



Switch of fungal to bacterial degradation in natural, drained and rewetted oligotrophic peatlands reflected in d15N and fatty acid composition

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Abstract. During the last centuries major parts of European peatlands were degraded along with drainage and land use changes. Peatland biodiversity and essential ecosystem functions (e.g. flood prevention, groundwater purification and CO2 sink) were dramatically impaired. Moreover, climate change threatens peatlands in the near future. Increasing pressure to

- 15 peatland ecosystems calls for a more cost-efficient method to indicate the current state of peatlands and the success of restoration effort. Metabolism processes in peatland soils are imprinted in stable isotope signatures due to differences in microorganism communities and their metabolic pathways. Therefore we hypothesize that depth profiles of nitrogen stable isotope values provide a promising opportunity to detect peatland decomposition or restoration. We studied five peatlands: Degerö Stormyr (Northern Sweden), Lakkasuo (Central Finland) and three mires in the Black Forest (Southern Germany).
- 20 At all locations cores were taken from adjacent drained (or rewetted) and natural sites to identify $\partial 15N$ trends that could indicate changes due to drainage and restoration. At all drained (and rewetted) sites we found a distinct peak ("turning point") of the $\partial 15N$ values in the center of the drained horizon. To verify our interpretation $\partial 13C$, the C/N ratio and the bulk density were measured and a microscopic analysis of the macro residuals in the peat cores was made. In addition we did a phospholipid fatty acid (PLFAs) analysis to link our results to microbial community composition. We distinguished between
- 25 fungal and bacterial-derived PLFAs. In accordance with other studies, our results suggest, that fungi dominate the microbial metabolism in the upper, aerobic peat horizon. This is reflected by depleted ∂15N values. Downwards the drained horizon conditions slowly switch to oxygen limitation. In consequence fungal-derived PLFAs decreases whereas bacterial-derived PLFAs are rising. The highest diversity of microbial-derived PLFAs is indicated by the ∂15N turning point. Below the ∂15N turning point, oxygen is increasingly limited and concentrations of all microbial-derived PLFAs are decreasing down to the
- 30 onset of the permanently waterlogged, anaerobic horizon. Peatland cores with restoration success show, above the formerly







drainage-affected horizon, again no depth trend of the isotopic values. Hence, we conclude that $\partial 15N$ stable isotope values reflect microbial community composition, which differ between drained and natural peatlands.

1 Introduction

Even though peatlands cover only 3-4 % of the Earth's land surface (Leifeld and Menichetti, 2018), they act as an enormous

- 35 sink for greenhouse gases in natural conditions (Yu et al., 2011; Joosten, 2008). In Europe 70% of the peatlands are currently degraded (Joosten and Couwenberg, 2001). These degrading peatlands account for five percent of the anthropogenic CO2 emission (Leifeld and Menichetti, 2018; Zedler and Kercher 2005). Despite this dramatic peat decline, we lack reliable and transferable tools providing time- and cost-efficient information of peatland condition. We hypothesize that nitrogen (N) stable isotopes could serve as such a tool.
- 40 In natural peatlands most biological metabolism and nutrient cycling takes place in the acrotelm (uppermost aerobic peat horizon with living vegetation) (Asada et al., 2005; Artz, 2013). In the water-saturated catotelm (deeper, anaerobic horizon) organic substrates are decomposed at much smaller rates owing to anoxic conditions and low pH values (Asada et al., 2005; Artz, 2013; Lin et al., 2014). In the mesotelm, a peat horizon situated between acrotelm and catotelm, water table levels and oxygen content constantly fluctuate, resulting in shifting oxic and anoxic conditions and shifting metabolism processes
- 45 (Asada et al., 2005; Artz, 2013; Lin et al., 2014). Clymo and Bryant (2008) therefore defined the mesotelm as a "transition horizon".

Stable C and N isotopes are correlated with vegetation composition and microbial decomposition processes. As decomposition induces an enrichment of heavy isotopes (15N, 13C), vegetation is mostly more depleted of 15N and 13C than microbial and recycled substrate. Alewell et al. (2011) and Krüger et al. (2014) reported distinct changes in ∂ 13C values

- 50 for palsa peat with the onset of decomposition of hummocks and various authors observed the same trend with decomposition in peatlands of other climate conditions (Krüger et al., 2016a; Novak et al, 1999; Hobbie et al., 2017; Biester et al., 2014). The distinct ∂13C depth pattern is a consequence of the use of different sources by fungi and bacteria as investigated by Kohl et al. (2015) for peat profiles. They conclude that an increasing ∂13C signal is caused by differences in biomass synthesis and carbon sources used by fungi and bacteria, which was also reported by Lichtfouse et al. (1995).
- 55 We suggest that changing microbial abundance must also be reflected in specific nitrogen stable isotope depth patterns. But, in contrast to the well-studied carbon isotope depth pattern, there are less data available for nitrogen. Caused by changing microbial abundance, which were also mentioned by Tfaily et al. (2014) or Hobbie and Ouimette (2009), we hypothesize that a peak of ∂15N values will be present at the same depth where the maximum diversity of microbial metabolism can be found.
- 60 To indicate the abundance of specific microbial communities in the peat horizons phospholipid fatty acids (PLFAs) are valid markers, because they are specific and persistent compounds of cell membranes of different species (Willers, Jansen van







Rensburg and Claassens, 2015; Reiffarth et al., 2016). Therefore PLFAs enable us to make qualitative and quantitative statements about the relative abundance of different microbial communities.

- Here, we use specific terms for the change points in the stable isotope depth pattern and the horizon descriptions. The points where the stable isotope signals undergo a sudden directional shift with depth are called "turning points" according to Alewell et al. (2011). Furthermore, the mesotelm is enlarged in drainage-affected cores and we will use the term "upper mesotelm" for the uppermost part of a drainage-affected horizon and the term "lower mesotelm" for the deeper part of a drainage-affected horizon. The bottom of the drainage-affected horizon and the onset of the underlying catotelm are marked by the ∂13C turning point. If rewetting processes are present above the mesotelm, the horizon is called "rewetting horizon".
- 70 Our aim is to evaluate ∂15N depth trends as indicators of specific peatland conditions and to study whether ∂15N depth trends of natural and drainage-affected sites indicate, in parallel to ∂13C, a shift in dominant microbial communities, reflected by specific PLFAs. We are also interested to see if there are distinct changes of ∂13C and ∂15N isotope signatures with the onset of rewetting processes. We match isotope depth trends (∂13C, ∂15N) with bulk density (BD) and carbon/nitrogen ratio (C/N). BD acts as an indicator for decomposition, because decomposition processes are leading to
- 75 higher compaction of the peat soil and therefore increasing BD values. The C/N ratio gives us also information about the degree of decomposition. With increasing decomposition the C/N ratio decreases. With an additional microscope analysis of the macro-residuals in the peat horizon, we will get information of the humification indices (HI) and the vegetation assemblages. To test our hypothesis of changing dominant microbial communities as drivers for isotope patterns, we do a PLFA analysis of four investigated sites two drainage-affected and two natural sites in Degerö Stromyr and Lakkasuo. We
- 80 will test the existence of two Gram positive bacterial (i-C15:0; a-C15:0) markers and one fungal (C18:2ω9c) marker (Sundh, Nilsson and Borga, 1997; Elvert et al., 2003). In the Swedish site Degerö Stormyr we add information of tree ring development as an indicator of peatland dynamics.

2 Material and methods

2.1 Site description

- 85 We studied five oligotrophic peatlands (Tab. 1, Tab. 2). Degerö Stormyr (200 m above see level (a.s.l.)) is situated in Northern Sweden, at the Kulbäcksliden Experimental Forest near Vindeln, between the rivers Umeälven and Vindelälven (Eurola, Hicks & Kaakinen, 1984). It is an acidic mire with minerotrophic conditions and consists of interconnected small mire patches divided by ridges of glacial till. The climate is characterized as cold with no dry seasons and cold summers (Dfc-zone after Köppen-Geiger classification; Peel et al., 2007).
- 90 In Degerö ditches were installed at the beginning of the 20th century, were closed in 2017 and a naturally reestablishment of sphagnum took place afterwards. The water table is at the surface in the natural part (DNM) (Nielsson et al., 2008) and in around 10-15 cm depths at the drainage-affected location (DDC).







Lakkasuo (150 m a.s.l.), Central Finland, is an eccentric peatland complex with two parts. In the southern part the conditions are ombrotrophic, whereas the northern part is minerotrophic (Minkkinen et al. 1999). Lakkasuo is also located in the cold

95 climate zone, with no dry seasons and cold summers (Dfc-zone after Köppen-Geiger classification; Peel et al. 2007). The 1961 installed ditches (70 cm depth, spacing of 40 m – 60 m) affect approximately 50 % of the peatland (Minkkinen et al., 1999). In the ombrotrophic natural site (LON) the water table was around 13 cm below ground surface. The ombrotrophic drained site (LOD) had a water table of 26 cm depth (average), whereas the water table is near the surface at the minerotrophic natural site (LMN) and in an average depth of 36 cm in the minerotrophic, drained site (LMD) (Minkkinen et al.)

100 al., 1999) (Tab. 1, Tab. 2).

In the Black Forest three mires were investigated: Breitlohmisse, Ursee and Rotmeer. They are located in the temperate climate zone with no dry seasons and warm summers (Cfb-zone after Köppen-Geiger; Peel et al., 2007). In the mires of the Black Forest ditches were installed in the middle of the 20th century. Breitlohmisse (810 m a.s.l., 50 km southeast of Baden-Baden) is minerotrophic and is located in the Northern part of the Black Forest. The mire is mostly lanced with ditches for

- 105 huntsman ships (Br2). The ditches are naturally refilled with Sphagnum. The water table is at an average depth of 15 cm in the natural center (Br1, Br2), and is found at lower depths near the degraded edges of the mire (Br3, Br4). Rotmeer (960 m a.s.l., 40 km southeast of Freiburg i.B.) and Ursee (850 m a.s.l., 45 km southeast of Freiburg i.B.) are both in the Southern Black Forest. Rotmeer consists of an ombrotrophic center (Ro1) (water table at the surface), surrounded by a minerotrophic part with signs of decomposition (Ro2, water table around 12 cm depth) and without mosses at the edges (Ro3, water table
- 110 below 12 cm depth). Urmeer is minerotrophic. A quaking bog forms the center with the water table at the surface (Ur2), whereas the edges had a lower water table (Ur1) (Tab. 1, Tab. 2).

2.2 Soil sampling and bulk analyses

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In May 2012 (Breitlohmisse), June 2012 (Rotmeer), July 2012 (Ursee) and September 2013 (Degerö and Lakkasuo) three volumetric peat cores were drilled per site with a Russian peat corer (Eijkelkamp, The Netherland) at a medium stage of

- 115 small-scale topography. In Degerö cores were sampled in the assumed natural center of the mire (DNM) and in one-meter distance to a drainage ditch (one meter depth) (DDC). In Lakkasuo we took cores at the natural sites (ombrotrophic natural (LON), minerotrophic natural (LMN)) and the drainage-affected locations (ombrotrophic drained (LOD), minerotrophic drained (LMD)). For Ursee two cores were taken, one in the natural center (Ur2) and one at the drainage-affected edge of the mire (Ur1). In Breitlohmisse and Rotmeer we took cores in a transect from natural (Br1, Ro1) to strong drainage-affected
- 120 (Br4, Ro3) sites. Each core has a composite length of one meter. Here, we focus on the uppermost 60 cm because this part included the drainage-affected horizon and no major changes in isotopic composition were observed at the natural sites below the acrotelm.

Directly after drilling HI were determined for each horizon with the von Post scale. The von Post scale has a range form 1 to 10. HI 1 indicates natural condition with undecomposed, completely visible vegetation residuals. HI 10 represents a strongly decomposed horizon without visible vegetation residuals. (Silc and Stanek, 1977)





The cores were encased in plastic shells and covered with plastic wrap, stored in coolers, and transported to the laboratory. The cores were sliced in 2 cm sections and every second layer was analysed, giving a 4 cm depth resolution. Samples were oven-dried at 40 °C for 72 h, and homogenized with a vibrating ball mill (MM400, Retsch, Germany). Stable C and N isotopic signatures were measured an elemental analyser combined with an isotope ratio mass spectrometer (EA-IRMS)

- 130 (Inegra2, Sercon, Crewe, UK). Carbon isotopic composition (13C/12C) was expressed relative to Vienna Pee-Dee Belemnite (VPDB) standard and reported in delta notation (‰), stable nitrogen isotopes were expressed relative to the atmospheric nitrogen standard and reported in delta notation (‰). C/N was determined with the mass relationship of the measured bulk content of C and N. Bulk density was measured with volumetric samples, which were weighted before and after drying. In Degerö tree rings of seven individual trees were analysed (Pinus sylvestris) to obtain information of growth conditions
- 135 and to enhance therefore our knowledge of drainage history.

2.3 Fatty acid analysis

Four cores (per site one drainage-affected and one natural core) were selected to do a fatty acid analysis: two sites in Lakkasuo, LOD 1 and LON 3 and two sites for Degerö Stromyr, DDC 3 and DNM 1. We took subsamples in all cores in the acrotelm (respectively at the end of the mesotelm in DDC) and in the catotelm. At the drainage-affected sites DDC 3 and LOD 1 we took also samples in the middle and at the end of the mesotelm.

We processed 0.2 - 1.1 g of sample for the phospholipid extraction with a mixture of CH2Cl2 : MeOH (9:1 v/v) in an Accelerated Solvent Extractor (Dionex ASE 350). 50 µl of an internal standard with nonadecanoic acid was added before processing each sample.

The total lipid extracts (TLE) were saponified by adding 2 ml of KOH dissolved in MeOH (12%) and putting it in the oven for 3 hours at 80° C

145 for 3 hours at 80°C.

Following the method of Elvert et al. (2003) TLE was afterwards pooled with 1 ml KCl (0.1 mol) and the neutral fraction was extracted by agitating three times with hexane. Neutral fraction in the supernatant was separated, dried under a stream of N2, and stored in the fridge for later analysis. We acidified the rest of the TLE with fuming hydrochloric acid to a pH of 1. The acid fraction was extracted by agitating again three times with hexane. The acid fraction in the supernatant was

- 150 separated and hexane was reduced to almost dryness under a stream of N2. Then the acid fraction was methylated by adding 1 ml Boron-Trifluoride (BF3) in MeOH (12-14%) and putting it in the oven for 1 hour at 60°C. Afterwards the PLFA fraction was pooled with KCl (0.1 mol) and transferred in 2 ml vials by agitating three times with hexane. The PLFAs were quantified with a Trace Ultra gas chromatograph (GC) equipped with a flame ionization detector (FID) (Thermo Scientific, Waltham, MA, USA). The carrier gas (helium) had a constant flow of 1.2 ml per minute and the GC-FID was set to splitless
- 155 mode. Detector temperature was 320°C and the samples (dissolved in hexane) were injected by 300°C. The starting temperature of the oven was 50°C. The temperature was increased by 10°C per minute to 140°C. The temperature was held for 1 minute before it was increased up to 300°C. This temperature was held for 63 minutes.







To identify the fungal and bacterial markers, we used the Bacterial Acid Methyl Esters standard (BAME, Supelco Mix), which includes as markers the PLFAs i-C15:0 and a-C-15:0 for Gram positive - bacteria (Zelles, 1997, O'Leary and

160 Wilkinson, 1988; Vestal and White, 1989) and C18:2ω9c for fungi (Andersen et al., 2010; Sundh, Nilsson and Borga, 1997; Zelles, 1997, O'Leary and Wilkinson, 1988; Vestal and White, 1989). Quantification of the PLFAs was done using the internal standard, C19:0 FA, after correcting for the methyl group, added during methylation reaction.

2.4 Data evaluation and statistical analysis

- As we were interested in comparing the depth trends of all single profiles with each other, we first normalized the depths of the cores. This was done using the depth of the $\partial 15N$ turning point (see chapter 3.1) in each drainage-affected profile as the anchor point serving as normalized depth (normD). The normalized depth of this anchor point was set to 20 cm depth (normD = 20 cm, fig. 1) in each single core. In the corresponding natural cores, we have transferred the values from the same depth related to the drained core into the same norm depth. For example the values of the natural site (DNM) in depth of 13 cm (depth of the turning point of $\partial 15N$ in the corresponding DDC core) were set to 20 cm normD.
- 170 In a second step, because we were mainly interested in trends and not the absolute values, we normalized the isotopic values themselves, because the range of ∂15N varied considerably between the sites, whereas the trends show consistent patterns (fig. 1). Therefore, to be able to do a meaningful comparison we set therefore the value of ∂15N at the turning point to zero in each profile:

175 normalized
$$\partial^{15}N[\%] = \partial^{15}N[\%] - \partial^{15}N[\%]$$
 at turning point

Using the same procedure, all other parameters (∂ 13C, C/N, BD) were normalized using the same anchor point (e.g., ∂ 15N turning point):

180 normalized value ($\partial^{l3}C$ [‰], BD, C/N) = value ($\partial^{l3}C$ [‰], BD, C/N) – value ($\partial^{l3}C$ [‰], C/N, BD) at $\partial^{l3}N$ turning point

Using the above procedures means to decide on the depth of the $\partial 15N$ turning points, which we backed up statistically with a t-test (p ≤ 0.05) and an integrated change point analysis with the software package "changepoint" in R (version 1.0.153). These analyses were done for each of the drained sites separately and also in addition with an average of all locations. For

185 the t-test, we analyzed for each depth if ∂15N values in the drainage-affected horizon are of the same population as the values of the natural sites (H0: drained and natural values are of the same population). For the changepoint analysis, the variance of ∂15N was evaluated with a linear gradient over the whole drainage-affected peat profile against the variance of three/ four separated linear gradients (rewetted part (if present), upper mesotelm, lower mesotelm, catotelm). Here, we define







the starting point of the drainage-affected horizon with the onset of a shift in the $\partial 15N$ values upward and the end of this 190 horizon with the stabilization of the $\partial 15N$ values towards the surface.

We also determined the slopes of each single core to get information on the strength of differences of the isotopic values with depth. First, the whole peat profile of each drainage-affected core was analysed as one trend (called "overall profile"). Second, profiles were separated into different horizons: (i) rewetted horizon (if present), (ii) upper mesotelm, (iii) lower mesotelm and (iv) catotelm. If values were clearly changing with depth slopes were closer to zero. In horizons with 195 stabilized values slopes were distinct higher or lower zero.

In the following we present only the normalized data. Raw data without normalization are available in the supplementary information.

2.5 Tree Ring width and microscope analysis of peat

The investigation of the tree ring width of seven surrounding trees (Pinus sylvestris) in Degerö Stormyr was done with a 200 hand-operated wood driller (Djos/ Sweden, 5 mm diameter). Samples were fixed on wooden carriers. The tracheids (elongated cells of the xylem of vascular plants) were cut with a sharp carbon blade and analysed with an impinging light binocular (60x – 160x amplification).

Peat samples of four study sites were analysed using an impinging light binocular (60x - 160x amplification) to get an overview of the vegetation assemblages and to differentiate horizons. For detailed information (distinction of Sphagnum

species) the samples were elutriated with water, pigmented with methyl-blue and analysed under a transmitted light microscope (100x - 640x amplification).

3 Results and discussion

3.1 Stable nitrogen isotope depth trends as indicators for peatland conditions

While mineral soils have been shown to have continuous increasing values of ∂15N (Nadelhoffer et al, 1996, Högberg et al,
1997), we found increasing ∂15N and ∂13C values with depth down to particular, isotopic specific turning points in drainage-affected peatland soils (fig. 1). We defined three types of peatland conditions from the observed depth trends of ∂15N: (a) natural, (b) drainage-affected up to the surface, (c) profiles with a rewetted horizon above the drainage-affected horizon. (fig. 1)

Eight out of nine studied drained peatlands and the average trend confirm the existence of a $\partial 15N$ turning point, with a p-215 value < 0.05 for the difference in $\partial 15N$, in the center of the mesotelm in contrast to the $\partial 15N$ values in undisturbed horizons (Tab.3), with a non-significant difference for one drained site: Breitlohmisse (Br3). The latter was most likely related to generally higher $\partial 15N$ values of the natural site in Breitlohmisse (Br1) compared to a smaller increase of $\partial 15N$ in this





drainage-affected site (Br3). The depth of $\partial 15N$ turning point (center of the mesotelm) differs from $\partial 13C$ turning point (end of the mesotelm) for all investigated sites (fig. 2).

- 220 Changed slope values of the separated horizons indicate significant trend changes (Tab. 5). In anaerobic conditions (natural, catotelm) with stabilized isotopic values with depth, slopes were distinct different to 0 [cm/‰]. ∂15N values seems to change within the mesotelm rapidly and slope values were closer to zero. Most interesting was a switch to negative trend values at the ∂15N turning point in all investigated drainage-affected sites, which marks the beginning of the lower mesotelm. (Tab. 5)
- In natural conditions (type (a)), all investigated parameters had a low variability and indicated natural, wet mire conditions (fig. 1). There were two exceptions: Breitlohmisse 1 (Br1) (40 60 cm normD) and Rotmeer 1 (Ro1) (30 50 cm nromD), with trend instabilities of ∂15N. This might indicate some minor drainage or disturbance in the wetland sites we classified as "natural" (fig. 1).

In contrast, the values of the drained sites show significant trends. We found two different trend types in the drainageaffected sites: Type (b) and (c) (fig. 1). For type (b) we distinguished six sites: Lakkasuo ombrotrophic drained (LOD),

- 230 affected sites: Type (b) and (c) (fig. 1). For type (b) we distinguished six sites: Lakkasuo ombrotrophic drained (LOD), Breitlohmisse 2 (Br2), Breitlohmisse 3 (Br3), Breitlohmisse 4 (Br4), Rotmeer 2 (Ro2) and Rotmeer 3 (Ro3) with clear signs of decomposition up to the surface. Type (c) was visible in three sites: drainage-affected site Degerö Stromyr (DDC), minerotrophic drainage-affected site Lakkasuo (LMD) and Ursee 1 (Ur1). At type (c) sites the isotopic values, C/N and BD were stabilized again above the mesotelm. Therefore, they are assumed to be in a "new" natural status (fig. 1, fig. 2).
- Below 8 cm (normD, average profile) all drainage-affected profiles showed the typical signs of the upper mesotelm with increasing values of ∂15N, ∂13C and BD, down to the ∂15N turning point, and decreasing C/N. Below the ∂15N turning point, in the lower mesotelm, ∂15N values were decreasing. In this horizon ∂13C values, C/N and BD were increasing. The end of the lower mesotelm was mostly linked to a clear shift in ∂13C trend to either stable values or a slow decreasing trend; hence, we called this point ∂13C turning point (28 cm normD, average profile) (e.g. Krüger et al. 2014). Constant C/N, BD
- 240 and ∂15N values below the ∂13C turning point served also as indicators for reduced compaction and decomposition. Most likely the ∂13C turning point marked the onset of permanent waterlogged anaerobic conditions (e.g. Krüger et al. 2016a). The similarity in trends in these deeper parts of the drainage-affected sites to those of the catotelm in the natural sites supported the assumption of an intact catotelm below the ∂13C turning point (fig. 1, fig. 2.). (For the single ∂13C, C/N and BD values of all peat cores see supplementary information).

3.2 Depth profile of vegetation assemblage and water table in connection with isotope depth pattern

All sites, which we attributed as "natural" (type (a)), had a water table near the surface (<10 cm), and macro-residuals were highly visible throughout the profile, HIs were low and the main living vegetation was Sphagnum spp. (tab. 6, tab.7). All drainage-affected sites had higher HIs even if no direct modifications in the vegetation assemblage could be documented. For type (b), there was little or no Sphagnum visible at the surface and the water table was found at lower depths. Macro-

250 residuals were more affected by decomposition and HIs were high up to the surface. These results were in line with our







interpretation of isotope signatures of a drainage-affected horizon up to the surface. Especially the ombrotrophic-drained site (LOD) was influenced by drainage. Here Sphagnum species seem to have disappeared and were replaced by mosses of drier environments or mosses were completely absent (tab. 7).

For type (c), vegetation assemblages were mainly composed of Sphagnum spp. and the water table was near the surface. HIs 255 were low in the rewetted horizon and macro-residuals were preserved well (tab. 6, tab. 7). With the onset of the upper mesotelm, HIs and decomposition of macro-residuals was high. In the lower mesotelm, the HIs were decreasing and more macro-residuals were visible. In the catotelm, the quality of macro residuals was higher than in the mesotelm and the HIs were even lower.

3.3 Tree ring width are verifying isotope signals of changing peatland conditions

- 260 Tree ring width is a marker for the wellbeing and/or growth rate of trees. Young trees have a small scope coupled with high growth rates, which leads to thicker tree rings. Tree rings get smaller with increasing age of the tree. If there are no environmental stressors like heat, increasing wetness or drought, tree rings are bigger and the cell lumen is higher compared to trees at sites with environmental stress. With increasing environmental stress tree ring width decreases (Stoffel et al., 2010). Before 1992, tree rings at the drainage-affected site (DDC) site showed only a slightly decreasing trend, which could
- 265 be due to aging (average of 1.3 mm width in the 1930s to an average width of 0.9 mm in the late 1980s). The draining ditches in Degerö Stormyr were established in the beginning of the 20th century, which supports these results, with dryer and therefore better growth conditions for trees. From 1992 onwards tree ring widths decreased, reaching 0.2 mm in 1998 and thereafter. These results were concurrent with the isotope analysis, because both suggest a restoration of natural wet peatland conditions. Rewetted conditions are no longer suitable for trees and lead to smaller tree ring width according to adverse
- 270 environmental conditions for tree growth. These findings underpin our suggestion of rewetted conditions at this site in Degerö.

3.4 Linkage of microbial abundance and isotopic signature

At the natural sites, all samples near the surface had relatively low, fungal-derived PLFA concentrations. In the catotelm the values were very low, dominated by bacterial-derived PLFAs (fig. 3).

- 275 In the drainage-affected sites microbial-derived PLFA abundance was increased over the whole mesotelm. This pattern could be caused by the improved conditions for metabolism processes by drainage: enhanced oxygen abundance and relatively high nutrient availability of the prior conserved plant material. In the upper mesotelm dominantly fungal-derived, but also bacterial-derived, PLFAs were visible. At the *∂*15N turning point lowered values of fungal markers and increased bacterialderived PLFAs could be found. In the lower mesotelm the abundance of microbial-derived PLFAs was generally decreased,
- 280 however more bacterial than fungal marker were detected. In the catotelm microbial-derived PLFA abundance was low, similar to the natural sites. (fig. 3)







The observed switch, from predominantly fungal abundance in the acrotelm to bacterial abundance in the lower mesotelm, is in agreement with Kohl et al., 2015 and Schmidt and Bölter, 2002. Also Andersen et al (2013) stated out, that fungi biomass is decreasing in peatland soils. Our investigations indicate, that this switch is linked to the depth pattern of isotopic N (fig.

- 4), which is in accordance with the findings of Wallander et al. (2009); Winsborough and Basiliko (2010) and Myers et al. (2012). With increasing depth microbial utilization of nitrogen and carbon via alternative pathways and from recycled sources (with enriched ∂15N values) is necessary (Dijkstra et al. 2006, Dröllinger et al. 2019). The reason for this is that increasing depth leads to increasing oxygen limitation and thus declining decomposability of the remaining substrate.
- The nitrogen isotopic signal of peat material is generally depleted compared to atmospheric nitrogen (which is, per definition, 0 ‰, if air is used as the nitrogen isotopic standard). As such, the average signal of the relatively undecomposed peat (e.g., the "natural" sites, the catotelm or the peat grown under rewetted conditions) is -10 to -4 ‰. The latter is parallel to the isotopic depletion of ∂13C in plant material, due to the general preference of plants for the lighter isotopes 12C and 14N. However, with the onset of drainage, decomposition of the organic plant material also below the acrotelm take place, resulting in a preferential mineralization of 14N and an enrichment of 15N in the remaining peat material. In acid bogs under
- 295 aerobic conditions, fungi will dominate the general metabolism (Thormann et al., 2003). This is pictured by the highest amount of fungal-derived PLFAs in the acrotelm and the upper mesotelm. Fungi are preferred decomposers of primary plant material (Wallander et al., 2009; Thormann et al., 2003) hence the depleted plant isotopic signal is relatively preserved in the upper most aerobic horizons. Furthermore, fungi have a relatively low nitrogen demand compared to bacteria (Myers et al. 2012). This explains, in line with Thormann (2005), our pattern of lower ∂15N isotope values in the acrotelm compared to
- 300 the mesotelm, where the incoming plant/ moss signal is more and more enriched with depth by decomposition processes and nitrogen turnover. With increasing depth and increasing oxygen limitation fungal metabolism decreases (Thormann, 2011). In parallel, bacterial metabolism increases as Lin et al. (2014); Hu et al. (2011) and Bauersachs et al. (2009) reported. They found evidence for bacterial-dominated decomposition in hypoxic conditions. Bacterial metabolism is generally faster than fungal metabolism (Brunner et al., 2013). Faster turnover is connected with less isotopic fractionation. In addition, bacterial
- 305 metabolism needs higher amounts of nitrogen. In summary, with increasing depth and increasing bacterial metabolism most of the available nitrogen will be immobilized, which results in no or low fractionation of the bulk material (e.g., no preferential loss of the lighter 14N). Hence, the ∂ 15N turning point could be caused by the N-limitation of peatland ecosystems with low oxygen availability. At this depth (20 cm normD) bacteria and fungi compete most over decomposable substrates (not necessarily nitrogen), resulting in the highest turnover rates with enrichment of ∂ 15N in the remaining peat,
- 310 as we know from mineral soils with aerobic decomposition (Alewell, et al. 2011, Nadelhofer, et al 1996). Tfaily et al. (2014) also reported the highest ∂15N values within the mesotelm. This pattern is reflected in the equilibrium of fungal- and bacterial-derived PLFAs at the ∂15N turning point (fig. 3).







In the lower mesotelm, oxygen limitation increases, leading to decreasing microbial metabolism and decreasing concentrations of microbial-derived PLFAs. The decreasing microbial metabolism leads to simultaneously decreasing $\partial 15N$

315 values because an increasing amount of undecomposed vegetation (with low ∂15N values) will be conserved. (fig. 4) Finally, with the establishment of permanently waterlogged anaerobic conditions (indicated by the ∂13C turning point), PLFA concentration decreases sharply. In the catotelm decomposition processes are mostly inhibited, which leads to stable ∂15N and ∂13C values, close to the original vegetation signals (Alewell et al., 2011; Krüger et al, 2015). (fig. 1, fig3)

4. Conclusion

- 320 Our results show significant differences in the nitrogen isotopic depth trends of natural, drainage-affected and rewetted peat profiles. We validated our isotopic hypothesis with microscope analysis of the vegetation remains in the cores as well as the investigation of tree rings as indicators for changing hydrological conditions in the past. An analysis of bacterial- versus fungal- derived PLFAs unravelled the changing microbial abundance with depth as characterized by high fungal abundance in the aerobic acrotelm with low nitrogen demand and turnover; transition to a mixture of decreasing fungal and increasing
- 325 bacterial abundance in the upper mesotelm, competing on organic substrates, resulting in an enrichment of ∂15N; decreasing microbial decomposition, dominated by bacterial abundance, in the lower mesotelm and impeded microbial metabolism and stabilized ∂15N values in the anaerobic catotelm.

Carbon isotope signatures are also changing with drainage, but there is neither a clear indication of a switch in microbial abundance within the drainage-affected horizon, nor a clear change in the trend with rewetting of the peatland, as it is visible

330 for the nitrogen signatures due nitrogen limitation and recycling processes. Summing up, ∂15N depth profiles in peat might give more insights into the degree of microbial transformation, because they reflect more precisely different microbial abundance than carbon isotope signatures do. Therefore, we conclude that ∂15N depth profiles could act as a reliable and efficient tool to get fast and easy information about peatland status, restoration success and drainage history.

335 Author contribution

Christine Alewell and Jens Leifeld are the supervisors of the project. Miriam Groß-Schmölders, Jan Paul Krüger, Axel Birkholz and Kristy Woodard were doing the measurements. Pascal von Sengbusch was doing the microscopy and vegetation analysis. Miriam Groß-Schmölders prepared the manuscript with contribution of all co-authors.







340 Competing interests

The authors declare that they have no conflict of interest.

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References

345 Alewell, C., Giesler, R., Klaminder, J., Leifeld, J., and Rollog, M.: Stable carbon isotope as indicator for environmental change in palsa peats, Biogeoscience, 8, 1769-1778, https://doi.org/10.5194/bg-8-1769-2011, 2011.

Alexandersson, H., Karlström, C. and Larsson-Mccan, S.: Temperature and precipitation in Sweden 1961-1990. Reference normals, Swedish Meteorological and Hydrological Institute (SMHI), Meterologi, 81, 1991.

 Andersen, R., Grasset, L., Thormann, M., Rochefort, L. and Francez, A.-J.: Changes in microbial community structure and
 function following Sphagnum peatland restoration, Soil Biology & Biochemistry, 42, 291-301, https://doi.org/10.1016/j.soilbio.2009.11.006, 2010.

Armbruster, M., Abiy, M., and Feger, K-H.: The biogeochemistry of two forested catchments in the Black Forest and the eastern Ore Mountains, Biogeochemistry, 65, 341-368, https://doi.org/10.1023/A:1026250209699, 2003.

Artz, R. R. E.: Microbial Community Structure and Carbon Substrate use in Northern Peatlands, Carbon Cycling in Northern
 Peatlands, 111–129, https://doi.org/10.1029/2008GM000806, 2013.

Asada, T., Warner, B. G. and Aravena, R.: Nitrogen isotope signature variability in plant species from open peatland, Aquatic Botany, 82(4), 297–307, https://doi.org/10.1016/j.aquabot.2005.05.005, 2005.

Baumann, K., Dignac, M.-F., Rumpel, C., Bardoux, G., Sarr, A., Steffens, M. and Maron, P.-A.: Soil microbial diversity affects soil organic matter decomposition in a silty grassland soil, Biogeochemistry, 114(1-3), 201–212, https://doi.org/10.1007/s10533-012-9800-6, 2013.

Bauersachs, T., Schouten, S., Compaoré, Wollenzoen, U., Stal, L. and Sinninghe Damsté, J.: Nitrogen isotopic fractionation associated with growth on dinitrogen gas and nitrate by cyanobacteria, Limnological Oceanography, 54, 1403-1411, https://doi.org/10.4319/lo.2009.54.4.1403, 2009.





Biester, H., Knorr, K.-H., Schellekens, J., Basler, A. and Hermanns, Y.-M.: Comparison of different methods to determine
the degree of peat decomposition in peat bogs, Biogeosciences, 11, 2691 - 2707, https://doi.org/10.5194/bg-11-2691-2014, 2014.

Boyle, S., Yarwood, R., Bottomley, P. and Myrold, D.: Bacterial and fungal contributions to soil nitrogen cycling under Douglas fir and red alder at two sites in Oregon, Soil Biology and Biochemistry, 40, 443-451, https://doi.org/10.1016/j.soilbio.2007.09.007, 2008.

370 Brunner, B., Contreras, S., Lehman, M., Matantseva, Rollog, M., Kalvelage, T., Klockgether, G., Gaute, L., Jetten, M., Kartal, B. and Kuypers, M.: Nitrogen isotope effects induced by anammox bacteria, PNAS, 110, 18994-18999, https://doi.org/10.1073/pnas.1310488110, 2013.

De Boer, W., Folman, L., Summerbell, R. and Boddy, L.: Living in a fungal world: impact of fungi on soil bacterial niche development, FEMS Microbiology Reviews, 29, 795-811, https://doi.org/10.1016/j.femsre.2004.11.005, 2005.

375 Crump, J [Eds]: Smoke on Water - Countering Global Threats From Peatland Loss and Degradation - A UNEP Rapid Response Assessment. United Nations Environment Program and GRID-Arendal, Nairobi and Arendal, 2017.

Clymo, R. and Bryant, C.: Diffusion and mass of dissolved carbon dioxide, methane and dissolved organic carbon in a 7-m deep raised peat bog, Geochimica et Cosmochimica Acta, 72, 20488-2066, https://doi.org/10.1016/j.gca.2008.01.032, 2008.

Deutscher Wetter Dienst (DWD): Wetter und Klima vor Ort: https://www.dwd.de/DE/wetter/wetterundklima_vorort/baden-380 wuerttemberg/feldberg/_node.html, last access: 18 September 2018.

Dijkstra, P., Ishizu, A., Doucett, R., Hart, S., Schwartz, E., Manyailo, O. and Hungate, B.: 13C and 15N natural abundance of the soil microbial biomass, Soil Biology & Biochemistry, 38, 3257-3266, https://doi.org/10.1016/j.soilbio.2006.04.005, 2006.

Drollinger, S., Kuzyakov, Y. and Glatzel, S.: Effects of peat decomposition on $\partial 13C$ and $\partial 15N$ depth profiles of Alpine 385 bogs, Catena, 178, 1–10, https://doi.org/10.1016/j.catena.2019.02.027, 2019.

Elvert, M., Boetius, A., Knittel, K. and Jørgensen, B.: Characterization of Specific Membrane Fatty Acids as Chemotaxonomic Markers for Sulfate-Reducing Bacterial Involved in Anaerobic Oxidation of Methane, Geomicrobiology Journal, 20, 403-419, https://doi.org/10.1080/01490450303894, 2003.

•





Eurola S., Hicks S. T., Kaakinen E.: Key to Finnish mire types, in: European Mires, edited by: Moore, P. D., Academic 390 Press, London, Great Britain, 1–117, 1984.

Federle, T. W.: Microbial distribution in soil-new techniques, Perspective in Microbial Ecology, 493-498, 1986.

Högberg, P., Högbom, L., Schinkel, H., Högberg, M., Johannisson, C. and Wallmark, H.: 15N abundance of surface soils, roots and mycorrhizas in profiles of European forest soils, Oecologia, 108, 207-214, https://doi.org/10.1007/BF00334643, 1996.

395 Högberg, P.: 15N natural abundance in soil-plant systems, The New Phytologist, 137 (2), 179-203, https://doi.org/10.1046/j.1469-8137.1997.00808.x, 1997.

Hobbie, E., Weber, N. and Trappe, J.: Mycorrhizal vs saprotrophic status of fungi: the isotopic evidence, The New Phytologist, 150, 601-610, https://doi.org/10.1046/j.1469-8137.2001.00134.x, 2001.

Hobbie, E.: Evidence that saprotrophic fungi mobilize carbon and mycorrhizal fungi mobilize nitrogen during litter 400 decomposition, The New Phytologist, 173, 447-449, https://dx.doi.org/10.1111/j.1469-8137.2007.01984.x, 2007.

Hobbie, E. A., and A. P. Ouimette: Controls of nitrogen isotope patterns in soil profiles, Biogeochemistry, 95 (2-3), 355-371, https://doi.org/10.1007/s10533-009-9328-6, 2009.

Hobbie, E. and Högberg, P.: Nitrogen isotopes link mycorrhizal fungi and plants to nitrogen dynamics, The New Phytologist, 196, 367-382, https://doi.org/10.1111/j.1469-8137.2012.04300.x, 2012.

405 Hobbie, E. A., Chen, J., Hanson, P. J., Iversen, C. M., Mcfarlane, K. J., Thorp, N. R., and Hofmockel, K. S.: Long-term carbon and nitrogen dynamics at SPRUCE revealed through stable isotopes in peat profiles, Biogeosciences, 14(9), 2481– 2494, https://dx.doi.org/10.5194/bg-14-2481-2017, 2017.

Hu, B.-I., Rush, D., Van der Biezen, E., Zheng, P., Van Mullekom, M., Schouten, S., J., Sinninghe Damasté, A. Smolders, Jetten, M. and Kartal, B.: New Anaerobe, Ammonium-Oxidizing Community Enriched from Peat Soil, Applied and
Environmental Microbiology, 77, 966-971, http://dx.doi.org/10.1128/AEM.02402-10, 2011.

Joosten, H and Couwenberg, J. (2001): Das Beispiel Europa, in: Landschaftsökologische Moorkunde, edited by: Succow, M. and Joosten, H, Stuttgart, Germany, 406-408, 2001.





Joosten H.: Peatlands and carbon, Assessment on peatlands, biodiversity and climate change, in: Global Environment Centre/Wetlands International, Kuala Lumpur/ Wageningen, 99–117, 2008.

415 Kohl, L., Laganierè, J., Edwards, K., Billings, S., Morril, P., Van Biesen, G. and Ziegler, S.: Distinct fungal and bacterial ∂13C signatures as potential drivers of increasing ∂13C od soil organic matter with depth, Biogeochemistry, 124,13-26, https://doi.org/10.1007/s10533-015-0107-2, 2015.

Krüger, J. P., Leifeld, J. and Alewell, C.: Degradation changes stable carbon isotope depth profiles in palsa peatlands, Biogeoscience, 11, 3369-3380, https://doi.org/10.5194/bg-11-3369-2014, 2014.

420 Krüger, J. P., Leifeld, J., Glatzel, S., Szidat, S. and Alewell, C.: Biogeochemical indicators of peatland degradation - a case study of a temperate bog in northern Germany, Biogeoscience, 12, 2861-2871, https://doi.org/10.5194/bg-12-2861-2015, 2015.

Krüger, J. P., Alewell, C., Minkkinen, K., Szidat, S. and Leifeld, J.: Calculating carbon changes in peat soils drained for forestry with four different profile-based methods, Forest Ecology and Management, 381, 29-36, https://doi.org/10.1016/j.foreco.2016.09.006, 2016a.

Krüger, J. P.: Peatland degradation indicated by stable isotope depth profiles and soil carbon loss, Ph.D. thesis, University of Basel, Switzerland, 143pp., 2016b.

Krüger, J. P., Conen, F., Leifeld, J. and Alewell, C.: Short communication: Palsa Uplift Identified by Stable Isotope Depth Profiles and Relation of ∂15N to C/N Ratio, Permafrost and Periglacial Processes, 28, 485-492, 430 https://doi.org/10.1002/ppp.1936, 2017.

Laine, J., Komulainen, V., Laiho, R., Minkkinen, K., Rasinmäki, A., Sallantus, T., Sarkkola, S., Silvan, N., Tolonen, K. and Tuittila, E, Vasander, H., Pälvänen, J. (Eds.): Lakkasuo: A Guide to a Mire Ecosystem, Department of Forest Ecology, University of Helsinki, Helsinki, 2004.

Leifeld, J. and Menichetti, L.: The underappreciated potential of peatlands in global climate change mitigation strategies, 435 nature communications, 9, 1071, https://doi.org/10.1038/s41467-018-03406-6, 2018.

Lichtfouse, E., Berthier, G., Houot, S.: Stable carbon isotope evidence for the microbial origin of C14-C18 n-alkanoic acids in soils, Organic Geochemistry, 23(9), 849-852, https://doi.org/10.1016/0146-6380(95)80006-D, 1995.





 Lin X., Tfaily, M., Green, S., Steinweg, J., Chanton, P., Imvittaya, A., Chanton, J., Cooper, W., Schadt, C. and Kostka, J.: Microbial Metabolic Potential for Carbon Degradation and Nutrient (Nitrogen and Phosphorus) Acquisition in an
 Ombrotrophic Peatland, Applied and Environmental Microbiology, 80, 3531-3540, https://doi.org/10.1128/AEM.00206-14, 2014.

Myers, B., Webster, K., Mclaughlin, J. and Basiliko, N.: Microbial activity across a boreal peatland nutrient gradient: the role of fungi and bacteria, Wetlands Ecological Management, 20, 77-88, https://doi.org/10.1007/s11273-011-9242-2, 2012.

Minkkinen, K., Vasander, H., Jauhiainen, S., Karsisto, M., and Laine, J.: Post-drain- age changes in vegetation composition
 and carbon balance in Lakkasuo mire, Central Finland, Plant and Soil, 207, https://doi.org/10.1023/A:1004466330076, 107–120, 1999.

Nadelhoffer, K., Shaver, G., Frey, B., Giblin, A., Johnson and L., MacKane, R.: 15N natural abundance and N use by tundra plants, Oecologica, 107, 386-394, https://doi.org/10.1007/BF00328456, 1996.

Nilsson, M., Sagerfors, J., Buffam, I., Laudon, H., Eriksson, T., Grelle, A., and Lindroth, A.: Contemporary carbon accumulation in a boreal oligotrophic minerogenic mire - A significant sink after accounting for all C-fluxes, Global Change Biology, 14(10), https://doi.org/10.1111/j.1365-2486.2008.01654.x, 2008.

Novák, M., Buzek, F. and Adamová, M.: Vertical trends in *∂*13C, *∂*15N and *∂*34S ratios in bulk Sphagnum peat, Soil Biology and Biochemistry, 31, 1343-1346, https://doi.org/10.1016/S0038-0717(99)00040-1, 1999.

O'Leary, W.M. and Wilkinson, S.: Gram-positive bacteria, in: Microbial Lipids, edited by: Ratledge, C. and Wilkinson, S.G.
 Academic Press, London, Great Britain, 117-185, 1988.

Peltoniemi, K., Fritze, H. and Laiho, R.: Response of fungal and actinobacterial communities to water-level drawdown in boreal peatland sites, Soil Biology & Biochemistry, 41, 1902-1914, https://doi.org/10.1016/j.soilbio.2009.06.018, 2009.

Peel, M. C., Finlayson, B. L., & McMahon, T. A.: Updated world map of the Köppen-Geiger climate classification, Hydrology and Earth System Sciences, 11(5), https://doi.org/10.5194/hess-11-1633-2007, 2007.

460 Reiffarth, D., Petticrew, E., Owens, P. and Lobb, D.: Sources of variability in fatty acids (FA) biomarkers in the application and compound-specific stable isotopes (CSSIs) to soil and sediment fingerprinting and tracing: A review, Sci. Total Environ., 565, 8-27, https://doi.org/10.1016/j.scitotenv.2016.04.137, 2016.





Sanderman, J., Hengl, T. and Fiske, G.: Soil carbon debt 12,000 years of human land use, PNAS, 114 (36); 9575 – 9580, https://doi.org/10.1073/pnas.1706103114, 2017.

465 Schmidt, N. and Bölter, M.: Fungal and bacterial biomass in tundra soils along an arctic transect from Taimyr Peninsula, central Siberia, Polar Biology, 25, 871-877, https://doi.org/10.1007/s00300-002-0422-7, 2002.

Silc, T. and Stanek, W.: Bulk density estimation of several peats in northern Ontario using the von Post humification scale, Canadian Journal of Soil Science, 57, 75.1977.

Stoffel, M., Bollschweiler, M, Butler, D. R. and Luckman, B. H. (Eds): Tree Rings and Natural Hazards - A State-of-the-Art, 470 Advances in Global Change Research, Bern, Switzerland, 2010.

Strickland, M. and Rousk, J.: Considering fungal: bacterial dominace in soils - Methods, controls, and ecosystem implications, Soil Biology & Biochemistry, 42, 1385-1395, https://doi.org/10.1016/j.soilbio.2010.05.007, 2010.

Sundh, I., Nilsson, M. and Borgå, P.: Variation in Microbial Community Structure in Two Boreal Peatlands as Determined by Analysis of Phospholipid Fatty Acid Profiles, Applied and environmental Microbiology, 64 (4), 1476-1482, 1997.

475 Tfaily, M., Cooper, W., Kostka, J., Chanton, P., Schadt, C., Hanson, P., Iversen, C. and Chanton, J.: Organic matter transformation in the peat column at Marcell Experimental Forest: Humification and vertical stratification, Journal of Geophysical Research: Biogeoscience, 119, 661-675, https://doi.org/10.1002/2013JG002492, 2014.

Thormann, M., Currah, R. and Bayley, S.: The mycorrhizal status of the dominant vegetation along a peatland gradient in southern boreal Alberta, Canada, Wetlands, 19, 438-450, https://doi.org/10.1007/BF03161775, 1999.

480 Thormann, M., Currah, R. and Bayley, S.: Succession of mircofungal assemblages in decomposing peatland plants, Plant and Soil, 250, 323-333, https://doi.org/10.1023/A:1022845604385, 2003.

Thormann, M., Currah, R. and Bayley, S.: Patterns of distribution of microfungi in decomposing bog and fen plants, Canadian Jornaul of Botanic, 82, 710-720, https://doi.org/10.1139/b04-025, 2004.

Thormann, M.: Diversity and finction of fungi in peatlands: A carbon cycling perspective, Canadian Journal of Soil Science, 281-293, https://doi.org/10.4141/S05-082, 2005.

Thormann, M., Rice, A. and Beilman, D.: Yeast in Peatlands: A Review of Richness and Roles in Peat Decomposition, Wetlands, 27, 761-773, https://doi.org/10.1672/0277-5212(2007)27[761:YIPARO]2.0.CO;2, 2007.





Thormann, M.: In vitro decomposition of Sphagnum-derived acrotelm and mesotelm peat by indigenous and alien basidiomycetous, Mires and Peat, 8, 1-12, 2011.

490 Vestal, J.R. and White, D.C.: Lipid analysis in microbial ecology, Biogeoscience, 39, 535-541, https://doi.org/10.2307/1310976, 1989.

Wallander, H., Mörth, C. and Giesler, R.: Increasing abundance of soil fungi is driver for 15N enrichment in soil profiles along a chronosequence undergoing isostatic rebound in northern Sweden, Oecologica, 160, 87-96, https://doi.org/10.1007/s00442-008-1270-0, 2009.

495 Willers, C., Jansen van Rensburg, P. J. and Claassens, S.: Phospholipid fatty acid profiling of microbial communities – a review of interpretations and recent applications, Applied Microbiology, 119, 1207-1218, https://doi.org/10.1111/jam.12902, 2015.

Winsborough, C. and Basiliko, N.: Fungal and Bacterial Activity in Nothern Peatlands, Geochemistry Journal, 27, 315-320, https://doi.org/10.1080/01490450903424432, 2010.

500 Y u, Z., Beilman, D.W., Frolking, S., MacDonald, G. M., Roulet, N. T., Camil, P., and Charman, D. J.: Peatlands an Their Role in the Global Carbon Cycle, EOS, 92, 97–106, https://doi.org/10.1029/2011EO120001, 2011.

Zedler, J. B. and Kercher, S.: WETLAND RESOURCES: Status, Trends, Ecosystem Services, and Restorability, Annual Review of Environment and Resources, 30(1), 39–74, https://doi.org/10.1146/annurev.energy.30.050504.144248, 2005.

Zelles, L.: Phospholipid fatty acid profiles in selected members of soil microbial communities, Chemosphere, 35, 275-294, 505 https://doi.org/10.1016/S0045-6535(97)00155-0, 1997.









Figure 1: ∂¹⁵N depth profiles in all natural and drained (or rewetted) sites; with normalized depth and normalized ∂¹⁵N values (see chapter 2.4); trend types: (a) natural (green), (b) drainage-affected up to the surface (orange) and (c) rewetted above drainage 510 (pink) (For single, non-normalized values see supplementary information).







Figure 2: Mean depth trends (∂^{15} N, ∂^{13} C, C/N and BD) of natural and drained sites of all nine investigated peatlands with normalized depth and normalization based on ∂^{15} N signatures (see chapter 2.4; For single ∂^{13} C, C/N and BD values of all peat cores see supplementary information).







Figure 3: PLFA amount of bacterial and fungal marker in LOD, LON, DDC and DNM in different horizons.







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Figure 4: Hypothesis of a microbial switch (fungi to bacteria) with depth, reflected by specific PLFAs, and its influence of the $\partial^{15}N$ depth trend; example photo and $\partial^{15}N$ values of the ombrotrophic, drained site in Lakkasuo (LOD) (note all isotope values are normalized to zero at turning point).







Table 1: Labeling of all drilling sites

Location	Labeling
Degerö	
natural mire	DNM
drained	DDC
Lakkasuo	
minerotrophic natural	LMN
minerotrophic drained	LMD
ombrotrophic natural	LON
ombrotrophic drained	LOD
Breitlohmisse	
natural mire	Br1
natural dry	Br2
drained	Br3
near the mire edge	Br4
Rotmeer	
natural mire	Ro1
drained, with Sphagnum	Ro2
drained, without Sphagnum	Ro3
Ursee	
natural mire	Ur2
drained	Ur1







Country	Country Mire lat/long MAT P		Р	Main vegetation	ation on top	
			[°C]	[mm]	natural	drained
Sweden	Degerö Stromyr	64°11'lat.,	+1.2	523	Sph. majus	Sph. balticum
		19°33'long.				
Finland	Lakkasuo	61°48'lat.,	+3	700	Sph.	Sph.
		24°19'long.			angustifolia	angustifolia
Germany	Breitlohmisse	48°41'lat.,	+7	835	Sph.	Sph.
(Black		8°25'long.			capilifolium	capilifolium
Forest)	Ursee	47°51'lat.,	+7	1600	-	-
		8°25'long.				
	Rotmeer	47°52'lat.,	+7	1600	Sph. rubellum	Sph. rubellum
		8°6'long.				patches

Table 2: Overview of studied mires; coordinates (lat./long.); mean annual temperature (MAT); annual precipitation (P);530Sphagnum mosses (Sph.) (Laine et al., 2004; Nielsson et al., 2008; DWD 2018, Alexandersson et al., 1991; Armbruster et al., 2003)





Table 3: Level of error probability according to T-tests of ∂¹⁵N measurements to identify significant differences (p<0.05, marked in red) between the drainage-affected and the natural sites; black border: normalized turning point of ∂¹⁵N, drained sites: Degerö
 (DDC), Lakkasuo (ombrotrophic (LOD), minerotrophic (LMD), Rotmeer (Ro2, Ro3), Ursee (Ur1), Breitlohmisse (Br2, Br3, Br4)

depth	all	DDC	LOD	LMD	Ro2	Ro3	Ur1	Br2	Br3	Br4
2	0.01	0.18	0.38	0.01	0.21	0.28	0.00	0.12	0.03	0.06
4	0.00	0.25	0.85	0.07	0.19	0.19	0.36	0.17	0.30	0.04
5	0.03	0.52	0.13	0.25	0.34	0.23	0.66	0.20	0.59	0.06
7	0.28	0.89	0.03	0.25	0.08	0.34	0.77	0.09	0.72	0.04
8	0.50	0.40	0.00	0.35	0.12	0.88	0.83	0.04	0.83	0.15
10	0.79	0.06	0.00	0.40	0.05	0.99	0.97	0.13	0.55	0.07
12	0.55	0.28	0.00	0.70	0.34	0.94	0.34	0.11	0.77	0.06
14	0.02	0.16	0.00	0.47	0.26	0.28	0.02	0.99	0.63	0.02
17	0.02	0.35	0.00	0.60	0.01	0.48	0.10	0.33	0.13	0.61
20	0.00	0.02	0.01	0.21	0.07	0.05	0.03	0.03	0.12	0.06
24	0.05	0.26	0.05	0.05	0.39	0.25	0.05	0.34	0.24	0.59
28	0.61	0.42	0.98	0.01	0.12	0.41	0.08	0.40	0.35	0.94
32	0.39	0.84	0.61	0.02	0.36	0.41	0.19	0.44	0.17	0.72
36	0.96	0.90	0.20	0.04	0.90	0.84	0.34	0.73	0.47	0.19
41	0.07	0.21	0.06	0.13	0.70	0.76	0.78	0.11	0.32	0.09
45	0.19	0.27	0.11	0.14	0.65	0.68	0.18	0.12	0.52	0.09
50	0.58	0.37	0.40	0.24	0.44	0.41	0.26	0.29	0.15	0.23
54	0.66	0.65	0.28	0.63	0.93	0.61	0.80	0.53	0.76	0.16
59	0.85	0.54	0.88	0.26	0.12	0.14	0.29	0.29	0.26	0.59
62	0.74	0.42	0.09	0.12	0.59	0.66	0.73	0.74	0.92	0.98
66	0.56	0.54	0.17	0.11	0.37	0.40	0.34	0.56	0.35	0.28
68	0.36	0.31	0.93	0.08	0.46	0.22		0.45	0.73	0.66





Table 4: Correlation coefficient r (Spearman) of trends for the whole core (= overall slope) and for separated sections; Degerö
(DDC), Lakkasuo (ombrotrophic (LOD), minerotrophic (LMD), Breitlohmisse (Br2, Br3, Br4), Rotmeer (Ro2, Ro3), Ursee (Ur1),
rewetted horizon (RW), upper mesotelm (UM), lower mesotelm (LM), catotelm (Cat)

	natural			drained		
site	overall	overall		section	I	
		_	RW	UM	LM	Cat
DDC	0.35	0.17	0.98	0.99	0.95	0.29
LOD	0.25	0.07		0.95	0.99	0.15
LMD	0.52	0.08	0.50	0.91	1.00	0.45
Br2	0.56	0.29		0.99	0.87	0.22
Br3	0.56	0.01		0.85	0.94	0.26
Br4	0.56	0.23		0.89	1.00	0.07
Ro2	0.00	0.11		0.73	0.50	0.63
Ro3	0.11	0.04		0.93	0.99	0.56
Ur1	0.13	0.00	0.01	0.73	0.98	0.04
average	0.34	0.11	0.50	0.89	0.91	0.30





Table 5: Trend values of all sites [‰/cm]; slopes are given for the whole core (=overall) and for separated sections (rewetted
horizon (RW), upper mesotelm (UM), lower mesotelm (LM), catotelm (Cat)); Degerö (DDC), Lakkasuo (ombrotrophic (LOD),
minerotrophic (LMD), Breitlohmisse (Br2, Br3, Br4), Rotmeer (Ro2, Ro3), Ursee (Ur1))

	natural			drained		
site	overall	overall	-	sec	tion	
			RW	UM	LM	Cat
DDC	9.66	5.83	-7.08	1.58	-2.04	11.24
LOD	8.47	6.14		4.06	-5.10	-9.49
LMD	25.18	-6.00	-16.95	3.17	-1.78	-10.53
Br2	23.82	8.39		2.19	-2.06	-8.80
Br3	23.82	-3.28		4.34	-2.25	6.43
Br4	23.82	9.06		3.22	-3.66	3.34
Ro2	3.21	-4.58		2.09	-2.78	-9.24
Ro3	4.83	-2.57		0.38	-1.89	-11.39
Ur1	7.76	0.98	2.52	13.76	-4.82	2.78



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Table 6: Humification Indices (HI) after von Post for all investigated sites; Degerö (natural (DNM), drained (DDC)), Lakkasuo (ombrotrophic natural (LON), ombrotrophic drained (LOD), minerotrophic natural (LMN), minerotrophic drained (LMD)), Breitlohmisse (natural (Br1), drained (Br2, Br3, Br4)), Ursee (natural (Ur2), drained (Ur1)), Rotmeer (natural (Ro1), drained (Ro2, Ro3))

site	natural	rewetted horizon	mesotelm	catotelm
DNM	H1-2			H3
DDC		H2-H3	H4	H3
LON	H2			H3
LOD			H4-H5	H3
LMN	H2			H3
LMD		H3	H4	H3
BR1	H3			H3
BR2			H4	H2-H3
BR3			H4	H3
BR4			H4	H3
UR1		H3	H3-H4	H2-H3
UR2	H2			H2
Ro1	H2			H3
Ro2			H4	H3
Ro3			H4	H3





Site	Horizon	Main species	Description
		species	Yellow, good preserved Sphturf, detached Sph.
	rewetted	Sph.	cymbifolia, Vaccinium oxycoccos Eriophorum
	horizon	balticum	vaginatum, Andromeda polifolia & Cladopodiella
			fluitans
Degerö			Darker, grayish; Sphturf, some Eriophorum
(DDC)	mesotelm	Sph.	vaginatum, detached Sph. cymbifolia, Vaccinium
· /		balticum	oxvcoccos, Andromeda polifolia
			Yellow: Sphturf. more Sph. cymbifolia, some
	catotelm	Sph.	Eriophorum vaginatum detached Vaccinium
	eutotenni	balticum	arveaceas Andromeda nalifalia
	upper	Snh	Dark brown: Sph_turf_mostly_Sph_ruballum_with
	masotalm	spn.	Dark orown, Spin-turi, mostly Spin. rubenam with
Lakkasuo	lower	Sph	Dark brown gravish: Sph turf mostly Sph
	magatalm	spn.	mikellum and Suk heltigum
(LOD)	mesoterm	rubellum	rubelium and spn. ballicum
	catotelm	Sph.	Light brown, yellow; Sphturf, mostly Sph
		rubellum	rubellum
	upper meostelm	Sph.	Brown; Sphturf mostly Sph. capilifolium and
		capilifolium	Sph. cymbifolia, much Ericaceous roots and some
	1 5		Eriophorum vaginatum stems
Breitlohmisse		Sph	Dark brown; Sphturf mostly Sph. capilifolium
(Br2)	mesotelm	cvmbifolia	and Sph. cymbifolia, some Ericaceous roots and
(512)		cymogona	Eriophorum vaginatum
		Snh	Lighter, reddish; Sphturf mostly Sph.
	catotelm	spn.	capilifolium and Sph. cymbifolia, some Sph.
		cymbijolia	acutifolia
		C L	Brown-reddish, yellow; Sphturf, mostly Sph.
	upper	Spn.	acutifolia, some Sph. rubellum, detached
	mesotiem	acutijolia	Eriophorum vaginatum
		C 1	Dark brown, grayish; Sphturf, mostly Sph.
Rotmeer (Ro2)	Sph.	cymbifolia, and Sph. acutifolia, some Sph.	
		cymbifolia	rubellum, detached Eriophorum vaginatum
			Reddish, yellow; Sphturf, mostly Sph.
	catotelm Sph. cymbifolia	Sph.	cymbifolia, some Sph. acutifolia, detached
		Erionhorum vaginatum	

Table 7: Description of vegetation of four of the study sites; Sphagnum mosses (Sph.)

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