



1 **Identification of new microbial functional standards for soil quality assessment**

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20 **Abstract.** The activity of microorganisms in soil is important for a robust functioning soil and related
21 ecosystem service. Hence, there is a necessity to identify the indigenous soil microbial community for its
22 functional properties using soil microbiological methods in order to determine the natural properties,
23 functioning and operating range of soil microbial communities, and to assess ecotoxicological effects due
24 to anthropogenic activities. Numerous microbiological methods currently exist in the literature and new,
25 more advanced methods continue to be developed; however, only a limited number of the methods are
26 standardized. Consequently, there is a need to identify the most promising non-standardized methods for
27 assessing soil quality and develop these into standards. In alignment with the “Ecosystem Service
28 Approach”, new methods should focus on soil microbial function, including nutrient cycling, pest control
29 and plant growth promotion, carbon cycling and sequestration, greenhouse gas emission, and soil
30 structure. The few existing, function-related standard methods available focus on the estimation of
31 microbial biomass, basal respiration, enzyme activities related to nutrient cycling, and organic chemical
32 biodegradation. This paper sets out to summarize and expand on recent discussions within the
33 International Organization for Standardization (ISO), Soil Quality - Biological Characterization sub-
34 committee (ISO TC 190/SC 4) where a need was identified to develop scientifically sound methods which
35 would best fulfil the practical needs of future users for assessing soil quality. Of particular note was the
36 current evolution of molecular methods in microbial ecology that uses quantitative real time PCR (qPCR)
37 to produce a large number of new endpoints and is more sensitive as compared to ‘classical’ methods.
38 Quantitative PCR assesses the activity of microbial genes that code for enzymes that catalyse major
39 transformation steps in nitrogen and phosphorus cycling, greenhouse gas emissions, chemical
40 transformations including pesticide degradation, and plant growth promotion pathways. In the
41 assessment of soil quality methods, it was found that fungal methods were significantly underrepresented.
42 As such, techniques to analyse fungal enzyme activities are proposed. Additionally, methods for the
43 determination of microbial growth rates and efficiencies, including the use of glomalin as a biochemical



44 marker for soil aggregation, are discussed. Furthermore, field methods indicative of carbon turnover,
45 including the litter bag test and a modification to the tea bag test, are presented. As a final note, it is
46 suggested that endpoints should represent a potential function of soil microorganisms rather than actual
47 activity levels, as the latter can largely be dependent on short-term variable soil properties such as
48 pedoclimatic conditions, nutrient availability, and anthropogenic soil cultivation activities.



49 **1 Introduction**

50 Soils are one of the world's hotspots for biodiversity (Parker, 2010). Biota – both micro- and macro-
51 organisms - in soil form strong networks and complex food webs, which determines the efficacy of the soil
52 ecosystem functions (e.g. nutrient cycling, C storage and turnover, water retention, and modulation of soil
53 structure) (Creamer et al., 2016). These functions support a range of ecosystem services that are
54 indispensable for soil use in agri-, horti- or silviculture (Nannipieri et al., 2017). At the same time, soil biota
55 are strongly impacted by various anthropogenic activities including ongoing global and climate change,
56 pollution, and degradation and destruction of the terrestrial environment (Gomiero, 2016; Montgomery,
57 2008; Wagg et al., 2014). Consequently, investigations of the soil biome structure and function became an
58 emerging topic in soil and environmental sciences (Griffiths and Philippot, 2013). As such, the number of
59 publications on soil ecology and ecosystem functioning has increased significantly over the past few
60 decades and has resulted in the development of new methods (e.g. Guillaume et al., 2016; Tian et al.,
61 2018). In comparison, the ecotoxicological assessment of human impacts (e.g. chemical pollution and
62 mechanical compaction) using single species tests, which are well-established methods, has remained
63 constant (Brookes, 1995; Joergensen and Emmerling, 2006).

64 Characterizing the natural state of a soil's biome is quite a challenging task. In addition to its huge
65 structural and functional diversity, the soil biome is influenced by strong temporal dynamics including
66 seasonal weather conditions and the enormous spatial heterogeneity which ranges from field scale to
67 microscale (Kuffner et al., 2012; Regan et al., 2014; Suriyavirun et al., 2019). All of these intrinsic properties
68 hinder the interpretation of data obtained from the analysis of soil biomes and the measurement of their
69 functional traits.

70 Despite the fundamental methodological advances over the past years, which allow for an in-
71 depth analysis of microbiomes and, to some extent, other soil-living organisms (e.g. Joergensen and
72 Emmerling, 2006; Paul, 2015; Yates et al., 2016), only a limited number of soil biological methods have



73 been standardized (for details see section 3). As a result, large deviations are observed between non-
74 standardized method protocols (e.g. Strickland and Rousk, 2010). Therefore, comparability between
75 datasets generated by different laboratories using different methods or modified protocols of the same
76 method are problematic. Consequently, the development of quality indices and threshold values,
77 respectively, for assessing soil quality is nearly impossible (Bastida et al., 2008). Presumably, this is why
78 the number of meta-analyses in soil biology remains small.

79 Given that there is a lack of harmonization between existing methods and, at the same time, a
80 proliferation of new methods, there is a need to identify the most promising methods described in the
81 literature that can be standardized to produce reliable indicators for soil quality (e.g. Philippot et al., 2012).
82 At the Annual International Organization for Standardization (ISO) meeting of TC 190 (Soil Quality) in
83 Fukuoka, Japan in October 2013, a decision was made to compile a list of available methods and to identify
84 those that would be suitable for assessing soil quality. Additionally, during a subsequent meeting of ISO
85 TC 190/SC 4/WG 4 (Microbiological Methods) held in Paris, France in March 2014, further discussions
86 focused the criteria for suitable methods to be comprised of microbial functional indicators. In this paper,
87 we summarize the major outcomes of the discussions which took place over the past several years within
88 ISO TC 190/SC 4. Therefore, besides collating a list of criteria for the selection of test methods for the
89 future analysis of microbial functions in soil, the aim of this paper is to present our opinion, as members
90 of the ISO TC 190 committee, to initiate further discussion on possible methods that should be
91 standardized for future soil quality assessments.

92

93 **2 Criteria for the selection of methods**

94 Several papers addressing the task to identify suitable methods to be used as biotic indicators (usually
95 including faunal indicators) were published in the last few years, mainly in the context of EU research
96 projects (e.g. Bispo et al., 2009; Faber et al., 2013; Ritz et al., 2009; Römbke et al., 2010). Here, we propose



97 to base the selection of soil quality methods more on the “Ecosystem Service Approach” (MEA 2005) which
98 is increasingly recognized by both environmental scientists and regulatory agencies (Breure et al., 2012;
99 Galic et al., 2012) and that soils have been raised to the rank of a natural resource to be protected. As a
100 consequence, and in addition to method development and application (including the assessment of
101 biodiversity as a prerequisite for soil function), the focus of future activities should be the determination
102 of soil microbial function as recommended endpoints (Kvas et al., 2017; Nienstedt et al., 2012; van der
103 Putten et al., 2010; TEEB, 2010). Consequently, we propose to assess both existing and new methods for
104 the selection of microbial functional tests that support various soil ecosystem services. This structures our
105 approach and simplifies the identification of ecologically relevant methods, as well as, presumably
106 increasing their acceptance by users, including the regulatory and stakeholder community. The following
107 soil **Functions** and ecosystem services have been defined and are proposed to be used as a starting point
108 for the development of future methods (MEA, 2005; Ockleford et al., 2017):

- 109 (1) Biodiversity, genetic resources, cultural services;
- 110 (2) Food web support;
- 111 (3) Biodegradation of pollutants;
- 112 (4) Nutrient cycling (for example N and P);
- 113 (5) Pest control and plant growth promotion;
- 114 (6) Carbon cycling and sequestration;
- 115 (7) Greenhouse gas emissions; and
- 116 (8) Soil structure affecting soil water, gas balance and filtration function.

117 A second major criterion for selecting methods for standardization is its usability. The method should be
118 applicable in regulations (e.g. European and National agencies registering chemicals or products) and for
119 the evaluation of soil ecology and functioning as a fundamental aspects of soil quality (e.g. by stakeholders
120 and researchers). Moreover, the routine use of methods to inform farmers and site owners on soil quality



121 as continuous assessments of their land and land-use practises, could be an additional condition that
122 would require the choice of easy-to-use methods or possibly encourage the simplification of existing
123 methods. Frequently used methods generate more data, which in turn is of high importance for the
124 validation of threshold values.

125 To assess possible methods, a list of criteria was used based on the ‘logical sieve’ approach (Ritz
126 et al., 2009). The list of criteria for the identification of functional indicators and associated methodologies
127 (Table 1) was an outcome of the EU FP7 EcoFINDERS project (Faber et al., 2013). The criteria were compiled
128 after sending a questionnaire to 25 partner institutions primarily working in the field of environmental
129 science; mainly representing academia but also regulators and subcontracting laboratories. These criteria
130 are applicable for different kinds of indicators and methods, including those addressing the functions of
131 soil microbial communities. In the following sections, we assume that existing ISO standardized methods
132 already fulfil these criteria. Additionally, the most appropriate new methods, including those proposed in
133 this article, need to be evaluated using the same criteria required for the standardization of ISO methods.
134 Therefore, the aim of this whole process is to identify methods which are scientifically sound and that best
135 fulfil the practical needs of future users.

136

137 **3 Existing and new methods**

138 Current methods that have already been implemented as ISO standards are found in Table 2, whereas
139 methods that might be considered for future standardization are in Table 3. The compilation in Table 2
140 comprises methods to quantify microbial biomass (e.g. through fumigation extraction of microbial biomass
141 carbon (MBC) and DNA) (**Function 6**) as well as for (further) analysis of structural microbial diversity (e.g.
142 determination of microbial fingerprints by phospholipid fatty acids (PLFA) analysis) (**Function 1**).
143 Additionally, microbial biomass, measured as respiratory activity, has been included in Table 2, although
144 not directly linked to one of the ecosystem services, as it provides important information on the activity



145 of the complete microbiome (i.e. microflora and microfauna). Soil basal respiration normalized to MBC
146 (ISO 14240-1 and ISO 14240-2, 1997, Table 2) yields the metabolic quotient qCO_2 , which is a sensitive
147 indicator for microbial carbon use efficiency (Anderson and Domsch, 1993). Interestingly, its use as an
148 endpoint to assess anthropogenic and natural impacts on the soil microbiome has been controversially
149 discussed in literature (Wardle and Ghani, 1995).

150 The biodiversity function (**Function 1**) addresses parameters related to the structural diversity of
151 the soil microbiome. Here, respective ISO guidelines analysing PLFA, phospholipid ether lipids (PLEL)
152 (ISO/TS 29843-1, 2010; ISO/TS 29843-2, 2011) and DNA (ISO 11063, 2012; ISO 17601, 2016), have already
153 been well implemented into guidelines (Table 2). Although microbial diversity, per se, is not strongly
154 correlated with a particular functional capacity, it is clear that the loss of diversity can have an impact on
155 microbial function (Thiele-Bruhn et al., 2012); at least for relatively specific functions performed by narrow
156 microbial guilds or taxa. This applies even more, when certain taxa are closely linked to very specific
157 functions including nitrifiers, methanogens, arbuscular- and ecto-mycorrhizal fungi, and biocontrol
158 microorganisms like *Trichoderma* (e.g. Hartmann et al., 2009; Hayat et al., 2010; Lugtenberg and Kamilova,
159 2009; Peng et al., 2008; Singh et al., 2007; Xia et al., 2011). Therefore, the interpretation of the outcomes
160 from microbial community-based testing tends to be straightforward and closely linked to **Function 4** and
161 **Function 5**.

162 Food web support (**Function 2**) of higher trophic levels no doubt starts from soil microorganisms
163 and propagates through the trophic levels (e.g. earthworms) that are consumed by birds and mammals
164 (Haynes, 2014; Scheu et al., 2002; Scheu et al., 2005). However, the role of the microbiota in the soil food
165 web is not fully understood, since many eukaryotic organisms can be considered as meta-organisms, which
166 carry their “own microbiome” that itself is essential for life supporting functions. From this, it is unclear if
167 environmental microbiomes and host specific microbiomes complement one another. So far there are no
168 comprehensive methods (especially not those addressing microbial functions) or standards available to



169 assess this problem. A future considering may employ the use of stable isotope labelling of select carbon
170 sources as a promising approach to follow food webs and degradation pathways (e.g. Coban et al., 2015;
171 Traugott et al., 2013).

172 Methods to assess the biodegradation of pollutants (**Function 3**), as described above, are already
173 implemented into ISO guidelines (Table 2) and are part of legal frameworks including pesticide directives
174 (EU Regulation 1107/2009/EC; European Commission, 2009). A number of standard methods for the
175 determination of the potential of soils to degrade organic chemicals (**Function 3**) under both aerobic (ISO
176 14239, 2017) and anaerobic (ISO 15473, 2002) conditions are available. This emphasizes that in the past,
177 the development of standard methods was mainly driven by the need to assess the ecotoxicological effects
178 of anthropogenic activities, such as chemical contamination of soils, rather than to describe and
179 understand the natural properties and functions of soils. However, defining methods for the
180 determination of adverse effects of contaminants on soil biota was not only being done in ISO, but it was
181 also a major task of other organizations such as the Organization for Economic Co-Operation and
182 Development OECD). For example, there are OECD guidelines, tests No. 216 and 217, for testing the long-
183 term effects of single exposure chemicals on soil microbial nitrogen and carbon transformation,
184 respectively (OECD, 2000a; 2000b). As a result, it was decided early that the standardization of methods
185 for toxicity testing would not be the primary aim of the ISO sub-committee (ISO TC 190/SC 4).

186 Some of the existing standard methods that are listed in Table 2 focus on the estimation of enzyme
187 activities useful for soil quality assessment, which mainly contribute to **Function 4**. Here, the potential
188 dehydrogenase activity measurement is an indicator for general (potential) oxidoreductase activity in soil.
189 Since this measurement has been frequently used, there are large amounts of baseline data available on
190 the toxic effects of a range of pollutants in soil. Recently, additional potential enzyme activities related to
191 the C, N, P and S cycle have been used and are either standardized or are in the process for standardization.

192



193 The current evolution of molecular methods in microbial ecology has resulted in a large number
194 of new endpoints. It is well known that many of the new endpoints (e.g. using quantitative real-time PCR
195 (qPCR)) are more sensitive than classical methods that had been standardized in the past (Ribbons et al.,
196 2016; Schulz et al., 2016). This new metagenomics approach will be of high importance in the future, as it
197 allows for the implementation of information on new functional traits that can be standardized into an
198 analytical pipeline. For the assessment of new methods linked to **Functions 4 to 8**, qPCR from soil DNA
199 extracts (ISO 17601, 2016) plays a very important role in determining the abundance of single marker gene
200 sequences, which are indicative of specific transformation processes or soil functions. For example, the
201 quantification of nitrogen fixing microbes, nitrifiers and denitrifiers has been successfully implemented
202 using the *nifH*, *amoA* and *nirS/nirK* genes as markers, respectively (Henry et al., 2004; Hirsch et al., 2010;
203 Ollivier et al., 2010; Sessitsch et al., 2006). Similarly, the quantification of microorganisms involved in the
204 β -keto adipate pathway has been implemented by targeting *pcaH* (El Azhari et al., 2008) and *catA* (El Azhari
205 et al., 2010) gene sequences. Various methods for the assessment of soil microbial **Function 4** (nutrient
206 cycling), **Function 5** (pest control and plant growth promotion) and **Function 7** (greenhouse gas emissions)
207 are proposed based on the qPCR analysis of gene sequences coding for enzymes which trigger the
208 respective function (e.g. Fish et al., 2013; Ribbons et al., 2016; Smith and Osborn, 2009). Additionally, it
209 should be noted that molecular methods based on the assessment of specific marker genes for estimating
210 the degradation potential in soil have already been proposed both for PAHs (e.g. Cebron et al., 2008) and
211 individual pesticides (e.g. Martin-Laurent et al., 2004). These could be interesting for future
212 standardization; however, if a method is very compound-specific and targeted, this could limit its
213 application range. Thus, these specific approaches will not be discussed further in this article.

214 Major advantages of qPCR assays to quantify gene sequence numbers, which can be used as
215 proxies for a given microbial process, are that they are: (i) highly standardized, sensitive, selective and
216 reproducible, (ii) designed for high throughput analysis, (iii) available for a wide range of targets, and (iv)



217 methods that are relatively cheap once the necessary analytical devices are on hand. Some training on the
218 method is required, however, once trained the assays are easy to perform. For example, numerous studies
219 have already used the microbial functional genes involved in nitrogen cycling to determine the status and
220 to assess induced changes in the soil microbial community (Levy-Booth et al., 2014; Nannipieri and Eldor,
221 2009; Wallenstein et al., 2006). Consequently, the number of functional genes that are suited for use as
222 specific indicators of soil function are continuing to grow in the literature as researcher gain experience in
223 this field and data becomes more prevalent.

224 Disadvantages, on the other hand, are that: (i) the quality of qPCR data depends on soil DNA
225 extracts (PCR inhibition), (ii) primer pairs, even degenerated ones, might not successfully amplify all
226 microbes of the functional group of interest, (iii) only genetic potential is resolved, and (iv) there is no
227 differentiation between active, dormant or dead microorganisms, when working with DNA as a template
228 for the qPCR reaction. The analysis of total RNA and of mRNA, which could help to overcome the latter
229 problem, is currently not a suitable alternative as it is highly dynamic in time and space and needs special
230 care to stabilize the RNA extracted from complex environmental matrices to avoid its degradation.
231 Another problem of DNA analysis is the biological representativeness of the results is solely based on a
232 relatively small amount of soil (from few hundred mg to ten g of soil) from which the DNA extracted. The
233 use of small soil samples (< 1 g) simplifies the sample preparation process for molecular biologists;
234 however, it provides a poor representation of the indigenous soil microbial community in the naturally
235 inhomogeneous soil. Typically, the α -biodiversity declines with sample size while that of β -biodiversity
236 increases (Nicol et al., 2003; Penton et al., 2016). Lastly, it must be noted that the high repeatability and
237 reproducibility of molecular biology methods, including qPCR assays, depends on extraction, purification
238 and amplification of DNA or RNA. This is typically performed using commercial extraction kits; however,
239 by simply changing the commercial supplier of a kit can substantially change the results (Brooks et al.,



240 2015; Feinstein et al., 2009). This clearly challenges standardization since standard methods must not
241 hinge on a specific supplier.

242 Recently, successful examples of microbial phosphorous turnover (**Function 4**) have been
243 published (Bergkemper et al., 2016) where metagenomics data have been used for the construction of
244 primers for P mineralization, transport and uptake. As another example, the relevance of anaerobic
245 ammonium oxidation (anammox) for N cycling in soils has increased (Levy-Booth et al., 2014) along with
246 the development of analytical methods for high throughput analysis. Among the microorganisms in soil
247 that substantially govern pest control and plant growth promotion (**Function 5**), the most numerous
248 organisms are the arbuscular mycorrhizae and ectomycorrhizal fungi. These microorganisms are especially
249 abundant in the rhizosphere (Hartmann et al., 2009; Hayat et al., 2010; Lugtenberg and Kamilova, 2009).
250 Methods related to **Function 5** are listed in Table 3.

251 Several options exist for (additional) standardized methods to test **Function 6** (carbon cycling and
252 sequestration) (Table 3). For **Function 6**, there is a need to implement more fungal activity analysis as most
253 tests described this far only assess bacterial activities. Thus, the integration of more fungal enzyme
254 activities into the suite of standardized methods for soil quality assessment is essential (for example
255 determining the turnover of complex natural compounds such as lignin) (Baldrian, 2006). The ligninolytic
256 enzymes laccase and Mn-peroxidase, as well as the chitin degrading 1,4- β -N-acetylglucosaminidase, are
257 typical fungal enzymes of interest for ecosystem services (Jiang et al., 2014; Šnajdr et al., 2008). However,
258 since other organisms also produce these enzymes, including bacteria and plants (Bollag, 1992; de Gonzalo
259 et al., 2016), current methods do not specifically target fungal enzyme activities. As a result, the
260 implementation of molecular methods for assessing fungal communities are far less developed than those
261 for bacterial communities (Table 3).

262 The method of community level physiological profiling (CLPP) using the BiologTM system (Biolog,
263 Hayward CA, USA) was first developed in the late 1980s to identify bacteria of clinical importance by



264 assessing the consumption of 95 different carbon sources in a microtiter plate. The technique was then
265 extended to identify bacterial strains from environmental mixed microbial communities samples using
266 select carbon sources (Garland, 1997). Currently, the technique is frequently used to assess the effects of
267 contaminants on soil microbial activity (Bloem and Breure, 2003; Schmitt et al., 2004). As such, the CLPP
268 method has become a measure of microbial functional diversity in soil (e.g. Gomez et al., 2006) and was
269 used to distinguish the biodiversity of soil microbial communities in monitoring programs (Rutgers et al.,
270 2016). Even though the method is easy to use, it does have some drawbacks (Winding and Hendriksen,
271 2007). The technique is based on the utilisation of select carbon sources, which when consumed result in
272 reduction, and thus colour change, of a tetrazolium indicator dye (Garland and Mills, 1991). This reaction
273 is based on the dehydrogenase enzyme activity of cultivable, fast growing, aerobic, eutrophic
274 microorganisms (largely bacteria). Consequently, this technique does not reflect the full spectrum of
275 microbial species within a mixed soil community. Additionally, due to the artificial growth conditions
276 required in the test, it is argued that the method does not reflect the microbial community diversity and
277 its function of a given soil (Glimm et al., 1997). On the other hand, however, standardized conditions allows
278 for direct comparisons between microbial communities in different sites, for example, independent of the
279 abiotic conditions, thus making CLPP a popular method for toxicology testing (Preston-Mafham et al.,
280 2002).

281 Isothermal micro-calorimetry is another technique that involves the direct measurement of
282 energetics in soil and provides a functional link between energy flow and the composition of belowground
283 microbial communities at a high taxonomic level (Herrmann et al., 2014). With this method, an integrative
284 determination of the metabolic activity of soil bacteria and fungi is achieved. The integrated assessment
285 of substances' and energy turnover has high potential to elucidate the regulation of soil ecological
286 functions. However, the substantial costs for the acquisition of this very specific instrumentation is
287 considered as a major drawback. Furthermore, the measurement requires water saturation of the soil and,



288 thus, the samples are modified. Since calorimetry has been rarely used and data and publications are few,
289 this method is considered not ready for standardization.

290 The methods targeting thymidine or leucine incorporation into microbial biomass can be used to
291 determine microbial growth rates and efficiencies (Bååth et al., 2001; Rousk, 2016). Growth rate is a
292 fundamental reference for numerous other microbial properties and functions. For example, it is required
293 to calculate microbial carbon use efficiency (CUE) as a key-parameter describing C-substrate turnover and
294 storage in soil (Liu et al., 2018; Spohn et al., 2016; Takriti et al., 2018). Furthermore, the method can be
295 used to assess the adverse effects of toxic chemicals on the microbial community (Modrzyński et al., 2016;
296 Rousk et al., 2009a). The drawbacks of these two methods are: (i) specific training is required, (ii)
297 laboratories must have a permit to manipulate radioactive isotopes, and (iii) there are higher costs for
298 proper handling and disposal of ^3H -labelled radioactive material. As an alternative, the incorporation of
299 the stable isotope ^{18}O from labelled water into soil microbial DNA can be used to distinguish growing and
300 non-growing microorganisms based on the gradient-separation of ^{18}O DNA and ^{16}O DNA (Schwartz,
301 2007). The ^{18}O stable isotope method has been improved by sequencing a marker gene from fractions
302 retrieved from ultracentrifugation to produce taxon density curves; thus enabling researchers to estimate
303 the percent isotope composition of each microbial taxon's genome (Schwartz et al., 2016). This method
304 continues to be advanced and, although not used often, could have a high potential for future
305 standardization.

306 There are simplistic methods available to determine organic matter decomposition which are
307 indicative of C cycling (**Function 6**). The tests listed in Table 3 are based on measuring the weight loss of
308 introduced organic materials of different complexity in soil over time. The tests are relatively easy to
309 perform and inexpensive, however, degradation activity is not exclusive to microorganisms but can also
310 include invertebrates. The OECD litter bag test (OECD, 2006) for site specific assessment of organic matter
311 decomposition uses wheat straw as the substrate and provides clear evidence of cellulose degradation. In



312 general, the litter bag tests provides evidence for the degradation of naturally occurring plant material in
313 soil. Results do, however, depend on the mesh size of the litter bags (increasing exclusion of soil animals
314 with decreasing mesh size). On the other hand, plant material or litter is hard to standardise with the
315 results largely depending on the composition of the plant material. As such, artificial cellulose has been
316 successfully used for a laboratory procedure to assess organic matter decomposition (Kvas et al., 2017).
317 Another alternative to the litter bag test is the use of tea bags (Keuskamp et al., 2013). Tea bags can be
318 purchased to contain a consistent quality of material, and so this method is preferred by citizen science
319 (e.g. farmers to assess the soil quality of their land). In order to better distinguish the degrading abilities
320 of different soil microbiomes, the test could be modified to use different types of tea that contain
321 recalcitrant material to a different extent. Another test for future method development is the Bait Lamina
322 test (ISO 18311, 2016) used to assess the degradation of organic matter in field soil by grazing
323 invertebrates (Jänsch et al., 2013; Kvas et al., 2017). It is a simple test that can easily be adapted for use
324 under controlled laboratory conditions (Jänsch et al., 2017).

325 Methods for the determination and assessment of greenhouse gas emissions from soil (**Function**
326 **7**) have already been standardized or are well advanced in the standardization process (Table 2). They are
327 mostly focused on measuring concentrations of greenhouse gases, like CO₂, CH₄ and N₂O, as well as their
328 fluxes as endpoints. In addition, molecular biology methods that estimate the relative abundance of
329 functional microbial guilds or taxa gives new insight into the ecology of microorganisms involved in the
330 formation of greenhouse gases. For example, the qPCR measurement of key N₂O functional genes has
331 allowed researchers to link N₂O reduction capacity to reduced greenhouse gas emissions in soil amended
332 with organic matter (Xu et al., 2018). Additionally, the quantification of functional gene sequences related
333 to methane generation and methane oxidation, respectively, yields detailed insights into the functional
334 potential of climate change-affected permafrost soils (Yergeau et al., 2010).



335 For **Function 8** (soil structure affecting soil water, gas balance and filtration function), there is clear
336 evidence that microbial activity and biomolecules substantially contribute to the formation and stability
337 of micro-aggregates, and thus to the structure, pore system and pre-consolidation stress of soils (Six et al.,
338 2004). While existing parameters, such as enzyme activities, are not clearly indicative in this regard (Beck
339 and Beck, 2000), glomalin can be considered as a biochemical marker of soil aggregation. This glycoprotein
340 is produced by microorganisms, especially arbuscular mycorrhiza fungi, and significantly increases
341 aggregate formation and stability (Rillig, 2004; Rillig and Mummey, 2006). The existing protocols for
342 extraction (chemical extraction combined with autoclaving) and determination of glomalin, either by using
343 the Bradford protein assay, enzyme-linked immunosorbent assay (ELISA), or LC-MS method (Bolliger et al.,
344 2008; Janos et al., 2008), open the possibility for its standardization in the near future. It should be noted,
345 however, that a well-equipped and experienced laboratory is required to perform this method.

346

347 **4 Transforming standardized methods into indicators of soil quality**

348 As recently underlined by the European Food Safety Agency (EFSA) in a scientific opinion ‘addressing the
349 state of the science on risk assessment of plant protection products for in-soil organisms’, there is an
350 urgent need to modernize pesticide risk assessment by implementing specific protection goals for in-soil
351 organisms which are key drivers of a wide range of functions supporting ecosystem services (Ockleford et
352 al., 2017). There currently exists a multitude of methods that can potentially be used for this task. Here,
353 we have identified in the body of this paper a number of methods that are presumably suitable for further
354 evaluation and standardization with regard to their scientific value and practical applicability. These
355 prospective standardized methods will not only be useful to identify adverse effects on the soil
356 microbiome, but also to conduct comparable studies in laboratories all over the world to define normal
357 operating ranges of microbial activity in soil and respective quality indices and threshold values.



358 It is clear that all parameters taken together reflect the potential of a microbial community to
359 perform a certain function and not solely a specific (actual) activity. This is important to understand to
360 interpret the values of a given endpoint in relation to both energy fluxes and compound transformation
361 rates, which can largely depend on intrinsic properties such as pedoclimatic conditions, and nutrient
362 availability as well as extrinsic properties such as anthropogenic effects, and soil cultivation measures. To
363 make use of these methods as indicators for soil quality, there are several requirements that need to be
364 included. These involves the assessment of the normal operating range of soil that include natural and
365 dynamic fluctuations of a given endpoint. The methods need to be implemented into a framework, which
366 takes into account site-specific conditions including soil type, pedoclimate and land-use. Additionally,
367 there is a requirement for the assessment of resistance and resilience of a given microbial endpoint to see
368 how much it is affected by a soil disturbance and whether or not it can recover (e.g. return to its original
369 state) after the disturbance has disappeared. Also, the use of a test battery to measure a range of
370 interconnected endpoints is recommended (Ockleford et al., 2017) to integrate different biological and
371 other parameters (e.g. soil pH, organic carbon content) into multiparametric indices (Bastida et al., 2008;
372 Kvas et al., 2017). Finally, to fully understand soil microbial functioning, a task was envisioned to
373 investigate the linkage between the genetic functional potential and the available resources, termed the
374 soil metaphenome (Jansson and Hofmockel, 2018). This will require even further integration and
375 assessment of multiple parameters and test methods. Reaching that goal will surely promote soil
376 ecological research but, at the current stage, may clearly go beyond the applied aim of standardization to
377 release easy-to-use targeted methods.

378 The critical evaluation of existing and non-standardized methods is required to further select and
379 standardize new methods to assess soil quality. For methods linked to the molecular analysis of soil
380 microbiomes; there is a need to ensure that worldwide activities are synchronized to propose important
381 standards that are well accepted by the scientific community. For example, recently, the Earth Microbiome



382 Project (www.earthmicrobiome.org) has proposed primer pairs to barcode soil bacteria in a standardized
383 manner. Furthermore, new bioinformatic pipelines have been developed that are being used more and
384 more as standard procedures. Finally, to improve the reproducibility of data it has been agreed that a
385 complex mixture of microorganisms (MOC) must be implemented as a control in every experiment. The
386 exact composition of the MOC is still under discussion, however, it is clear that if further developments of
387 microbial bar coding and/or metagenomics methods are to be implemented into ISO guidelines, an MOC
388 is required.

389 ISO standardization committees are open circles and the presented selection and valuation of
390 methods may not be complete. Environmental scientists are solicited to propose new work items enlarging
391 the current catalogue of biological methods for future standardization. Accordingly, this opinion paper
392 aims at initiating a broader discussion intended to improve the measurement of microbial functions for
393 soil quality assessment.

394

395

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397

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832 **Table 1.** List of criteria for the selection of indicators for microbial functional indicators, based on Faber et
 833 al. (2013) and Pulleman et al. (2012), with slight modifications by the authors.

Criteria	Measured by	Low Score	High Score
a) Practicability	Lab equipment	Very few labs have the equipment needed	All labs would be able to carry out the work
	Skills	Specialist skills are needed	General skills would suffice
b) Cost efficiency	Capital start-up	More than €100 000	Less than €2000
	Cost per sample	More than €100	Less than €2
	Labour needed in the lab	High labour demand	Low labour demand
	Labour needed in the field	High labour demand	Low labour demand
c) Policy relevance	Focus on ecosystem processes and services	Weak links with existing or planned legislation	Strong links with existing or planned legislation
d) Sensitivity	Effect of soil properties	No response or idiosyncratic response	The indicator responds characteristically to change
	Effect of land use	No response or idiosyncratic response	The indicator responds characteristically to change
	Effect of disturbance	No response or idiosyncratic response	The indicator responds characteristically to change
e) Selectivity		Endpoint affected by numerous variables	Endpoint only affected by parameter under investigation
f) Reproducibility		Low or largely varying reproducibility among replicates	Highly reproducible
g) Use as an indicator	Status quo	Not in use already	In use already
h) Handling and availability of organisms ¹		Rare and/or difficult to obtain Difficult to keep Largely varying quality/fitness Seasonal availability	Easy to obtain Easy to keep Easy to provide with constant quality/fitness Year-round availability
i) Fit for use as an indicator	Significance / explanatory power	Weak relationship to ecological function	Strong relationship to ecological function
	Standardized	Methods are not ready for general use or standardization (i.e. low experience, no SOPs ²)	Methods are already in general use, preferably as standard (e.g. OECD)
	Spatio-temporally relevant	Spatio-temporally only relevant for a small plot at one point in time	Representative for more than one site and/or more than one point in time
	Understandable	Difficult to explain in a policy situation	Easily understood in a policy situation



j)	Experience	Literature data	Low amount of information on performance and outcome, e.g. <10 publications	High amount of information on the performance and outcome, e.g. >10 publications, existing ring test(s)
k)	Data evaluation	Database	No or hardly any existing data available or not freely available	Freely available and sound database for data evaluation

834 ¹ Only relevant for faunal species. Does not apply to soil microorganisms that are tested with their natural
835 abundance in mixed communities.

836 ² Standard operating procedures



837 **Table 2.** Methods already validated and published as ISO standards for determining potential microbial
 838 biomass and activities for soil quality.

Microbial biomass and respiration (some relations to Functions 1 and 6)	
ISO 14240-1	Determination of soil microbial biomass – Part 1: Substrate induced respiration method
ISO 12240-2	Determination of soil microbial biomass – Part 2: Fumigation – extraction method
ISO 16072	Laboratory method for determination of microbial soil respiration
ISO 17155	Determination of the activity of the soil microflora using respiration curves
ISO 11063	Direct soil DNA extraction
ISO 17601	Quantification of the abundance of microbial groups in soil DNA extract
ISO/TS 29843-1	Method by phospholipid fatty acid analysis (PLFA) and phospholipid ether lipids (PLEL) analysis (
ISO/TS 29843-2:	Method by phospholipid fatty acid analysis (PLFA) using the simple PLFA extraction method
(Potential) microbial enzymatic activities: C, N and P turnover (Functions 4 and 6)	
ISO/TS 22939 ¹	Measurement of enzyme activity patterns in soil samples using fluorogenic substrates in micro-well plates
ISO/DIS 20130 EN ²	Measurement of enzyme activity patterns in soil samples using colorimetric substrates in micro-well plates
ISO/TS 23753-1	Determination of dehydrogenase activity in soils – Part 1: Method using triphenyltetrazolium chloride (TTC)
ISO/TS 23753-2	Determination of dehydrogenase activity in soils – Part 2: Method using iodotetrazolium chloride (INT)
ISO 14238	Biological methods – Determination of nitrogen mineralization and nitrification in soils and the influence of chemicals on these processes
ISO 15685	Determination of potential nitrification and inhibition of nitrification – Rapid test by ammonium oxidation
Potential microbial activities: biodegradation of pollutants (Function 3)	
ISO 11266	Guidance on laboratory testing for biodegradation of organic chemicals in soil under aerobic conditions
ISO 14239	Laboratory incubation systems for measuring the mineralization of organic chemicals in soil under aerobic conditions
ISO 15473	Guidance on laboratory testing for biodegradation of organic chemicals in soil under anaerobic conditions
Potential microbial activities: turnover greenhouse gases (Function 7)	
ISO/DIS 20951	Guidance on methods for measuring greenhouse gases (CO ₂ , N ₂ O, CH ₄) and ammonia (NH ₃) fluxes between soils and the atmosphere
ISO/TS 20131-1	Easy laboratory assessments of soil denitrification, a process source of N ₂ O emissions -- Part 1: Soil denitrifying enzymes activities
ISO/TS 20131-2	Easy laboratory assessments of soil denitrification, a process source of N ₂ O emissions -- Part 2: Assessment of the capacity of soils to reduce N ₂ O
Potential microbial activities: organic matter decomposition (Function 6)	
ISO/CD 23265 ³	Test for measuring organic matter decomposition in contaminated soil

839 ¹ Measured enzyme activities: Arylsulfatase E.C. 3.1.6.1; α-glucosidase E.C. 3.2.1.20; β-glucosidase E.C.
 840 3.2.1.21; β-xylosidase E.C. 3.2.1.37; cellobiosidase E.C. 3.2.1.91; N-acetylglucosaminidase E.C. 3.2.1.52;



841 phosphodiesterase E.C. 3.1.4.1; phosphomonoesterase E.C. 3.1.3.2; leucine-aminopeptidase E.C.
842 3.4.11.1; alanine-aminopeptidase E.C. 3.4.11.12.

843 ² Measured enzyme activities: Arylamidase E.C. 3.4.11.2; arylsulfatase E.C. 3.1.6.1; α -glucosidase E.C.
844 3.2.1.20; β -glucosidase E.C. 3.2.1.21; β -galactosidase E.C. 3.2.1.22; N-acetylglucosaminidase E.C.
845 3.2.1.52; phosphatase E.C. 3.1.4.1; acid phosphatase E.C. 3.1.4.1; alkaline phosphatase E.C. 3.1.4.1;
846 urease E.C. 3.5.1.5.

847 ³ Degradation of cellulose under laboratory conditions.



848 **Table 3.** Potential new methods for the ISO standardization process and assessment according to the
 849 “logical sieve” selection criteria (described in Table 1).

Method	Source	Function addressed	Assessment ¹											
Function 4. Nutrient cycling (N and P)			a	b	c	d	e	f	g	h	i	j	k	
Functional genes assessed by real time qPCR			1-2	3	5	5	5	5	5	5	na ²	4	4	3
Ammonium monooxygenase gene (<i>amoA</i>)	Levy-Booth et al., 2014	quantify the abundance of nitrifying microbes	1-2	3	5	5	5	5	5	5	na ²	4	4	3
Ammonium monooxygenase gene (<i>amoB</i>)	Norton et al., 2002	quantify the abundance of nitrifying microbes	1-2	3	5	5	5	5	5	5	na	4	4	3
Nitrogenase gene (<i>nifH</i>)	Gaby and Buckley, 2012	quantify the abundance of N fixing microbes	1-2	3	5	5	5	5	5	5	na	4	4	3
Various genes driving P turnover	Bergkemper et al, 2016	quantify the abundance of microbes driving P transformation	1-2	3	5	5	5	5	5	5	na	4	4	3
Function 5. Pest control and plant growth promotion														
Specific mtDNA sequences assessed by real time qPCR	Voříšková et al., 2017	quantify the abundance of arbuscular mycorrhiza	1-2	3	5	5	5	5	5	5	na	4	3	2
Specific ITS sequences assessed by real time qPCR	Sakakibara et al., 2002	quantify the abundance of ectomycorrhizal fungi	1-2	3	5	5	5	5	5	5	na	4	3	2
Specific ITS sequences assessed by real time qPCR	Savazzini et al., 2008	quantify the abundance of biocontrol active Trichoderma fungi	1-2	3	5	5	5	5	5	5	na	4	3	2
Function 6. Carbon cycling and sequestration														
Enzyme activity of fungi	Eichlerová et al., 2012	determine activity of laccases	4	4	5	5	3	3	5	5	na	4	5	4
	Bach et al., 2013	determine activity of phenoloxidases	4	4	5	5	3	3	5	5	na	4	5	4
Community level physiological profiling (CLPP, “Biolog”)	Garland and Mills, 1991	determine degradation of a set of carbon sources	3	4	3	1	1	5	3	na	1	5	3	
Microcalorimetry	Prado and Airoidi, 2001; 2003	quantify microbial energy turnover	1	2	1	3	3	5	1	na	3	3	1	
[³ H]-leucine or [³ H]-thymidine incorporation	Bååth, 1998; Bååth et al., 2001; Rousk et al., 2009b	quantify microbial growth rate and efficiency	1	2	5	5	4	5	5	na	4	4	2	
[¹⁸ O] incorporation into DNA from labelled water	Schwartz, 2007; Schwartz et al., 2016	quantify microbial growth rate and efficiency	2	3	5	5	4	5	5	na	3	2	2	
Organic matter decomposition	OECD, 2006; Knacker et al., 2003	assess organic matter degradation and therefore C cycling	5	5	4	4	5	4	5	na	5	5	5	
Litter bag technique	Bockhorst and Wardle, 2013	assess the degradation of plant litter material	5	5	4	4	5	4	5	na	5	5	5	
Tea bag technique	Keuskamp et al., 2013	assess the degradation of tea leaves	5	5	4	4	5	4	5	na	5	5	5	

850



851 **Table 3.** Continued

Method	Source	Function addressed	Assessment ¹
Funct. genes within C cycle assessed by real time qPCR	El Azhari et al., 2008	quantify the abundance of microbes able to degrade protocatechuate (<i>pcaH</i>) a key intermediary metabolite of the β -ketoacid pathway	1-3 5 5 5 5 5 na 4 4 3
	El Azhari et al., 2010	quantify the abundance of microbes able to degrade catechol (<i>cat A</i>) a key intermediary metabolite of the β -ketoacid pathway	1-3 5 5 5 5 5 na 4 4 3
Function 7. Greenhouse gas emissions			
Methyl coenzyme M reductase (<i>mcrA</i>) assessed by real time qPCR	Steinberg and Regan, 2009	quantify the abundance of methane producing microbes	1-3 5 5 5 5 5 na 4 4 3
N ₂ O reductase gene (<i>nosZ</i>) assessed by real time qPCR	Jung et al., 2013	quantify the abundance of N ₂ O reducing microbes	1-3 5 5 5 5 5 na 4 4 3
Methane reductase gene (<i>pmoA</i>)	Kolb et al., 2003	quantify the abundance of methane reducing microbes	1-3 5 5 5 5 5 na 4 4 3
Nitric oxide reductase gene (<i>cnorA</i>) assessed by real time qPCR	Dandie et al., 2007	quantify the abundance of methane reducing microbes	1-3 5 5 5 5 5 na 4 4 3
Function 8. Soil structure affecting soil water, gas balance and filtration capacity			
Determination of glomalin	Bolliger et al., 2008; Janos et al., 2008; Wright et al., 1998	determine the content of glomalin in soil as a proxy of soil aggregation	3 3 4 4 4 5 3 na 4 4 2

852 ¹ Overall scoring in case of several measures for one criterion. Fulfilment of criterion described by
 853 numbering (colour code): 1 (red) very low; 2 (orange) low; 3 (yellow) medium; 4 (light green) good; 5
 854 (dark green) very good.

855 ² na = not applicable

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