



1           **Efficient Eco-Friendly Organic Wastes Mixed with Growth Promoting Bacteria to**  
2           **Remediate and Increase Fertility of Saline Sodic Soil in Egypt**

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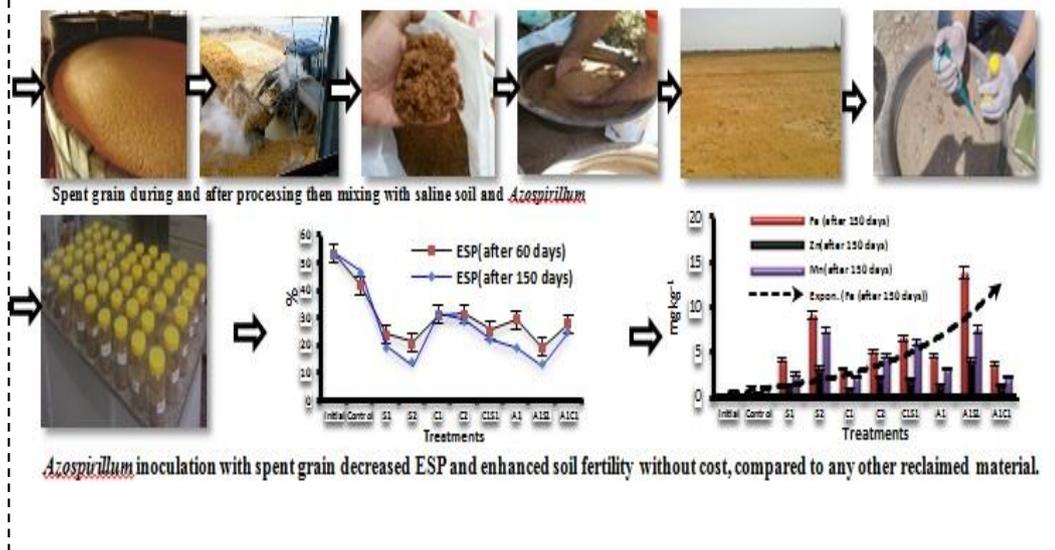
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**HIGHLIGHTS:**

- 1- Spent grain is a costless by-product which can be used for saline-sodic soil reclamation.
- 2- The combined application of *Azospirillum* with spent grain decreased the ESP in soil.
- 3- SOC, pH, Dehydrogenase, and Total acidity were the main factors affecting ESP and then influencing, CEC and Ex-Na<sup>+</sup>.
- 4- Variable-rate compost application was more expensive than the spent grain and *Azospirillum*.

**GRAPHICAL ABSTRACT:**



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28 **ABSTRACT**

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30 In Egypt, the total area of agricultural land is 3.36 million acres, which is 3.8 % of the entire  
31 territory of the country. One of the main obstacles to agricultural production in Egypt is soil  
32 salinization and degradation. Therefore, saline-sodic soil reclamation in arid regions is highly relevant.  
33 This study aimed to use *Azospirillum* inoculation with eco-friendly organic wastes for free remediation  
34 of saline-sodic soils. In this work nine treatments included two levels of spent grain (S1 and S2), two  
35 levels of compost (C1 and C2), a mix of both sources (C1S1), one level of *Azospirillum* (A1), a mix of  
36 both sources with *Azospirillum* (A1S1 and A1C1) and an untreated control. The treatments were  
37 previously incubated with soil at field capacity for five months under laboratory conditions at 28° C.  
38 The most relevant chemical and biological parameters were analysed every month for five months.  
39 Results indicate that *Azospirillum* inoculation with spent grain increased soil organic carbon (SOC),  
40 dehydrogenase and urease enzymes, micro-nutrients (Fe<sup>2+</sup>, Zn<sup>2+</sup>, Mn<sup>2+</sup>, Cu<sup>2+</sup> and B<sup>+</sup>), and macro-  
41 nutrients (N, P and K); while decreased exchangeable sodium percentage (ESP), pH and EC, by 75 %,  
42 12%, and 43 % respectively, compared to initial conditions. The significant variation was observed in  
43 chemical and biological properties among all treatments in the order of  
44 S2>A1S1>A1>C2>C1S1>A1C1>C1>control. In conclusion, the addition of *Azospirillum* with spent  
45 grain is highly recommended for amelioration of the saline-sodic soil and is more effective compared  
46 with compost to remediate saline-sodic soils.

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48 **Keywords:** Soil incubation, *Azospirillum*, amelioration, dehydrogenase, spent grain, saline-sodic

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**INTRODUCTION**

51 One of the main obstacles to agricultural production in Egypt is soil salinization, which lead  
52 to land degradation (Rashad *et al.*, 2010). Soil salinization and degradation subsequent to salinization  
53 is a key environmental impediment that severely influences agricultural productivity and sustainability  
54 in arid and semi-arid regions (Qadir *et al.*, 2008). Globally salt-affected soils account for  
55 approximately 46 million hectares (20%) of agricultural land (Ahmad *et al.*, 2013). More than 60 % of  
56 salt-affected soils classified as saline-sodic soils, the saline-sodic soils possess EC >4 dSm<sup>-1</sup>, pH >8.2  
57 and exchangeable sodium percentage (ESP) >15 % (Rath and Rousk, 2015). The remediation of saline-  
58 sodic soils requires removal of Na<sup>+</sup> ions from the exchange areas and leaching of the substituted Na<sup>+</sup>  
59 out of the root zone in percolating water (Sastre-Conde *et al.*, 2015; Zhang *et al.*, 2019). The chemical  
60 amendments as desulfurization gypsum open a source of Ca<sup>2+</sup> to substitute the exchangeable sodium  
61 from the cation exchange areas (Luo *et al.*, 2018), which plays a very important role in the ameliorant  
62 of saline-sodic soils. The worldwide expansion of saline-sodic soils is increasing due to



63 mismanagement and using low quality of irrigation water. The soil called saline-sodic due to its high  
64 ESP, cultivation of saline-sodic soils faces many challenges, such as poor soil structure, low soil  
65 stability aggregates, soil crusting, low hydraulic conductivity, low water infiltration, low water holding  
66 capacity, bulk density is high, alkalinity, high  $\text{Na}^+$  concentrations in soil solution, low soil fertility  
67 (Dodd *et al.*, 2013; Rath and Rousk, 2015), low biological activity, nutrient deficiencies, microbial  
68 community and soil fertility (Zhang *et al.*, 2019). Effects of salinity, including low agricultural  
69 productivity, low economic returns and soil erosion (Hu and Schmidhalter, 2005). The salt effects are  
70 the results of complex interactions between the physiological, biochemical and morphological  
71 processes, including plant growth, nutrients, seed germination, and water absorption (Shrivastava and  
72 Kumar, 2015; Zhang *et al.*, 2019). Soil salinity imposes ionic toxicity, osmotic pressure and nutrients  
73 (N, Ca, K, P, Fe, and Zn) deficiency and oxidative stress on plants, thus limiting the absorption of  
74 water from the soil. Soil salinity significantly reduces the absorption of phosphorus (P) due to  
75 phosphate ions are deposited with calcium ions (Bano and Fatima, 2009). Some elements, such as  $\text{Na}^+$ ,  
76  $\text{Cl}^-$ , and  $\text{B}^+$ , have specific toxic effects to plants. Excessive accumulation of  $\text{Na}^+$  in cell walls can  
77 quickly lead to osmotic pressure and cell death (Munns, 2002). Soil organic amendments can improve  
78 soil physical and chemical properties and microbial activity in saline-sodic (Luo *et al.*, 2018; Zhen *et*  
79 *al.*, 2014)

80         Application of agro-industrial organic wastes as soil amendments such as the Brewers' spent  
81 grain, 85 % of the total by-products generated and compost with bacteria inoculation to reclamation  
82 saline soil. Organic additives have been used from the residues of the beryl industry, these wastes are  
83 inexpensive, and their accumulation causes severe damage to the environment (Hafez *et al.*, 2019;  
84 Mussatto, 2014; Nassar *et al.*, 2014). Spent grain has a negative value for environmental quality due to  
85 its accumulation as industrial waste. It has been recently applied as an environment-friendly organic  
86 waste to the soil, as well as composting of agricultural biomass, to improve physical and chemical  
87 properties of calcareous and sandy soils (Hafez *et al.*, 2019; Rashad *et al.*, 2016). Spent grain is acidic,  
88 rich in macro and micronutrients, amino acids, and vitamins as well as cellulose, hemicelluloses, lipid,  
89 protein, arabinoxylan, ash, and lignin (Mussatto and Roberto, 2006). Therefore, it is a good candidate  
90 for recycling in agriculture. The environmental systems using a low level of external input have  
91 become more important all over the world in the last few years from natural resources, and reduction  
92 of environmental systems degradation (Mader, 2002)

93         In this context, organic wastes incubation with bacteria encourages the use of agricultural  
94 practices and alternative technologies, according to economic situation, social and environmental  
95 conditions of the region the system is in particular because it refers to fertilization, which is usually



96 done by application of organic matter with *Azospirillum* bacteria such as spent grain, compost, plant  
97 residues, or manure, a lot of which are produced within the same farm (Hafez *et al.*, 2019).

98           On the other hand, application of eco-friendly organic wastes and bacteria to the saline-sodic  
99 soil not only generates a better food situation, but moreover, positively affects other properties such as  
100 soil particle assembly, water holding capacity and increase biological activity (Hafez *et al.*, 2019;  
101 Hayat *et al.*, 2010; Scherer *et al.*, 2011), contribute to generating high production, even with low  
102 application or none of the fertilizer. In the aspect incubation system for organic wastes with micro-  
103 organisms are found in soil (Zhuang *et al.*, 2007) incubated decreased soil salinity and increased the  
104 content of nitrogen, phosphorus, and potassium (Rashad *et al.*, 2016; Rojas-Tapias *et al.*, 2012). The  
105 application of biofertilizers, plant growth promoters and organic amendments enhancing soil fertility,  
106 microbial activities and salt-affected. In addition, plant growth-promoting bacteria (PGPB) have been  
107 recently applied as a biofertilizer in soil to increase nitrogen fixation and phosphate availability and  
108 reduce soil salinity (Hafez *et al.*, 2019; Rojas-Tapias *et al.*, 2012; Zhen *et al.*, 2014; Shrivastava and  
109 Kumar, 2015; Zarea *et al.*, 2012).

110           One of the PGPB is the *Azospirillum brasilense*, this is free-living bacteria capable of  
111 affecting the growth and seed yield of numerous plant species and many agronomic and ecological  
112 significance. Plants treated with PGPB producing bacteria display increased resistance to salinity stress  
113 due to improved soil structure (Karagöz *et al.*, 2012; Zhuang *et al.*, 2007). PGPB can also bind to  
114 cations including Na<sup>+</sup> thus making it unavailable to plants under saline conditions. (Zhao *et al.*, 2018).  
115 Egypt, is located in the arid region therefore, the study of salinity in such dry land area has become of  
116 great interest. We hypothesized that the applied spent grain with *Azospirillum* inoculation, for tow  
117 period 60 and 150 days of incubation, would reclaimed saline-sodic soil while increasing soil fertility  
118 and reduce the ESP. The objectives of this work were to (i) the effect of eco-friendly organic wastes  
119 doses with *Azospirillum* inoculation on some chemical and biological properties of saline-sodic soil in  
120 the North-Western part of Egypt on reduce soil salinity, (ii) increase the content of beneficial microbes  
121 and activity of enzymes, and (iii) increase soil fertility in regards to the SOC improve the availability  
122 of macro and micro-nutrients in saline-sodic soil after 60 and 150 days of incubation.

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## 2-MATERIALS AND METHODS

### 125 2.1. Site Information

126 The study sites are located at the farm of El-Bangar village 28, El-Hamam City, Matrouh  
127 Governorate, North-Western part of Egypt (30°45'44.47"N 29°25'26.54"E). The soil samples are  
128 saline-sodic typically mixed with rocks. Due to its inadequate soil, water regime, salinity in arid  
129 regions can occur when the water table is between two and three meters from the surface of the soil.  
130 The use and potential effect of soil enhancer materials are important and current raise. The study site



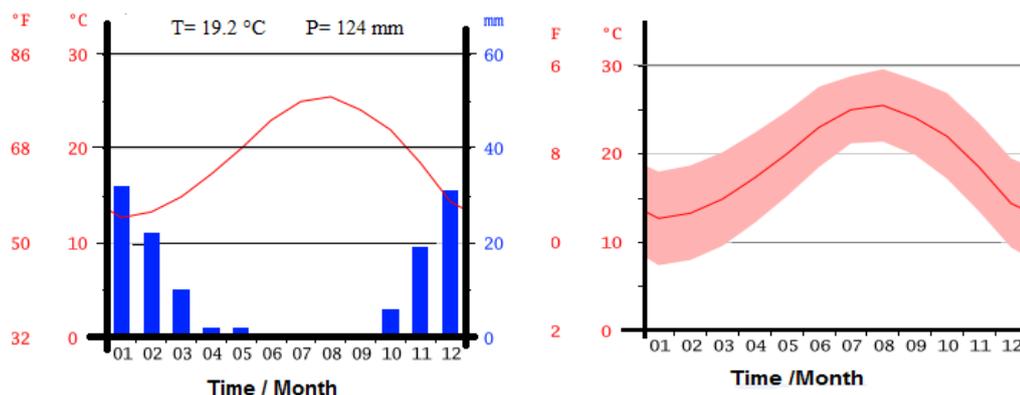
131 has a desert climate arid, hot summer and semi-cool and semi-wet winters. The climate is classified  
 132 as BWh by the Köppen-Geiger system (Climate-Data.org, 2019). The average air temperature in El-  
 133 Hamam is 19.2 °C. In a year, Annual precipitation is 124 mm. The temperatures are higher on  
 134 average in August, at around 27.5 °C. In January, is the coldest month the average temperature is  
 135 12.7 °C. It is the lowest average temperature of the whole year (Fig 1). The soil samples were  
 136 granules with diameters less than 2- mm, and used after the sieving process. The samples of the  
 137 airdried soil were used for the physical and chemical parameters of the saline-sodic soil are shown in  
 138 (Table 1).

139 (Table.1) Initial physical and chemical parameters of saline-sodic soil used for the study

Soil parameter	Saline-sodic soil
pH (1:2.5 w: w)	8.84 ± 0.05
EC <sub>c</sub> (dS m <sup>-1</sup> )	5.43 ± 0.10
Total N (gkg <sup>-1</sup> )	0.02 ± 0.001
Available P (mgkg <sup>-1</sup> )	1.20 ± 0.07
Available K <sup>+</sup> (mgkg <sup>-1</sup> )	78.1 ± 0.44
Total CaCO <sub>3</sub> (%)	18.6 ± 1.35
CEC (Cmol <sup>+</sup> kg <sup>-1</sup> )	7.56 ± 0.24
Organic Matter (g kg <sup>-1</sup> )	1.78 ± 0.21
Soil Organic Carbon (g kg <sup>-1</sup> )	1.03 ± 0.01
Total DOC (%)	0.012 ± 0.003
ESP (%)	53.1 ± 1.93
C/N Ratio	49.14 ± 2.1
Sand (%)	31.2 ± 0.11
Silt (%)	23.5 ± 0.21
Clay (%)	45.3 ± 0.18
Texture	Clay Loam
<b>Micronutrients DTPA Extractible (mgkg<sup>-1</sup>)</b>	
Fe <sup>2+</sup>	0.08 ± 0.01
Zn <sup>2+</sup>	Nd*
Mn <sup>2+</sup>	0.11 ± 0.020
Cu <sup>2+</sup>	0.003 ± 0.01
B <sup>+</sup>	Nd*
Cl <sup>-</sup>	192 ± 0.166

Data correspond to means of four replicates ± standard deviation. Nd<sup>\*</sup>= not detected

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(Fig. 1). Annual precipitation and temperature study area of soil (Climate-Data.org, 2019).  
<https://en.climate-data.org/africa/egypt/alexandria-governorate/alexandria-515/>



153 **2.2. Soil Treatments and *Azospirillum* preparation**

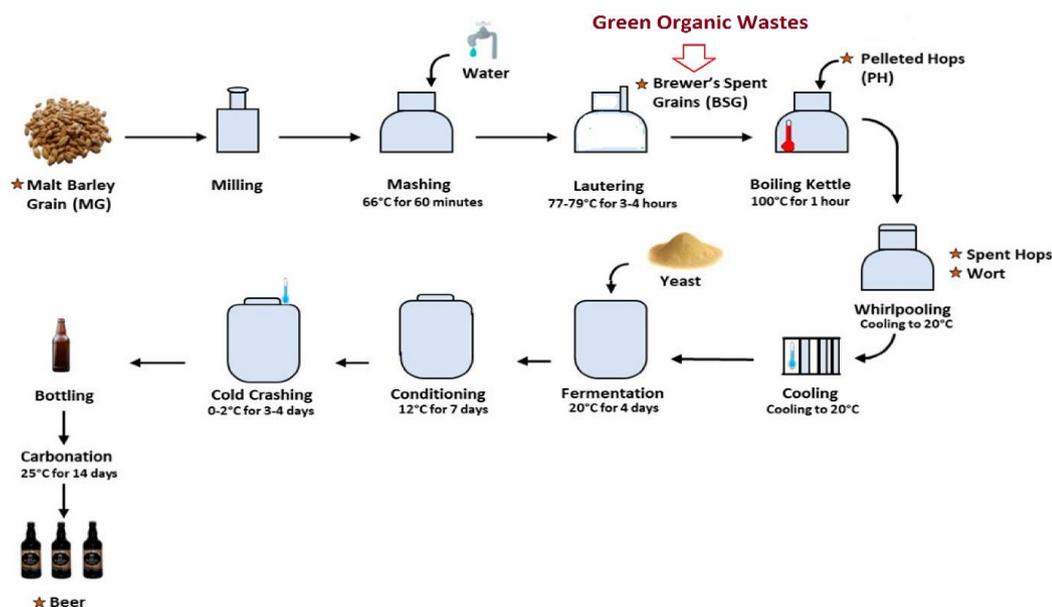
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155 Three types of treatments (*Azospirillum*, compost and spent grain) were used. The  
156 *Azospirillum* bacteria, it was from Faculty of Agriculture, Ain Shams, Cairo Governorate, Egypt. The  
157 compost consisted of plant and animal wastes from the national factory for the production of  
158 compost, Alexandria, Governorate, Egypt. The spent grain, a by-product from beer industry, was  
159 obtained from Al Ahram Beverages Company, Abu Hammad, Al Sharkia Governorate, Egypt  
160 (Schema, 1). The characteristics of the treatments were determined according to the standard  
161 procedures shown in (Table 2). *Azospirillum brasilense* (Sp245) was cultured; growth and  
162 inoculation were performed as described by (Pii *et al.*, 2015b). Briefly, the *Azospirillum* bacterium  
163 was grown for 4 days in LB medium (10g L<sup>-1</sup> triptone, 5g L<sup>-1</sup> NaCl, 10g L<sup>-1</sup> yeast extract) with  
164 continuous shaking at 30°C. Bacteria were then centrifuged for 15 min at 4500 x g and washed four  
165 times with sterile saline-sodic solution (NaCl 0.85% w/v) after that the *Azospirillum* bacteria became  
166 ready for inoculation with the soil.

167 **2.3. Experimental design and treatments management**

168 The experiment was carried out during the summer of 2017 laboratory of the Department of  
169 Land and Water Technologies, City of Scientific Research and Technological Applications (SRTA-  
170 City), and analyses were done between the (SRTA-City) and laboratory of the Department of Soil  
171 Science and Soil Ecology, Saint Petersburg State University (SPBSU). The incubation experiment  
172 was nine treatments included two levels of spent grain, were referred to 12.8 gkg<sup>-1</sup> soil (S1) and 25.5  
173 gkg<sup>-1</sup> soil (S2); two levels of compost, were referred to 23.5 gkg<sup>-1</sup> soil (C1), and 47 gkg<sup>-1</sup> (C2); one  
174 level of *Azospirillum brasilense* inoculation (A1) was 5 % by weight of the soil; a mix of both  
175 sources spent grain and compost (C1S1); a mix of both sources with *Azospirillum* (A1C1) and  
176 (A1S1); all treatments were used and compared to the control without amendments. All treatments  
177 were incorporated with bottle 400g of saline-sodic soil. To let added treatments to affect soil  
178 properties, all bottle treatments were incubated under laboratory conditions, without plants, for five  
179 months at 28° C. The bottles were distributed in a randomized complete block design with four  
180 replications of each treatment. All bottles were watered with tab deionized-water once to field  
181 capacity of the soil. The most relevant chemical and biological parameters were analyzed every  
182 month for five months.

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(Schema 1). Schematic beer brewing process and product eco-friendly organic wastes “spent grain” description of extraction method by (Lordan *et al.*, 2019).

(Table, 2). Organic wastes characteristics (oven-dry weight basis).

Soil parameter	Compost	Spent grain
pH (1:5 w:w)	7.20 ± 0.01	4.16 ± 0.03
EC (dS/m, 1:5 w:w)	5.81 ± 0.21	1.45 ± 0.21
Organic Matter (g kg <sup>-1</sup> )	332 ± 1.23	750 ± 0.57
Total N (%)	2.10 ± 0.32	3.12 ± 0.68
Total P (%)	1.03 ± 0.52	1.86 ± 0.54
Total K (%)	0.57 ± 0.01	1.74 ± 0.63
C: N ratio	9.16 ± 0.35	13.9 ± 0.12
Organic carbon (%)	19.25 ± 0.12	43.5 ± 0.94
Fe <sup>2+</sup> (mgkg <sup>-1</sup> )	960 ± 2.97	1130 ± 3.87
Zn <sup>2+</sup> (mgkg <sup>-1</sup> )	220 ± 5.34	368 ± 2.34
Mn <sup>2+</sup> (mgkg <sup>-1</sup> )	100 ± 2.14	210 ± 1.98
Cu <sup>+</sup> (mgkg <sup>-1</sup> )	61 ± 1.21	98 ± 1.54

Data correspond to means of four replicates ± standard deviation

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### 3.4. Sampling

#### 3.4.1. Soil physicochemical analyses

The total thirty-six samples referred to nine treatments × four replications, soil samples it was taken from each bottle after the removal of visible roots and fresh litter material, the composite samples were sieved (2 mm) and then stored at room temperature for less than three months until chemical analyses were performed. Samples were air-dried and ready for measured. The soil pH was measured potentiometrically in a soil : water suspension 1:2.5 w/v (Page, *et al.*, 1982). The electrical



199 conductivity ( $EC_e$ ) was measured in saturated paste extracts using an EC meter (Corwin and Yemoto,  
200 2017). The total nitrogen (N), available phosphorus (P), and available potassium (K) were measured  
201 using (Anderson, *et al.*, 1982). Micronutrients ( $Fe^{2+}$ ,  $Zn^{2+}$ ,  $Mn^{2+}$ ,  $Cu^{2+}$ , and  $B^+$ ) were extracted by  
202 DTPA solution as explained by (Soltanpour and Schwab, 1977) then measured with inductively  
203 coupled plasma (ICP) atomic emission spectroscopy. Exchangeable sodium percentage (ESP) of soil  
204 was determined using the relationship  $ESP = (\text{exchangeable sodium (Ex Na}^+) / \text{cation exchange}$   
205  $\text{capacity (CEC)}) * 100$ . Extractable sodium was extracted with 1 M ammonium acetate solution  
206 (Normandin *et al.*, 1998). Soil CEC was estimated following the Bower saturation method as outlined  
207 by (Richards, 1954).

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#### 209 **3.4.2 Biological analyses**

210 Dehydrogenase activity was determined by triphenyltetrazolium chloride (TTC) as described by  
211 (Casida *et al.*, 1964). Soil urease (S-UE) was determined using methods previously published by  
212 (Kandeler and Gerber, 1988). Soil organic carbon (SOC) concentration was determined by  
213 oxidization with  $K_2Cr_2O_7$  (Ouyang *et al.*, 2013).

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#### 215 **Statistical analysis**

216 Statistical analysis was carried out using of the SPSS v.16 software (Visauta, 2007). Data  
217 were submitted to a normality test before the analysis of variance. When statistical significance was  
218 found ( $P \leq 0.05$ ), a comparison of the means was carried out by using the Tukey test. Furthermore, a  
219 Pearson correlation analysis was carried out to observe the degree of association between some of the  
220 studied variables.

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## **RESULTS AND DISCUSSION**

223 Several factors could control eco-friendly organic wastes decomposition. In the present  
224 study, organic types, decomposition condition, and soil properties are the dominant factors that  
225 control the decomposition processes to reclamation saline-sodic soil. The concentrations of nutrients  
226 in the incubated soils were determined as an indication for soil fertility. The chemical and biological  
227 properties are discussed and presented separately. Chemical properties, including soil acidity (pH),  
228 electrical conductivity in saturated paste extract ( $EC_e$ ), exchangeable sodium percent (ESP), cation  
229 exchange capacity (CEC), exchangeable sodium (Ex- $Na^+$ ), soil organic carbon (SOC), humic  
230 substance (HS), total acidity (TA), available phosphorus (P), available potassium ( $K^+$ ), total nitrogen



231 (N), micro-nutrients ( $\text{Fe}^{2+}$ ,  $\text{Zn}^{2+}$ ,  $\text{Mn}^{2+}$ ,  $\text{Cu}^{2+}$ , and  $\text{B}^+$ ), dehydrogenase and urease enzymes in saline-  
 232 sodic soil after 60 and, 150 days of incubation.

233 **3.1. Chemical properties changed as a soil reclamation indicator after the incubation period**

234 The soil chemical properties including, pH, ESP,  $\text{EX-Na}^+$ ,  $\text{EC}_e$  and CEC after 60 and 150  
 235 days of incubation, were significantly changed with increasing eco-friendly organic wastes (Table 3).  
 236 Soil soluble and exchangeable salts content statistically significantly decreased with increasing eco-  
 237 friendly organic wastes and *Azospirillum* application ( $P < 0.05$ ).

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239 Table (3): Soil salinity indicator in saline-sodic soil after 60 and 150 days of incubation

Treatments	Incubation Period	pH	$\text{EC}_e$	$\text{Ex-Na}^+$	CEC	ESP
	days					
<b>Initial</b>	<b>Zero time</b>	<b>8.84±0.05</b>	<b>5.43±0.10</b>	<b>4.00±0.18</b>	<b>7.56±0.24</b>	<b>53.12±1.93</b>
Control	60	8.76±0.05	5.42±0.10	3.39±0.21	7.93±0.24	42.78±1.93
	<b>150</b>	<b>8.94±0.08</b>	<b>5.48±0.02</b>	<b>3.83±0.28</b>	<b>8.26±0.64</b>	<b>46.49±4.16</b>
S1	60	8.34±0.12	4.11±0.07	3.24±0.29	14.23±0.49	22.76±1.46
	<b>150</b>	<b>8.03±0.06</b>	<b>3.66±0.23</b>	<b>2.90±0.26</b>	<b>15.03±0.25</b>	<b>19.28±1.57</b>
S2	60	8.10±0.11	3.10±0.12	3.06±0.11	16.20±0.83	18.91±1.73
	<b>150</b>	<b>7.83±0.10</b>	<b>3.00±0.17</b>	<b>2.43±0.06</b>	<b>17.80±0.2</b>	<b>13.66±0.2</b>
C1	60	8.51±0.04	5.17±0.13	3.96±0.058	12.33±0.31	30.67±0.33
	<b>150</b>	<b>8.60±0.06</b>	<b>5.42±0.10</b>	<b>4.46±0.15</b>	<b>14.43±0.51</b>	<b>30.99±2.01</b>
C2	60	8.47±0.04	5.41±0.21	4.57±0.44	15.67±0.40	29.34±2.09
	<b>150</b>	<b>8.58±0.05</b>	<b>5.70±0.53</b>	<b>4.63±0.56</b>	<b>16.03±0.25</b>	<b>28.86±3.14</b>
C1S1	60	8.17±0.36	4.65±0.32	5.50±0.20	14.90±0.20	26.90±0.35
	<b>150</b>	<b>8.06±0.43</b>	<b>4.67±0.02</b>	<b>3.50±0.26</b>	<b>15.93±0.11</b>	<b>21.97±1.76</b>
A1	60	7.98±0.03	3.84±0.18	2.19±0.30	8.83±0.24	24.89±4.14
	<b>150</b>	<b>7.86±0.13</b>	<b>3.54±0.35</b>	<b>1.75±0.18</b>	<b>9.26±0.92</b>	<b>19.13±3.72</b>
A1S1	60	7.93±0.06	4.51±0.08	3.06±0.11	15.83±0.29	19.36±0.63
	<b>150</b>	<b>7.89±0.05</b>	<b>4.03±0.07</b>	<b>2.07±0.11</b>	<b>16.12±0.33</b>	<b>12.85±0.49</b>
A1C1	60	8.24±0.18	4.674±0.03	4.00±0.15	14.25±0.35	28.07±0.83
	<b>150</b>	<b>8.23±0.01</b>	<b>4.485±0.03</b>	<b>3.54±0.36</b>	<b>14.30±0.42</b>	<b>24.72±1.83</b>
<b>LSD<sub>0.05</sub></b>	60	0.275	0.892	2.051	0.827	13.72
	<b>150</b>	<b>0.278</b>	<b>0.413</b>	<b>0.475</b>	<b>0.801</b>	<b>4.180</b>

240 Data correspond to means of four replicates ± standard deviation.  $\text{LSD}_{0.05}$ = least significant difference (LSD) test  
 241 at ( $p \leq 0.05$ ).

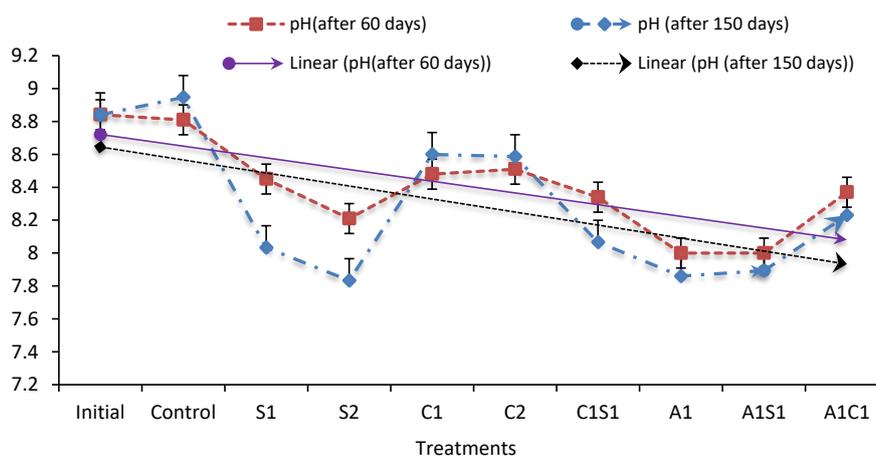
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243 **3.1.1. The soil pH changed after the incubation period**

244 The influences of the eco-friendly organic wastes incubation with *Azospirillum* inoculation  
 245 on the decrease soil pH compared to initial condition and control treatment followed the order: S2  
 246 (spent grain  $25.5 \text{ gkg}^{-1}$ ) > A1 (5% of soil weight) > A1S1 (Mix A1 and S1) > S1 (spent grain  $12.8$   
 247  $\text{gkg}^{-1}$ ) > C1S1 (Mix C1 and S1) > C2 ( $47 \text{ gkg}^{-1}$ ) > C1 (Compost  $23.5 \text{ gkg}^{-1}$ ). Shown (Fig 2) soil pH  
 248 after 60 days of incubation was higher in the soil initial condition compared to the spent grain (S2)  
 249 and *Azospirillum* (A1), treatments were 8.10, 7.93 and 7.98, for the treatments S2, A1S1, and A1  
 250 respectively, ( $P < 0.05$ ). While the corresponding values after 150 days of incubation were 7.83, 7.86  
 251 and 7.89, for the S2, A1 and A1S1 respectively, ( $P < 0.05$ ). The inherent soil pH values after 150 days



252 were between 7.8 and 7.89 are ideal for P-availability, while pH values below 5.5 and between 7.5  
253 and 8.5 limits P-availability to plants due to fixation by aluminum, iron, or calcium (Rashad *et al.*,  
254 2016). The S2 treatment affected positively the soil pH greater than compost treatments and control.  
255 Accordingly, this reduction in the pH values might be due to increase of the groups of carboxyl and  
256 phenol as a result of decomposing of eco-friendly organic wastes and the reduction of bicarbonate in  
257 the soil. (Hafez *et al.*, 2019) reported that a decrease of pH might be due to the addition of spent grain  
258 only or with *Azospirillum* inoculation in calcareous and sandy soils reduces soil pH sharply. The  
259 difference in the pH values between the S2 and the A1S1 was a highly significant camper to initial  
260 and control. The remediation of saline-sodic soils by organic matter sources has long been known to  
261 facilitate the reclamation of saline-sodic soils (Sastre-Conde *et al.*, 2015).



262  
263 Figure (2): Effect of *Azospirillum* and eco-friendly organic wastes on pH in saline-sodic soil after 60  
264 and 150 days of incubation. Data points and error bars represent means and standard errors (n = 4).

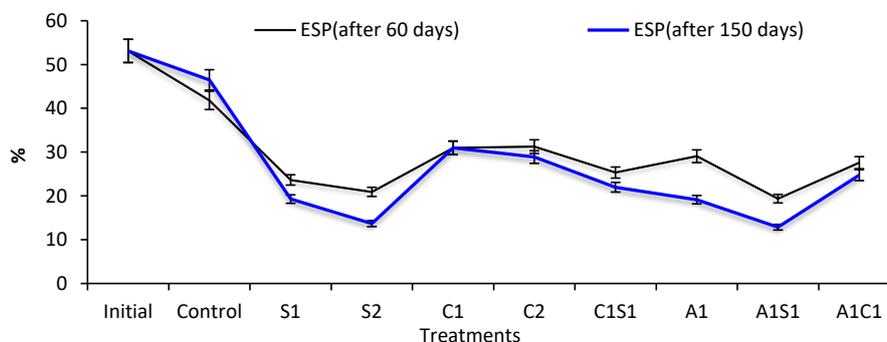
### 265 3.1.2. Soil ESP indicator

266 The soil ESP was decreased with the spent grain and *Azospirillum* increasing (P<0.05). The  
267 soil ESP was statistically significantly higher in the inoculation *Azospirillum* with spent grain in  
268 saline-sodic soil than that in the compost and control treatments (P<0.05). In the present study, the  
269 middle and high salinities exhibited greater salt content and ESP, and less Table (3) and Figure (3).  
270 The ESP after 60 days of incubation in the saline-sodic soil was 53.10, 19.35, 20.91 and 23.64 % for  
271 the initial condition, A1S1, S2, and S1, respectively. The corresponding percentages for the 150 days  
272 was similar to the trend of pH were 53.10, 12.85, 13.66 and 19.28 % for the initial condition, A1S1,  
273 S2, and S1, respectively. The addition of spent grain with *Azospirillum* (A1S1) to the saline-sodic  
274 soil contributed a statistically significantly lower amount of salt to the soil. The spent grain decreased  
275 the amount of exchangeable Na<sup>+</sup> and salinity than the compost and initial condition, this reducing for



276 ESP due to mixing *Azospirillum* with spent grain enhancement of the soil activity of carbon  
277 degrading extracellular enzymes and microbial decomposition rates. There was a significantly  
278 positive correlation between soil salt content and eco-friendly organic wastes amendments. Soil  
279 microbes are essential in soil carbon mineralization and accumulation, which determines soil fertility.

280 The *Azospirillum* increased microbial biomass carbon which decreased the exchangeable  
281  $\text{Na}^+$  in soil solution, this result was positively correlated with organic wastes incubation in saline-  
282 sodic soil. The addition of organic matter played the very important role for decreased soil salinity  
283 and enhancement the soil microbial biomass, this result agreement with (Zhang *et al.*, 2015, 2019;  
284 Zhen *et al.*, 2014) demonstrated that the low soil microbial carbon was attributed to the low soil  
285 organic carbon that resulted from little carbon input into the soils for the scarcity of vegetation under  
286 high salt condition. (Mahmoodabadi *et al.*, 2013; Sastre-Conde *et al.*, 2015) attributed the decrease of  
287 soil EC, SAR, and ESP by organic wastes and chemical amendments under saline-sodic soil. (Meena  
288 *et al.*, 2019) further demonstrated that lower organic matter in saline soil is more likely to be due to  
289 the decreased organic sources than the increased decomposition rate.



290

291 Figure (3): Effect of *Azospirillum* and eco-friendly organic wastes on ESP in saline-sodic soil after 60  
292 and 150 days of incubation. Data points and error bars represent means and standard errors ( $n = 4$ ).  
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### 294 3.1.3. Soil exchangeable $\text{Na}^+$ and electrical conductivity

295 The soil  $\text{Ex-Na}^+$  was statistically significantly higher in the inoculation *Azospirillum* with  
296 spent grain in saline-sodic soil than that in the other treatments ( $P < 0.05$ ). In the present study, the  
297 saline-sodic soil before amendments incubation was contained high content of the  $\text{Ex-Na}^+$  and ESP,  
298 and less organic matter compared to the after amendments incubation (Table 3). The effect of spent  
299 grain levels and *Azospirillum* treatments on  $\text{Ex-Na}^+$  concentration were consistently higher ( $P < 0.05$ )  
300 than compost levels, control and the initial (Table 3). The general influences of the *Azospirillum*  
301 inoculation with eco-friendly organic wastes incubation on the decrease in the soil  $\text{Ex-Na}^+$   
302 concentration compared to initial condition and control treatment, followed the order:  $\text{A1} > \text{A1S1} > \text{S2}$   
303  $> \text{S1} > \text{C1S1} > \text{control} > \text{initial} > \text{C2} > \text{C1}$ . Shown (Table 3), the soil  $\text{Ex-Na}^+$  for compost treatments



304 after 60 days of incubation were 5.0, 3.80 and 4.0 meq100g<sup>-1</sup>, for the C2, C1 and A1C1 respectively,  
305 (P<0.05). While the corresponding values after 150 days of incubation was higher in the compost  
306 only and compost with *Azospirillum* treatments compared to the initial condition and control, were  
307 4.6, 4.4 and 3.54 meq100g<sup>-1</sup> for the treatments C2, C1 and A1C1 respectively. In the control and  
308 initial Ex-Na<sup>+</sup> ranged between 3.4 and 4 meq100g<sup>-1</sup> after 150 days of incubation, while in A1, after  
309 150 days Ex-Na<sup>+</sup> was 1.75 meq100g<sup>-1</sup>. This confirms that soil salinization was effectively maintained  
310 through the control treatment irrigation with saline water, low rainfall since Ex-Na<sup>+</sup> decreased in A1  
311 from 4 in initial to 1.75 meq100g<sup>-1</sup> for spent grain and *Azospirillum* treatments. The relative  
312 reduction in Ex-Na<sup>+</sup> concentration after 150 days corresponded to 56.25 and 48.25% for A1 and  
313 A1S1 respectively, compared to the initial condition. Obviously, as the application levels of compost  
314 increased, the EC<sub>e</sub> and Ex-Na<sup>+</sup> concentrations increased (Table 3). There were significant differences  
315 among the treatments in the Ex-Na<sup>+</sup> concentration.

316 On the other hand, the higher concentrations of EC<sub>e</sub> and Ex-Na<sup>+</sup> can lead to deterioration of  
317 soil physical properties. The compost treatments was negative a contributed statistically significantly  
318 higher amount of EC<sub>e</sub> and Ex-Na<sup>+</sup> to the soil compared the control. I think this result due to the  
319 compost treatments is contained very high salts content before adding to soil, resulting in increased  
320 Na<sup>+</sup> concentration in the compost residues. Therefore, the application of pig compost could increase  
321 soil salinity in areas of low precipitation and drainage where the salts are not leached through the soil  
322 profile or on borderline saline soils. Therefore, the characterization of organic wastes is mandatory  
323 before adding to soil.

324 On the other hand the *Azospirillum* with spent grain treatments decreased amount of EC<sub>e</sub> and  
325 Ex-Na<sup>+</sup> than the compost and initial condition, this reducing for the Ex-Na<sup>+</sup> due to spent grain with  
326 *Azospirillum* enhancement of the soil microbial activity of carbon-degrading extracellular enzymes,  
327 increased humic substance, chelating the sodium in soil to sodium humate form and enhancement of  
328 CEC under A1S1 and A1 treatments in saline-sodic soil incubated. Similar results were reported by  
329 (Mahmoodabadi *et al.*, 2013; Rojas-Tapias *et al.*, 2012; Shrivastava and Kumar, 2015; Zhang *et al.*,  
330 2015) who reported that the applied organic wastes and PGPB, depending on their chemical  
331 composition might have a positive impact on soil properties, especially Na<sup>+</sup> and EC. (Ashraf *et al.*,  
332 2004), who found that inoculation with exopolysaccharide- producing bacteria could restrict Na<sup>+</sup>  
333 influx into roots. Further, (Zhang *et al.*, 2008) reported that inoculation with *Bacillus* bacteria could  
334 mediate the level of salt tolerance in soil.

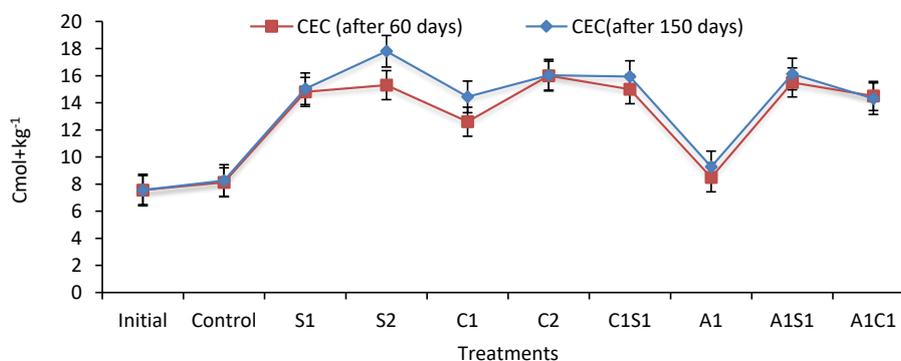
### 335 3.1.4. Cation Exchangeable Sodium Percentage (CEC)

336 A progressive increase in CEC was observed in all treatments due to the accumulation of  
337 Ca<sup>2+</sup> and Mg<sup>2+</sup> release with the eco-friendly organic wastes amendments. In particular, CEC



338 increased significantly during the spent grain and *Azospirillum* treatments ( $P < 0.05$ ), when it reached  
339 15.5 and 15.3  $\text{Cmol}\cdot\text{kg}^{-1}$  in A1S1 and S2 (Table 3) after 60 days of incubation. The A1S1 treatment  
340 increased CEC and decreased  $\text{Ex}\text{-Na}^+$  led to a significant reduction ( $P < 0.05$ ) for the ESP indicator  
341 (Table 3). On the other hand, the corresponding percent for the 150 days was similar to the trend of  
342 SOC were 7.50, 17.80, 16.12 and 16.01  $\text{Cmol}\cdot\text{kg}^{-1}$  for the initial condition, S2, A1S1, and C2,  
343 respectively. As a result, in soil, no significant differences were observed between *Azospirillum*, eco-  
344 friendly organic wastes, and the control. The effect of *Azospirillum* inoculation with spent grain on  
345 the CEC was confirmed when comparing to CEC of the different treatments of compost, (Fig 4).  
346 CEC increased by 137.3, 115 and 113.7% in S21, A1S, and C2 respectively, after 150 days of soil  
347 incubation. Generally, the reported concentrations in organically treated soil with application rate of  
348 *Azospirillum*, spent grain and compost are safe and sufficient for macronutrient consumption. This is  
349 results in agreement with other researchers (Shrivastava and Kumar, 2015; Trivedi *et al.*, 2017).

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Figure (4): Effect of *Azospirillum* and eco-friendly organic wastes on CEC in saline-sodic soil after 60 and 150 days of incubation. Data points and error bars represent means and standard errors ( $n = 4$ ).

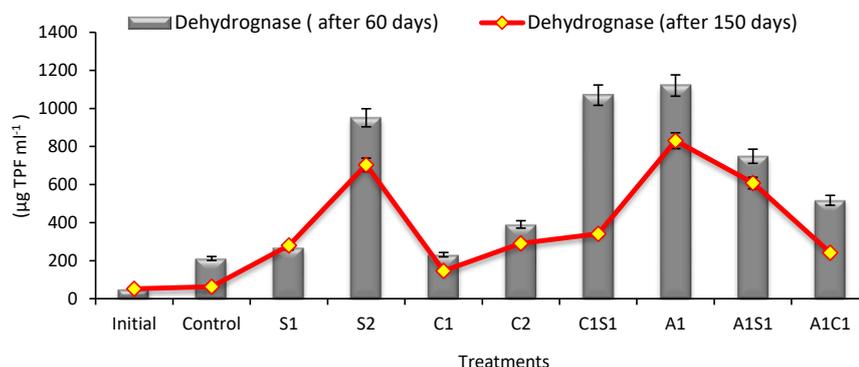
### 356 3.2. Soil biological parameters after incubation period

#### 357 3.2.1. Dehydrogenase and urease activities

358 The soil biological parameters dehydrogenase and urease activities showed in Figure  
359 (5) and figure (6) revealed that there were significant differences in dehydrogenase and urease  
360 activities in saline-sodic soil between all treatments, and also significant differences were recorded  
361 when using different levels of eco-friendly organic wastes and *Azospirillum* concentrations followed  
362 the order  $\text{A1S1} \geq \text{S2} \geq \text{A1} > \text{C1S1} > \text{C2} > \text{A1C1} \geq \text{S1} > \text{C1} > \text{control} > \text{initial}$ . Also, the differences in  
363 dehydrogenase and urease activities were highly significant ( $P < 0.05$ ) when combined with different  
364 levels of spent grain with *Azospirillum* or only spent grain, compared with different levels of compost  
365 and control. This is results in agreement with other researchers (Liang *et al.*, 2014; Meena *et al.*,



366 2019), in soils treated with farmyard manure; dehydrogenase activity is increasing. The high organic  
 367 matter content enhanced the production of microbial activity as a result of soil organic carbon during  
 368 the mineralization process of organic matter. Generally, the mineralized organic carbon increased as  
 369 the application rate of spent grain with *Azospirillum* increased. (Liang *et al.*, 2014) reported that  
 370 application of organic amendments increased microbial activity and urease activities in all particle-  
 371 size fractions as compared to those in control. Our results are in the same direction with the findings  
 372 of (Rojas-Tapias *et al.*, 2012; Scherer *et al.*, 2011; Shi *et al.*, 2019) which proposed that the  
 373 treatments with high organic amendments might exhibit the greatest microbial activity and  
 374 dehydrogenase activity. The soil enzyme activities play a very important role in nutrients recycling,  
 375 making them accessible to plants and microorganisms (Ai *et al.*, 2015).

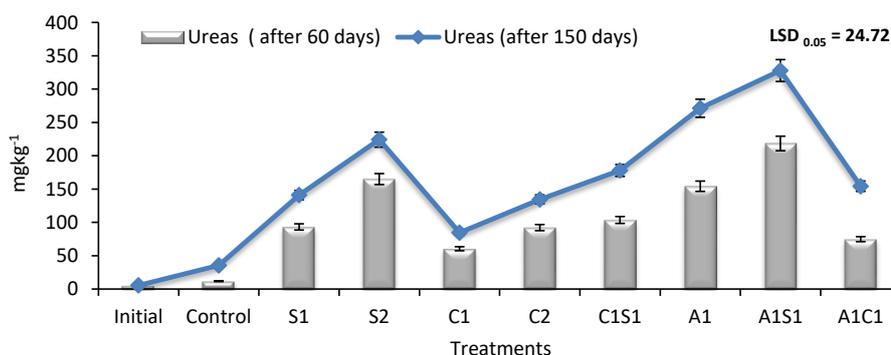


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Figure (5): Effect of *Azospirillum* and eco-friendly organic wastes on Dehydrogenase enzyme ( $\mu\text{g TPF ml}^{-1}$ ) in saline-sodic soil after 60 and 150 days of incubation. Data points and error bars represent means and standard errors ( $n = 4$ ).



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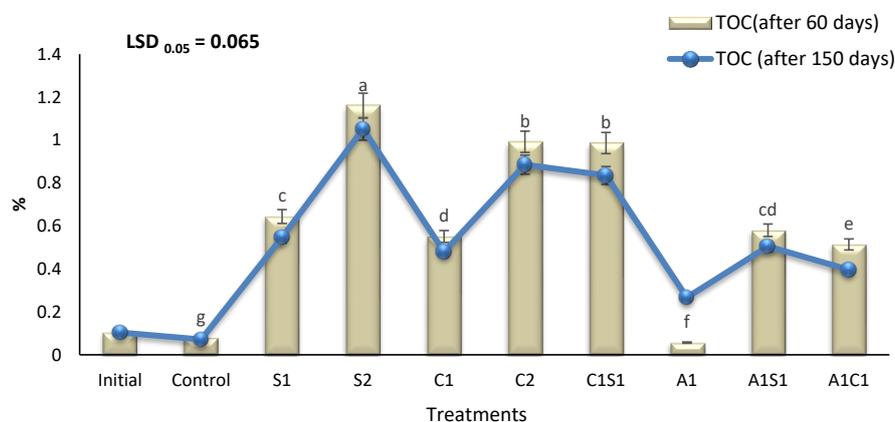
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Figure (6): Effect of *Azospirillum* and eco-friendly organic wastes on Urease enzyme in saline-sodic soil after 60 and 150 days of incubation. Data points and error bars represent means and standard errors ( $n = 4$ ).



### 384 3.2.2. Soil organic carbon

385 The soil organic carbon (SOC) values are shown in Table (4) and Figure (7). The SOC  
386 percentages after 60 days of incubation were 0.1, 0.081, 0.64, 1.16, 0.55, 1.0, 0.98, 0.056, 0.58 and  
387 0.51 for initial condition, control, S1, S2, C1, C2, C1S1, A1, A1S1 and A1C1, respectively. There  
388 were significant differences among the treatments. The trend of SOC after 60 days was similar to the  
389 trend of SOC after 150 days of incubation, where the SOC percentages were 0.1, 0.0712, 0.54, 1.05,  
390 0.47, 0.88, 0.83, 0.26, 0.50 and 0.39 for initial condition, control, S1, S2, C1, C2, C1S1, A1, A1S1  
391 and A1C1, respectively. The spent grain (S2) possessed the highest values of SOC while the A1 and  
392 control showed the lowest value. The spent grain amendments increased the carbon contents in  
393 saline-sodic soil up to 12 gkg<sup>-1</sup>. The use of organic wastes with *Azospirillum* increased saline-sodic  
394 soil organic carbon contents. On the other hand, the *Azospirillum* bacteria increased the SOC after  
395 150 days greater than after 60 days, this result *Azospirillum* increased of decomposition rate was  
396 highly significant after 150 days compared with 60 days of incubation. The large increase of SOC  
397 and TN amounts with organic amendments application might be attributed to the increased  
398 particulate organic matter derived from manure since the organic material added to the soil was  
399 initially located in the coarser fraction (Liang *et al.*, 2014; Zhang *et al.*, 2015). The high organic  
400 matter content enhanced the production of soil organic carbon as a result of micro-organisms activity  
401 during the mineralization process of organic matter. Generally, the mineralized organic carbon  
402 increased as the application rate of spent grain and compost increased. The presented results of SOC  
403 are in agreement with the results presented by (Luo *et al.*, 2018; Nadeem *et al.*, 2014; Nassar *et al.*,  
404 2014; Scherer *et al.*, 2011; Zhang *et al.*, 2015)



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Figure (7): Effect of *Azospirillum* and eco-friendly organic wastes on SOC in saline-sodic soil after 60 and 150 days of incubation. Data correspond to means of four replicates  $\pm$  standard error. Different letters indicate statistically significant differences ( $p \leq 0.05$ ) among treatments.



410 Table (4): Soil N, P, K and SOC in saline-sodic soil after 60 and 150 days of incubation

Treatments	Incubation Period Day	Total N	Ava P	Ava K	SOC
		%	mgkg <sup>-1</sup>		%
<b>Initial</b>	<b>Zero time</b>	<b>0.0021</b>	<b>1.2</b>	<b>78.0</b>	<b>0.1</b>
Control	60	0.0030	1.95	86.31	0.096
	<b>150</b>	<b>0.0047</b>	<b>2.06</b>	<b>92.66</b>	<b>0.071</b>
S1	60	0.0280	3.73	158.47	0.597
	<b>150</b>	<b>0.0950</b>	<b>5.72</b>	<b>179.33</b>	<b>0.547</b>
S2	60	0.0633	6.81	268.06	1.175
	<b>150</b>	<b>0.1482</b>	<b>14.28</b>	<b>284.33</b>	<b>1.052</b>
C1	60	0.0160	2.66	121.71	0.576
	<b>150</b>	<b>0.0840</b>	<b>3.72</b>	<b>145.20</b>	<b>0.471</b>
C2	60	0.0360	5.45	206.24	0.968
	<b>150</b>	<b>0.1090</b>	<b>8.45</b>	<b>237.66</b>	<b>0.885</b>
C1S1	60	0.0567	7.22	200.00	1.169
	<b>150</b>	<b>0.1250</b>	<b>10.00</b>	<b>237.66</b>	<b>0.835</b>
A1	60	0.0460	8.94	88.366	0.055
	<b>150</b>	<b>0.1437</b>	<b>13.13</b>	<b>117.66</b>	<b>0.267</b>
A1S1	60	0.0583	8.81	166.14	0.576
	<b>150</b>	<b>0.1717</b>	<b>9.63</b>	<b>195.00</b>	<b>0.504</b>
A1C1	60	0.0324	5.22	139.20	0.501
	<b>150</b>	<b>0.099</b>	<b>7.15</b>	<b>180.50</b>	<b>0.396</b>
<b>LSD<sub>0.05</sub></b>	60	0.005	0.452	11.0	0.208
	150	0.025	0.761	17.74	0.065

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413 **3.3. Soil fertility indicator after incubation periods**

414 **3.3.1. N, P, and K<sup>+</sup> concentration in saline-sodic soil**

415 The effects of different amendments on soil N are shown in Table (4). *Azospirillum*  
 416 inoculation not only alleviated the inhibitory effect of salinity on soluble and total macro-nutrients  
 417 but also induced a marked and progressive increase in N, P and K concentrations. Figure (8) show  
 418 total N concentrations. Total nitrogen is the sum of nitrate (NO<sub>3</sub><sup>-</sup>), nitrite (NO<sub>2</sub>), organic nitrogen and  
 419 ammonia (all expressed as N). The total N percentages for the saline-sodic soil after 60 days of  
 420 incubation were 0.0021, 0.003, 0.028, 0.063, 0.016, 0.036, 0.056, 0.046, 0.058 and 0.032% for the  
 421 initial condition, control, S1, S2, C1, C2, S1C1, A1, A1S1 and A1C1 treatments, respectively. The  
 422 corresponding percentages for the saline-sodic soil after 150 days of incubation were 0.0021, 0.0047,  
 423 0.095, 0.148, 0.084, 0.109, 0.125, 0.143, 0.171 and 0.099%. The spent grain with *Azospirillum*  
 424 inoculation was a good supplier for the total nitrogen at both incubation periods in comparison to the  
 425 initial total N.

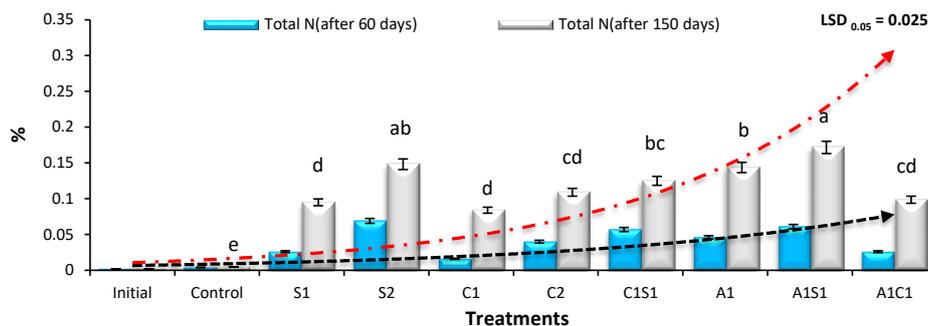
426 The compost treatment supplied a lower total N in comparison to the *Azospirillum* and spent  
 427 grain. This is due to the high initial nitrogen percentage in spent grain and enhancements  
 428 decomposition rate and mineralization process by *Azospirillum* bacteria. The total N was greater  
 429 after 150 days of incubation than after 60 days of incubation in saline-sodic soil. Is in accordance  
 430 with the report of (Liang *et al.*, 2014; Sastre-Conde *et al.*, 2015; Yu *et al.*, 2019; Zhang *et al.*, 2015),



431 that a majority part of organic fertilizer is released to the soil N, P and K<sup>+</sup> after application of  
432 agricultural waste compost and bio-fertilizer in saline-sodic soils. Accordingly, the spent grain with  
433 *Azospirillum* used in the present study can supply a growing plant with plentiful of nitrogen after 60  
434 and 150 days of incubation and this period was enough for decomposition organic matter and  
435 mineralization of nutrients. Soil microbial communities regulate principal ecosystem processes such  
436 as soil nutrient cycling and organic matter formation and decomposition (Sant'Anna *et al.*, 2018;  
437 Trivedi *et al.*, 2017).

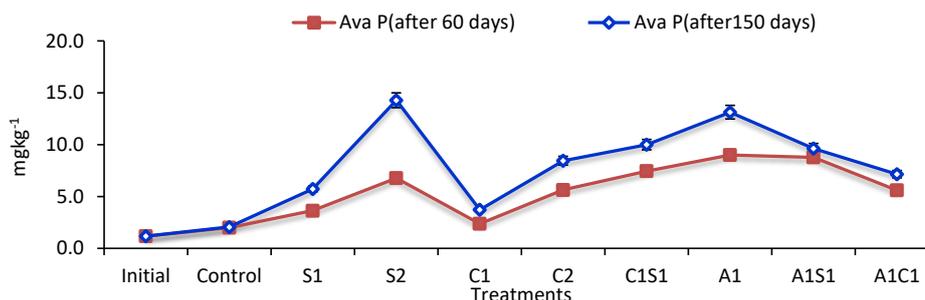
438 **The soil available P and K<sup>+</sup>** values are shown in Table (4) and Fig (9) were significantly  
439 increased by eco-friendly organic wastes with *Azospirillum* treatments compared to the initial  
440 condition. The trend of available P and K after 60 days was similar to the trend of available P and K  
441 after 150 days of incubation, the available P and K, concentrations followed the order S2 > A1 >  
442 A1S1 > C1S1 > C2> S1 > C1 > A1C1 > control > initial respectively. The spent grain (S2) possessed  
443 the highest values of P and K while the A1C1 and control showed the lowest value. The available P  
444 and K were significantly ( $P < 0.05$ ) increased by spent grain (S2) treatment compared to control  
445 treatment after 150 days 12 and 3.64 folds increase in P and K respectively. The spent grain  
446 amendments increased the available P and K contents in saline-sodic soil up to 12 mgkg<sup>-1</sup> in S2 and  
447 A1 respectively.

448 The spent grain application rates with *Azospirillum* were superior compared to the same rate  
449 of compost for soil P and K. The high rate of spent grain (S2) significantly produced larger increases  
450 in soil P than in the other organic wastes amendment, while A1 was statistically similar to S2 and  
451 both significantly increased soil P compared to other organic wastes amendments shown Figure (9).  
452 Generally soil N, P and K increased as the application rate of *Azospirillum* or spent grain increased,  
453 and all of the organic treatments were effective at adding N, P and K to saline-sodic soil.  
454 Furthermore, spent grain and *Azospirillum* should be applied before planting to give sufficient time  
455 for natural oxidation of organic wastes materials, which in turn enhances the soil available nutrients.  
456 Similar to the results of the present study, incubation of some organic sources resulted high total N.  
457 Our results also concur with those of (Barra *et al.*, 2019; Ouyang *et al.*, 2013; Trivedi *et al.*, 2017)  
458 who found that the application of composted poultry manure, organic fertilizer, and bio-fertilizers  
459 inoculation enhances the benefit of P-fertilization led to the increase of available P in saline-sodic  
460 soils. This increase in available P is due to the reduction in pH as a result of organic matter  
461 decomposition. Furthermore, the application of bio-fertilizer or FYM significantly increased total N,  
462 available P and K for saline soils. (Liang *et al.*, 2014; Yu *et al.*, 2019).



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Figure (8): Effect of *Azospirillum* and eco-friendly organic wastes on total (N) in saline-sodic soil after 60 and 150 days of incubation. Data correspond to means of four replicates  $\pm$  standard error. Different letters indicate statistically significant differences ( $p \leq 0.05$ ) among treatments.



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Figure (9): Effect of *Azospirillum* and eco-friendly organic wastes on available (P) in saline-sodic soil after 60 and 150 days of incubation. Data points and error bars represent means and standard errors ( $n = 4$ )

Table (5): Soil micronutrients in saline-sodic soil after 60 and 150 days of incubation

Treatments	Incubation Period	Fe <sup>2+</sup>	Zn <sup>2+</sup>	Mn <sup>2+</sup>	Cu <sup>2+</sup>	B <sup>+</sup>
Initial	Zero time	<b>0.08±0.01</b>	Nd*	<b>0.11±0.02</b>	<b>0.003±0.01</b>	Nd*
Control	60	0.74±0.01	0.332±0.06	0.619±0.07	0.093±0.01	0.058±0.01
	150	<b>1.27±0.08</b>	<b>0.477±0.04</b>	<b>1.089±0.10</b>	<b>0.264±0.02</b>	<b>0.081±0.009</b>
S1	60	3.33±0.19	0.925±0.03	2.049±0.09	0.742±0.08	0.445±0.021
	150	<b>4.27±0.32</b>	<b>1.335±0.13</b>	<b>2.625±0.05</b>	<b>1.020±0.13</b>	<b>0.768±0.151</b>
S2	60	8.01±0.41	2.633±0.10	5.989±0.66	1.344±0.11	0.716±0.012
	150	<b>9.15±0.44</b>	<b>3.273±0.13</b>	<b>7.461±0.72</b>	<b>1.776±0.02</b>	<b>1.110±0.18</b>
C1	60	2.62±0.05	0.753±0.03	1.657±0.13	0.403±0.02	0.378±0.04
	150	<b>3.12±0.06</b>	<b>1.059±0.14</b>	<b>2.334±0.22</b>	<b>0.501±0.08</b>	<b>0.510±0.02</b>
C2	60	4.40±0.14	1.439±0.11	3.522±0.05	0.851±0.02	0.574±0.02
	150	<b>5.16±0.023</b>	<b>2.178±0.07</b>	<b>4.719±0.19</b>	<b>1.131±0.05</b>	<b>0.858±0.05</b>
C1S1	60	5.96±0.55	1.473±0.07	4.602±1.64	0.868±0.30	0.591±0.18
	150	<b>6.64±0.65</b>	<b>2.010±0.18</b>	<b>6.159±0.27</b>	<b>1.329±0.10</b>	<b>1.143±0.14</b>
A1	60	4.03±0.03	1.113±0.13	2.595±0.07	0.711±0.11	0.350±0.02
	150	<b>4.68±0.21</b>	<b>1.488±0.12</b>	<b>3.183±0.18</b>	<b>0.993±0.16</b>	<b>0.501±0.07</b>
A1S1	60	12.58±0.40	2.904±0.04	5.956±0.11	1.531±0.17	0.916±0.04
	150	<b>13.80±0.36</b>	<b>4.125±0.39</b>	<b>7.600±0.37</b>	<b>1.945±0.24</b>	<b>1.275±0.36</b>
A1C1	60	2.38±0.11	0.865±0.09	1.789±0.11	0.467±0.05	0.439±0.05
	150	<b>3.81±0.28</b>	<b>1.308±0.35</b>	<b>2.193±0.16</b>	<b>0.660±0.04</b>	<b>0.534±0.08</b>
LSD <sub>0.05</sub>	60	0.484	0.140	1.022	0.228	0.114
	150	0.585	0.361	0.541	0.207	0.271

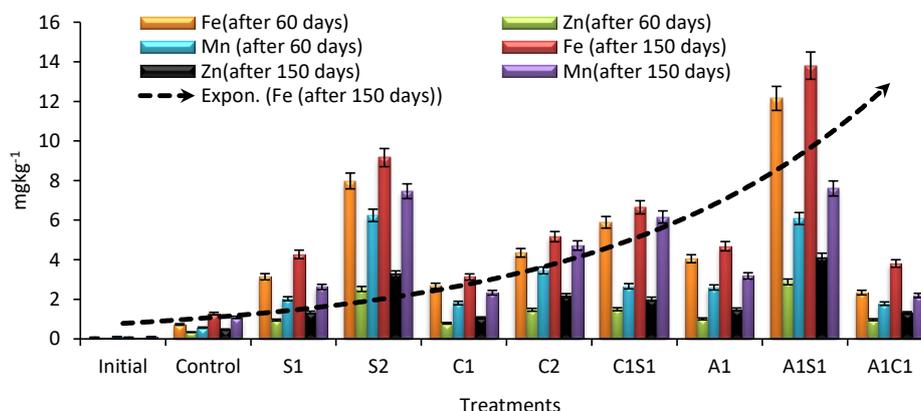
472 Data correspond to means of four replicates  $\pm$  standard deviation. LSD<sub>0.05</sub>= least significant difference (LSD) test  
 473 at ( $p \leq 0.05$ ). Nd\*= not detected  
 474



475 **3.3.2. Soil micro-nutrients ( $Fe^{2+}$ ,  $Zn^{2+}$ ,  $Mn^{2+}$ ,  $Cu^{2+}$  and  $B^+$ )**

476 The soil amendments were found significantly impact on micro-nutrients. Table (5) shows  
477 that  $Fe^{2+}$ ,  $Zn^{2+}$ ,  $Mn^{2+}$ ,  $Cu^{2+}$ , and  $B^+$  concentrations after 60 days of incubation had similar trend  $Fe^{2+}$ ,  
478  $Zn^{2+}$ ,  $Mn^{2+}$ ,  $Cu^{2+}$  and  $B^+$  concentrations after 150 days of incubation.  $Fe^{2+}$ ,  $Zn^{2+}$ ,  $Mn^{2+}$ ,  $Cu^{2+}$  and  $B^+$   
479 concentrations followed the order A1S1 > S2 > C1S1 > C2 > A1 > S1 > A1C1 > C1 > control > initial.  
480 The micronutrients, the largest increases were observed with A1S1 treatment while the lowest  
481 increases were observed with C1 and control treatment after 60 and 150 days of saline-sodic soil  
482 incubated. Available micronutrients were significantly ( $P < 0.05$ ) increased by *Azospirillum* with  
483 spent grain (A1S1) treatment compared to control treatment 18.9, 12.42, 12.66, 20.86 and 21.90 folds  
484 increase in  $Fe^{2+}$ ,  $Zn^{2+}$ ,  $Mn^{2+}$ ,  $Cu^{2+}$ , and  $B^+$ , respectively. The *Azospirillum* and spent grain application  
485 rates (A1S1 and S2) were superior compared to the same rate of compost for  $Fe^{2+}$ ,  $Zn^{2+}$ , and  $Mn^{2+}$   
486 (Fig 10). The superiority of the *Azospirillum* with spent grain was due to its effect on lowering soil  
487 pH as well as solubility and chelation effect of organic wastes in saline-sodic soil (Pii *et al.*, 2015a).  
488 Accordingly, the application of A1S1 to enrich soil with micronutrients is highly recommended in  
489 saline-sodic soils. The mineralization of eco-friendly organic wastes compounds leads to decrease pH  
490 in saline-sodic soil, which increases micronutrient availability in the soil. The presented results are in  
491 agreement with the results presented by (Rashad *et al.*, 2016; Wang and Cai, 2006; Zhang *et al.*,  
492 2019) who reported that the available micro-nutrients are increased by using agro-industrial organic  
493 wastes and microbial community.

494



495 Figure (10): Effect of *Azospirillum* and eco-friendly organic wastes on micro-nutrients ( $Fe^{2+}$ ,  $Zn^{2+}$ , and  
496  $Mn^{2+}$ ) in saline-sodic soil after 60 and 150 days of incubation. Data points and error bars represent means  
497 and standard errors ( $n = 4$ ).  
498  
499

500  
501



502 **CONCLUSION**

503 The results of the present study confirm our hypotheses that (i) Application of spent grain  
504 doses with *Azospirillum* inoculation reduced the ESP, increased the beneficial microbes and the  
505 activity of enzymes from decomposition of the eco-friendly organic wastes and (ii) Soil fertility  
506 increased by improvement of macro and micro-nutrients in saline-sodic soil after 60 and 150 days of  
507 incubation. The results also showed that the spent grain and *Azospirillum* treatments enhanced the  
508 production of soil organic carbon, the activity of enzymes, soil macro and micro-nutrients which  
509 decreased soil salinity in saline-sodic soil as a result of micro-organisms activity during the  
510 mineralization process of spent grain with *Azospirillum* treatments.

511 In the saline-sodic soil, the concentration of monovalent total N, available P, and available  
512  $K^+$  also  $Fe^{2+}$ ,  $Zn^{2+}$ ,  $Mn^{2+}$ ,  $Cu^{2+}$ , and  $B^+$  were affected significantly by the application of *Azospirillum*  
513 with spent grain under controlled laboratory conditions. In the final application of *Azospirillum* in  
514 saline-sodic soil has a negligible effect on the final SOC content but it has a strong effect on soil  
515 dehydrogenase, urease, pH, EC, ESP, and  $Ex-Na^+$ . At the end of experiments, monovalent cations  
516 concentrations, as well as soil EC and ESP, were smaller after 150 days than 60 days of incubation.  
517 Finally, the soil salinity indicator decreased significantly, even for eco-friendly organic wastes with  
518 *Azospirillum*, after 150 days of incubation. Results suggested that the applied spent grain with  
519 *Azospirillum*, depending on their biological and chemical properties might have a positive effect on  
520 soil properties, especially in the absence of compost.

521

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