing of this approach to be further tested is recommended in countries and regions that cover 529 both of these soil classes. Nonetheless, SVR had <u>a</u> better performance than RF and BRT, except when the number 530 of samples in training was increased by additional dataset. This was disadvantageous for SVR and advantageous 531 for BRT unless mineral and organic soils were modelled separately. In general, increasing the number of training 532 samples led to limited improvement of performance. Therefore, this approach should be done-adopted with-giving 533 consideration of the algorithm and the characteristics of the data. Furthermore, the better performance of SVR over 534 compared with that of RF and BRT was particularly highlighted when predicting SOC in organic soils. Thus, this 535 The good performance of algorithm-SVR suggests that this algorithm should therefore be taken into greater account 536 in DSM-when the number of samples is limited.

537 Data availability

- 538 The soil data used in this study are publicly available via: <u>https://doi.org/10.3220/DATA20200203151139</u> and
- 539 <u>https://esdac.jrc.ec.europa.eu/content/lucas-2009-topsoil-data</u>

540 Author contribution

- 541 AS and AD conceptualised and developed the methodology of the presented work, with input from ML.AS
- 542 gathered the predictors with contributions from AD. AS executed the programming, testing of existing code

- 543 components, formal analysis and visualizationvisualisation. AG contributed to the programming. The preparation
- of the paper was done by all authors.

545 Competing interests

- 546 The authors declare that they have no conflict of interest except the author AD is a member of the journal's
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- 1 Spatial prediction of organic carbon in German agricultural topsoil
- 2 using machine learning algorithms Performance of three machine
- 3 learning algorithms for predicting soil organic carbon in German
- 4 agricultural soil
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10 Supplements

- 11 Table S1: The range of parameters for tuning in full dataset (AP1 and AP1L) and mineral and organic soil subsets
- 12 (AP2 and AP2L)Tuning parameter ranges corresponding to the models trained by different algorithms, The ranges
- 13 differ considering the one-model (full dataset, AP1 & AP1L) or two-model approach (mineral and organic data
- 14 subset, AP2 & AP2L). BRT = boosted regression trees, RF = random forest, and SVR = support vector regression.

<u>Algorithm</u>	<u>Tuning</u> pParameter	Full dataset	Mineral soil <u>data</u> <u>subset</u>	Organic soil <u>data</u> <u>subset</u>
	С	1-100	1-50	1-200
SVR	epsilon	0-5	0-1	0-5
	gamma	0.001-1	0.001-1	0.001-1
RF	mtry	3-13	3-13	3-13
BRT	number of trees	100-3000	100-3000	100-3000
	shrinkage	0.001-0.1	0.001-0.1	0.001-0.1
	interaction depth	1-5	1-5	1-5
	bag fraction	0.5-0.9	0.5-0.9	0.5-0.9

15 Table S2: Error metrics of the algorithmsPredictive model performance of the models trained with different machine 16 learning algorithms and datasets: A) built on the German Agricultural Soil Inventory, B) including LUCAS data in 17 the training set. BRT = boosted regression trees, RF = random forest, and SVR = support vector regression.

	Algorithm	<u>RMSE</u>	MAE	<u>%MAPE</u>	<u>%Bias</u>	<u>AIC</u>	BIC	<u>Approach</u>
	<u>BRT</u>	<u>32.9</u>	<u>12.4</u>	<u>50.9</u>	<u>-32</u>	<u>14865</u>	<u>14889</u>	<u>AP1</u>
	<u>RF</u>	<u>33.2</u>	<u>12.3</u>	<u>48.6</u>	<u>-30</u>	<u>14913</u>	<u>14919</u>	<u>AP1</u>
	<u>SVR</u>	<u>31.6</u>	<u>12.3</u>	<u>47.4</u>	<u>-20</u>	<u>14643</u>	<u>14661</u>	<u>AP1</u>
	BRT	<u>9.5</u>	<u>6.2</u>	<u>35.9</u>	<u>-20</u>	<u>7500</u>	7524	Mineral
	<u>RF</u>	<u>9.1</u>	<u>5.9</u>	<u>34</u>	<u>-20</u>	<u>7288</u>	<u>7294</u>	<u>Mineral</u>
	<u>SVR</u>	<u>9.2</u>	<u>5.8</u>	<u>31.8</u>	<u>-10</u>	<u>7331</u>	<u>7349</u>	<u>Mineral</u>
A	<u>BRT</u>	<u>107</u>	<u>90.4</u>	<u>48.5</u>	<u>-26</u>	<u>757</u>	<u>768</u>	<u>Organic</u>
	<u>RF</u>	<u>106.1</u>	<u>89.3</u>	<u>48.2</u>	<u>-28</u>	<u>750</u>	<u>753</u>	<u>Organic</u>
	<u>SVR</u>	<u>101.7</u>	<u>86.9</u>	<u>45.6</u>	<u>-22</u>	<u>746</u>	<u>754</u>	<u>Organic</u>
	<u>BRT</u>	<u>22</u>	<u>9.1</u>	<u>36.3</u>	<u>-20</u>	<u>12578</u>	<u>12602</u>	<u>AP2</u>
	<u>RF</u>	<u>21.7</u>	<u>8.8</u>	<u>34.5</u>	<u>-20</u>	<u>12496</u>	<u>12502</u>	<u>AP2</u>
	<u>SVR</u>	<u>21</u>	<u>8.6</u>	<u>32.3</u>	<u>-10</u>	<u>12310</u>	<u>12328</u>	<u>AP2</u>

	<u>Algorithm</u>	<u>RMSE</u>	<u>MAE</u>	<u>%MAPE</u>	<u>%Bias</u>	<u>AIC</u>	BIC	<u>Approach</u>
	<u>BRT</u>	<u>31.3</u>	<u>11.8</u>	<u>47.4</u>	<u>-30</u>	<u>14568</u>	<u>14592</u>	<u>AP1L</u>
	<u>RF</u>	<u>32.5</u>	<u>12.1</u>	<u>46.8</u>	<u>-30</u>	<u>14754</u>	<u>14759</u>	<u>AP1L</u>
	<u>SVR</u>	<u>32.6</u>	<u>12.3</u>	<u>46.4</u>	<u>-20</u>	<u>14775</u>	<u>14792</u>	<u>AP1L</u>
<u>B</u>	<u>BRT</u>	<u>9.4</u>	<u>6.2</u>	<u>35.6</u>	<u>-20</u>	<u>7429</u>	<u>7453</u>	Mineral
	<u>RF</u>	<u>9.1</u>	<u>6</u>	<u>34.6</u>	<u>-20</u>	<u>7268</u>	<u>7274</u>	<u>Mineral</u>
	<u>SVR</u>	<u>9.1</u>	<u>5.8</u>	<u>31.7</u>	<u>-10</u>	<u>7275</u>	<u>7293</u>	Mineral
	<u>BRT</u>	<u>105.4</u>	<u>88.4</u>	<u>45</u>	<u>-20</u>	<u>754</u>	<u>765</u>	<u>Organic</u>
	<u>RF</u>	<u>104.1</u>	<u>86.2</u>	<u>43.5</u>	-20	<u>745</u>	<u>748</u>	<u>Organic</u>
	<u>SVR</u>	<u>100.2</u>	<u>81.7</u>	<u>40.2</u>	<u>-12</u>	<u>741</u>	<u>749</u>	<u>Organic</u>
	<u>BRT</u>	<u>21.7</u>	<u>9</u>	<u>36</u>	<u>-20</u>	<u>12486</u>	<u>12510</u>	<u>AP2L</u>
	RF	<u>21.4</u>	<u>8.7</u>	<u>34.9</u>	<u>-20</u>	<u>12379</u>	<u>12385</u>	<u>AP2L</u>
	<u>SVR</u>	<u>20.7</u>	<u>8.4</u>	<u>31.9</u>	<u>-10</u>	<u>12191</u>	<u>12209</u>	<u>AP2L</u>

19Table S3: Percent change in predictive model performance comparing models trained with different machine learning20algorithms and data sets: A) and B) comparison of models trained by using data from the –German Agricultural Soil21Inventory, C) and D) comparison of models trained by using data from the German Agricultural Soil Inventory and22LUCAS, A) and C) comparison of models trained withcomparison with regards to the different machine learning23algorithms, B) and D) comparison of the one-model approach (AP1) to the two-model approach (AP2), E) comparison24of the approaches before and after including LUCAS, BRT = boosted regression trees, RF = random forest, and SVR =25support vector regression.

	<u>Algorithm</u>	<u>RMSE</u> (%)	<u>MAE (%)</u>	<u>MAPE (%)</u>	Approach
	BRT to RF	<u>0.9</u>	<u>-0.8</u>	<u>-4.5</u>	<u>AP1</u>
	<u>RF to SVR</u>	<u>-4.8</u>	<u>0.0</u>	<u>-2.5</u>	<u>AP1</u>
	BRT to SVR	-4.0	<u>-0.8</u>	<u>-6.9</u>	<u>AP1</u>
	BRT to RF	<u>-4.2</u>	<u>-4.8</u>	<u>-5.3</u>	<u>Mineral</u>
	RF to SVR	<u>1.1</u>	<u>-1.7</u>	<u>-6.5</u>	<u>Mineral</u>
٨	BRT to SVR	<u>-3.2</u>	<u>-6.5</u>	<u>-11.4</u>	Mineral
A	BRT to RF	<u>-0.8</u>	<u>-1.2</u>	<u>-0.6</u>	<u>Organic</u>
	RF to SVR	<u>-4.1</u>	<u>-2.7</u>	<u>-5.4</u>	<u>Organic</u>
	BRT to SVR	<u>-5.2</u>	<u>-4.0</u>	<u>-6.0</u>	<u>Organic</u>
	BRT to RF	<u>-1.4</u>	<u>-3.3</u>	<u>-5.0</u>	<u>AP2</u>
	<u>RF to SVR</u>	<u>-3.2</u>	<u>-2.3</u>	<u>-6.4</u>	<u>AP2</u>
	BRT to SVR	<u>-4.5</u>	<u>-5.5</u>	<u>-11.0</u>	<u>AP2</u>
	<u>Algorithm</u>	<u>RMSE</u>	MAE	<u>MAPE</u>	<u>Approach</u>
<u>B</u>	BRT	<u>-33.1</u>	<u>-26.6</u>	<u>-28.7</u>	AP1 to AP2
	RF	<u>-34.6</u>	<u>-28.5</u>	<u>-29.0</u>	AP1 to AP2
	<u>SVR</u>	-33.5	<u>-30.1</u>	-31.9	AP1 to AP2
	<u>Algorithm</u>	RMSE	MAE	MAPE	Approach

	BRT to RF	<u>3.8</u>	<u>2.5</u>	<u>-1.3</u>	AP1L
	<u>RF to SVR</u>	<u>0.3</u>	<u>1.7</u>	<u>-0.9</u>	<u>AP1L</u>
	BRT to SVR	<u>4.2</u>	<u>4.2</u>	<u>-2.1</u>	<u>AP1L</u>
	BRT to RF	<u>-3.2</u>	-3.2	<u>-2.8</u>	Mineral
	RF to SVR	<u>0.0</u>	<u>-3.3</u>	<u>-8.4</u>	Mineral
C	BRT to SVR	-3.2	<u>-6.5</u>	<u>-11.0</u>	Mineral
<u>C</u>	BRT to RF	<u>-1.2</u>	<u>-2.5</u>	<u>-3.3</u>	<u>Organic</u>
	RF to SVR	<u>-3.7</u>	<u>-5.2</u>	<u>-7.6</u>	<u>Organic</u>
	BRT to SVR	<u>-5.2</u>	<u>-8.2</u>	<u>-10.7</u>	<u>Organic</u>
	BRT to RF	<u>-1.4</u>	<u>-3.3</u>	<u>-3.1</u>	<u>AP2L</u>
	RF to SVR	<u>-3.3</u>	<u>-3.4</u>	<u>-8.6</u>	<u>AP2L</u>
	BRT to SVR	<u>-4.6</u>	<u>-6.7</u>	<u>-11.4</u>	<u>AP2L</u>
	<u>Algorithm</u>	<u>RMSE</u>	MAE	MAPE	<u>Approach</u>
	<u>BRT</u>	<u>-30.7</u>	<u>-23.7</u>	<u>-24.1</u>	AP1L to AP2L
<u>D</u>	<u>RF</u>	<u>-34.2</u>	<u>-28.1</u>	<u>-25.4</u>	AP1L to AP2L
	<u>SVR</u>	<u>-36.5</u>	<u>-31.7</u>	<u>-31.3</u>	AP1L to AP2L
	<u>Algorithm</u>	<u>RMSE</u>	MAE	<u>MAPE</u>	<u>Approach</u>
	BRT	<u>-4.9</u>	<u>-4.8</u>	<u>-6.9</u>	AP1 to AP1L
	<u>RF</u>	<u>-2.1</u>	<u>-1.6</u>	<u>-3.7</u>	AP1 to AP1L
	<u>SVR</u>	<u>3.2</u>	<u>0.0</u>	<u>-2.1</u>	AP1 to AP1L
	<u>BRT</u>	<u>-1.1</u>	<u>0.0</u>	<u>-0.8</u>	Mineral
	<u>RF</u>	<u>0.0</u>	<u>1.7</u>	<u>1.8</u>	Mineral
F	<u>SVR</u>	<u>-1.1</u>	<u>0.0</u>	<u>-0.3</u>	Mineral
<u> </u>	BRT	<u>-1.5</u>	<u>-2.2</u>	<u>-7.2</u>	Organic
	RF	<u>-1.9</u>	<u>-3.5</u>	<u>-9.8</u>	<u>Organic</u>
	<u>SVR</u>	<u>-1.5</u>	<u>-6.0</u>	<u>-11.8</u>	<u>Organic</u>
	BRT	<u>-1.4</u>	<u>-1.1</u>	<u>-0.8</u>	AP2 to AP2L
	RF	<u>-1.4</u>	<u>-1.1</u>	<u>1.2</u>	AP2 to AP2L
	<u>SVR</u>	<u>-1.4</u>	-2.3	-1.2	AP2 to AP2L

27 Table S<u>4</u>3: List of covariates, their abbreviations and **reference**<u>their SCORPAN ID</u>.

SCORPAN ID	Covariates	Abbreviation
S	Net erosion	Net-Ero
	Available water capacity	AWC
	Total nitrogen	TN
	pH	pН
	Soil organic map	Peat

	Clay content	Clay		
	Multi-annual grid of annual sunshine duration over Germany	Sun-Dur		
	Multi-annual grids of number of summer days over Germany			
С	Multi-annual grids of monthly averaged daily minimum air temperature (2m) over Germany	Min-temp		
	Multi-annual grids of precipitation height over Germany	Precip		
0	Landuse	DLM		
	Digital elevation model	EU-DEM		
	Slope	Slope		
	Aspect north south direction	Aspect-NS		
р	Aspect east west direction	Aspect-EW		
К	Plan Curvature	Plan-Curv		
	Profile curvature	Prof-Curv		
	Topographic wetness index	TWI		
	Geomorphographic map	GMK		
	Large-scale landscape unit map (Bodengrosslandschaft)	BGL		
Р	Large-scale soil climate region map (Bodenklima)	Bod-klim		
	Hydrological unit	HUK-HE		
N	X coordination	X		
1N	Y coordination	У		



Figure S1: Selected covariates: Sun-Dur) sunshine duration (DWD, 2017), Summ-D) summer days (DWD, 2018b), Min temp) minimum temperature (DWD, 2018a), Precip) precipitation (DWD, 2018c), EU-DEM) digital elevation model
 (European Union Copernicus Land Monitoring Service, 2016), Net-Ero) net soil erosion and deposition rates (Borrelli
 et al., 2018), AWC) available water capacity (Ballabio et al., 2016), N) total nitrogen (Ballabio et al., 2019), pH) map of
 pH (Ballabio et al., 2019), %Clay) % Clay (Ballabio et al., 2016), BGL) soil scapes unit (BGR, 2008) [Legend], Bod-

- Klim) soil-climate region (Roßberg et al., 2007), HUK-HE) hydrogeological unit of hydrogeological map-(BGR, SDG, 2019), GMK) geomorphographic map of Germany (BGR, 2007) [Legend], DLM) Land use (BKG, 2019), Peat) Organic soils (Roßkopf et al., 2015). 35 36





Figure S34: Boxplots comparing algorithm-model performance with regards to the three machine learning algorithms
 considering- LUCAS at the original sampling depth (20 cm) versus LUCAS with depth extrapolated (30 cm): A) RMSE
 (g kg⁻¹), B) MAE (g kg⁻¹) and C) MAPE (%). <u>BRT = boosted regression trees, RF = random forest, and SVR = support</u>
 vector regression.



47 Figure S4: Spatial distribution of relative residuals from the models trained with the different machine learning
 48 algorithms, A) AP1 approach, B) AP1L approach, C) AP2 approach and D) AP2L approach. BRT = boosted
 49 regression trees, RF = random forest, and SVR = support vector regression.



Figure S5: Spatial prediction of SOC content (g kg-1) of German agricultural soils based on the two-model approach for the three algorithms (BRT AP2L, RF AP2L, SVR AP2L). BRT = boosted regression trees, RF = random forest, and SVR = support vector regression. 53 54

- 55 Disclaimer: It is important to note that the provided spatial prediction of SOC content must not be used to identify
- 56 the organic soils of Germany or to determine their spatial distribution. One reason is low sample size of organic
- 57 soils and the systematic underestimation of their SOC content, which leads to an underestimation of their spatial
- 58 extent. Furthermore, organic soils might have been mixed with mineral soil, i.e. due to deep ploughing, or feature
 59 a mineral soil cover. Thus, organic soils might be present despite having a mineral topsoils. Therefore, this study
- 60 cannot nor intents to identify or classify organic soils.