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Editorial Support
on behalf of the SOIL Editorial Board

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Dear Natascha Töpfer, dear editorial board members,

As the corresponding author and on behalf of all coauthors of the manuscript soil-2019-42 "*Identification of new microbial functional standards for soil quality assessment*" I herewith re-submit a revised version of the manuscript. First of all we want to express our appreciation of the very constructive and sound suggestions and criticism of both reviewers. Following these advice, we significantly revised the manuscript and added text passages. Additionally we did some further polishing in style and language.

Please find our response to the reviewers' suggestions in the following text (marked as text in italics).

I very much hope that the revised version, which we think was substantially improved, finds your and the reviewers approval.

With kind regards,

S. Thiele-Bruhn

Reviewer #1 Robert Griffith

This paper discusses new methodologies and opportunities offered by molecular methodologies to provide microbiological indicators for assessing soil quality. In essence, it reports on key issues raised during recent discussions within the International Organization for Standardization (ISO), and identifies the need to focus on soil functions of relevance to ecosystem services as recognised for example in the Millennium Ecosystem Assessment. A key focus of the paper is in highlighting and scoring the potential of qPCR approaches for quantifying functional gene abundances of relevance to providing simple metrics relevant for quantification of biogeochemical fluxes (which are difficult to quantify directly).

The paper is generally well written, interesting, and delivers on synthesising the current broad status with respect to these issues, and additionally proposes some potentially new indicator approaches which could be implemented. As such, I feel it makes a useful contribution. The paper could be improved by offering more critical analyses of the approaches; as well as authoritatively defining the new science needed to facilitate the implementation of more robust soil microbiological indicators.

We especially added further aspects and discussions on soil microbial methods, which also relate to the points raised by reviewer # 2, E. Hannula.

Three areas which could be elaborated on further I feel are highlighted below. Perhaps fully covering them in detail extends beyond the remit of this manuscript, so I leave it to the editor to decide whether they should be expanded upon in the article (alternatively I guess these publically viewable comments may constitute a contribution to the “discussion” format of the journal. . .).

1. Indicator targets within the global soil geographic context. The paper briefly mentions this on line 365 (“methods need to be implemented into a framework, which takes into account site-specific conditions”), but offers no specific ways forward for this critical issue. Are elevated abundances of a functional indicator always “desirable”, and how might indicator target values, and indeed the indicators themselves differ for different soil systems? I’m not sure if we even have a good soil classification system or framework that allows us to set regionalised targets for the simple variable of soil carbon, and I sense this is what causes pushback on soil targets from industry and policymakers. Given this, could proposing even more microbiological variables be deemed somewhat premature?

This is a very good point and emphasizes the need to gather a database of soil microbial parameters in worldwide soils (with links to soil chemical, physical and pedoclimatic data) in order to get some idea about ‘normal’ value ranges. We fully agree that ‘higher’ values do not necessarily mean ‘better’ and such typical value ranges are needed to identify unusual aberrations. However, this fundamental discussion would lead a bit beyond the scope of this paper, which is focused on the methods that are required to receive the results, independent from how we store and assess these results. Consequently and pointing into the direction raised by R. Griffiths we added on lines 397-399: “Undoubtedly, this requires further joint efforts in order to generate comprehensive databases from which normal operating ranges of values for a given proxy can be read. Such a task calls for standardized methods to obtain comparable data”.

2. Relatedly, what is the evidence that gene abundance relates to functions of relevance to ecosystem services? It is often stated that you cannot infer anything about processes from gene abundance alone, but I feel there is little literature actually specifically addressing this with robust contrasts within an ES indicator context. For example comparative data for ammonia oxidation gene abundances does actually appear to relate to nitrification rates in certain studies, so do we need a critical meta-analyses of this now for a variety of indicators? Again, relating to the point above, do we always want high nitrification, high litter decomposition, high enzyme activity etc in all soil systems; and is there any evidence that molecular detection of elevated pathogens reliably informs on plant health...Essentially what do these measures really tell us about desirable ES outcomes, and if there is little information available, then what can be done to progress?

We share the view of the reviewer that we are only at the beginning to understand the meaning of microbial activity and functional gene abundance in terms of ecosystem services and functioning of soils. The following phrases of the added text mostly apply to this comment.

250-256 "Also, evidence is increasing that functional gene abundance and community structure are closely linked to related microbial activities and their increase or decrease, e.g. through agricultural fertilizer regime or soil contamination (Levy-Booth et al., 2014; Ouyang et al., 2018; Xue et al., 2018). However, also contrasting findings have been reported, pointing to the fact that functional gene abundance and diversity is less affected by short-term changes, e.g. due to soil moisture changes (Zhang et al., 2019). A critical meta-analysis of existing data and reports, respectively, would be timely to better identify and generalize the linkage of functional gene abundance and ecosystem services."

402-408: "Here the use of DNA based methods, which provide a measure of a microbial community's potential to perform a given process, might be more useful than using RNA. The RNA rather indicates actual activities, which may highly fluctuate in time and space, and thus are of less significance as an indicator. However, free DNA released from dead microbes is often highly resistant in soil, which might result in an over estimation of a potential function. This needs to be taken into account when interpreting the data. Recently, methods that extract DNA only from living cells have been described, but their use has not been yet introduced into recent standardization activities."

3. Standardisation: essential for policy, but bad for science? Given the paper's policy focus, it appears to heavily endorse standardisation. However molecular ecology is a rapidly growing field, and technologies change (eg sequencing platforms) which causes issues with implementing standardised protocols. Scientific developments must be free to progress in order to develop the deep and often complex understanding of processes required to implement meaningful process indicators. It would be useful to highlight this potentially conflicting issue.

We agree that standardization is a balancing act. We aim to receive largely comparable data (asking for defined methods) but want to use the latest methods for that purpose (asking for highest flexibility). So we added on lines 429-436: "Lastly, it must be noted that standardization of methods is inevitably a balancing act. On one hand, standardization provides defined methods that are essential to obtain comparable data, e.g. for integration in large, joint databases. On the other hand, it requires setting a specific method for several years. Consequently, scientific progress cannot be easily adopted, or at least with a delay, considering that standards are revised every five years, which may be a barrier to the introduction of new approaches

resulting from technological evolution, especially in the fast developing field of molecular biology methods. Hence, it is also the aim of this paper to have an open discussion to identify the best suitable methods with an assumed longer period of validity.”

Reviewer #2 Emilia Hannula

The topic of relating currently available measurement techniques to the soil functions is a very timely issue. Using logical sieve approach, this article investigates the suitability of commonly used methods to evaluate the soil functions and ecosystem services, one at the time. The paper is very clearly written and presents solid, interesting research. It could benefit from section on future directions and more updated list of currently available (molecular) techniques. Will these be better or do we already have a golden standard that will tell us all we want to know?

For sure we don't have the golden standard (otherwise new method development would be obsolete). We added further text and discussion especially on soil molecular biology methods and parameters.

200-210: metagenomics for degrading activities.

218-226: Problems with data handling and bioinformatics, respectively.

250-256: Discussion on linkage between functional gene abundance and function.

It is clear that the work has started already in 2013 and in the past six years huge developments have been achieved in the toolbox available to measure soil functions and especially diversity. Authors mention the 'Earth microbiome project' and standardization of primer sets and pipelines to study bacterial diversity as an emerging technique. However, in reality, this method is the most used method to study fungi and bacteria in soils and is considered fairly standard as all labs use the same regions (V4 and ITS2) and often same primers. There is a recent article on the methodological comparison on bacteria (Ramirez et al. 2018, Detecting macroecological patterns in bacterial communities across independent studies of global soils, Nature microbiology). There is much discussion on using 'mock' (not MOC)-communities to standardize the methods and most labs use these already.

Discussion on barcoding using high throughput sequencing and in special the use of primer pairs targeting the V4 region of the 16S rRNA gene and ITS2 region for bacterial and fungal barcoding was added to the text. Also we evaluate the suitability of bioinformatics pipelines for cross comparison. (Lines 157-167).

Furthermore, the field is moving towards true (shotgun) metagenomics sequencing which yields data on all soil organisms and their functions. This approach is emerging but will definitely be worth discussing in this context. It will have no bias of PCR but the quantity of soil used and DNA extraction efficiency related issues will remain. In short, the 'newer' methods should be discussed. For the methods used to study diversity and future avenues in that field, following paper can be cited (Geisen et al. 2019, A methodological framework to embrace soil biodiversity, SBB).

Metagenomics are further discussed now on lines 218-226. We thank for the helpful literature reference that was included in the new added text.

In the abstract, it is mentioned that especially fungal functions are difficult to measure. The article presents a suite of measures traditionally used and recommends to use enzyme production to measure fungal functions. For bacteria, qPCR based methods are recommended. This discrepancy in recommendations should be discussed. Why is it not feasible in all case to look at process rates (i.e. decomposition), but details on the amount of enzymes and/or amount of organisms performing the task are needed? In which scenarios and which scales which measurement is needed? Considering this would make the article stronger and bring more to the field. In the evaluation of function 6 (carbon cycling) it is simply stated that because these enzymes are often produced by other organisms, molecular methods are less developed. This is partly true but can be related also that the enzyme measurements are pretty good and give the actual rate of enzyme measured. Furthermore, there are existing primer sets for quite some of the CAZys (for example: Edwards et al. 2011: Simulated Atmospheric N Deposition Alters Fungal Community Composition and Suppresses Ligninolytic Gene Expression in a Northern Hardwood Forest, PlosOne Gorfer et al. 2011: Community profiling and gene expression of fungal assimilatory nitrate reductases in agricultural soil, ISMEJ Chen et al. 2013: Comparative analysis of basidiomycetous laccase genes in forest soils reveals differences at the cDNA and DNA levels. Plant and Soil Hannula & van Veen (2016) Primer Sets Developed for Functional Genes Reveal Shifts in Functionality of Fungal Community in Soils. Frontiers in Microbiology These measurements from DNA have problems as different fungi have different copy numbers/ types of genes and not everything that is in the DNA is expressed. Indeed, work is needed to get these methods ISO certified but more discussion on the future directions would be welcome.

Fungi: Information on methods targeted towards fungi was added and the recommended literature included. Fct. 6. The specific problem of linking copy numbers of fungal genes to the size of a functional population was emphasized (lines 291-297).

Decomposition rates: We present and propose molecular methods (being an indirect measure of activities, rather looking on the organism side) but also on activities such as the litter bag and the tea bag method, thus targeting much more the effect point of view (see Table 3). We also added new text on the discussion in how far functional gene abundance can serve as a measure of true soil microbial functioning (lines 250-256).

Scenarios and scales: We agree that a bundle of methods is proposed with differences regarding application range and purpose (despite the function represented). To discuss this, would lead much deeper into the criteria (Table 1) d) sensitivity (with effect of soil properties, of land use and of disturbance on the test result), e) selectivity, and g) use as an indicator. We feel that this topic is surely relevant but would require too much detail information and discussion, making the manuscript too voluminous. We decided to leave the manuscript more concise and hope to find the reviewers' consent for that.

1 Identification of new microbial functional standards for soil 2 quality assessment

3
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23 **Abstract.** The activity of microorganisms in soil is important for a robust functioning of soil and related ecosystem
24 services. Hence, there is a necessity to identify the composition, diversity and function of the soil microbiome
25 indigenous soil microbial community for its functional properties using soil microbiological methods in order to
26 determine its the natural properties, functioning and operating range of soil microbial communities, and as well as
27 to assess ecotoxicological effects due to anthropogenic activities. Numerous microbiological methods currently
28 exist in the literature and new, more advanced methods continue to be developed; however, only a limited number
29 of these methods are standardized. Consequently, there is a need to identify the most promising non-standardized
30 methods for assessing soil quality and develop to transform them ~~these~~ into standards. In alignment agreement with
31 the “Ecosystem Service Approach”, new methods should focus more on soil microbial functions, including nutrient
32 cycling and greenhouse gas emission, pest control and plant growth promotion, carbon cycling and sequestration,
33 greenhouse gas emission, and as well as soil structure development and filter function. The few existing, ~~function-~~
34 ~~related standard~~ standardized methods available that focus on the function of the soil microbiome mostly include
35 measurements, like basal respiration, enzyme activities, and biodegradation of organic matter, under well-defined
36 conditions in the lab, and include ~~focus on the estimation of microbial biomass, basal respiration, enzyme activities~~
37 ~~related to nutrient cycling, and organic chemical biodegradation of organic chemicals~~. This paper sets out to
38 summarize and expand on recent discussions within the International Organization for Standardization (ISO), Soil
39 Quality - Biological Characterization sub-committee (ISO TC 190/SC 4) where a need was identified to develop
40 scientifically sound methods, which would best fulfil the practical needs of future users for assessing soil quality,
41 going beyond the existing test systems. Of particular note ~~was is~~ the current evolution of molecular methods in
42 microbial ecology that uses quantitative real time PCR (qPCR) to produce a large number of new functional
43 endpoints ~~and which are~~ more sensitive as compared to ‘classical’ methods. Quantitative PCR assesses the
44 abundance ~~activity~~ of microbial genes that ~~code for enzymes that~~ catalyse major transformation steps in nitrogen
45 and phosphorus cycling, greenhouse gas emissions, chemical transformations including pesticide degradation, and
46 plant growth promotion pathways based on the assessment of marker gene sequences that drive the related
47 processes. In the assessment of soil quality methods, it was found that most methods focus on bacteria and related
48 endpoints. Techniques to describe fungal communities as well as their functional traits are far less
49 represented ~~methods were significantly underrepresented~~. As such, techniques to analyse fungal enzyme activities
50 are proposed. Additionally, methods for the determination of microbial growth rates and efficiencies, including
51 the use of glomalin as a biochemical marker for soil aggregation, are discussed. Furthermore, field methods
52 indicative of carbon turnover, including the litter bag test and a modification to the tea bag test, are presented.

53 [However, it is obvious, that with increasing developments in high throughput sequencing technologies and big](#)
54 [data analyses, including metagenomics analysis, it will be possible to implement these technologies into](#)
55 [the standardisation process for assessing the functions of the soil microbiome. As a final note, ~~Overall~~Overall, it is](#)
56 suggested that endpoints should represent a potential function of soil microorganisms rather than actual activity
57 levels, as the latter can largely be dependent on short-term variable soil properties such as pedoclimatic conditions,
58 nutrient availability, and anthropogenic soil cultivation activities.

59 **1 Introduction**

60 Soils are one of the world's hotspots for biodiversity (Parker, 2010). Biota – both micro- and macro-organisms –
61 in soil form strong networks and complex food webs, which determines the efficacy of the soil ecosystem functions
62 (e.g. nutrient cycling, C storage and turnover, water retention, and modulation of soil structure) (Creamer et al.,
63 2016). These functions support a range of ecosystem services that are indispensable for soil use in agri-, horti- or
64 silviculture (Nannipieri et al., 2017). At the same time, soil biota are strongly impacted by various anthropogenic
65 activities including ongoing global and climate change, pollution, ~~and as well as~~ degradation and destruction of
66 the terrestrial environment (Gomiero, 2016; Montgomery, 2008; Wagg et al., 2014). Consequently, investigations
67 of the soil biome structure and function became an emerging topic in soil and environmental sciences (Griffiths
68 and Philippot, 2013). As such, the number of ~~studies on describing the ecology of soils~~ ~~ecology~~ and ecosystem
69 functioning has increased significantly over the past few decades and has resulted in the development of new
70 methods ~~for the assessment of microbial communities~~ (e.g. Guillaume et al., 2016; Tian et al., 2018). In
71 comparison, the ecotoxicological assessment of human impacts (e.g. chemical pollution and mechanical
72 compaction) using single species tests, which are well-established methods, has remained constant (Brookes, 1995;
73 Joergensen and Emmerling, 2006).

74 Characterizing the natural state of a soil's biome is ~~a quite~~ ~~+~~challenging task. In addition to its huge
75 structural and functional diversity, ~~the~~ soil biomes ~~is~~are influenced by strong temporal dynamics including
76 seasonal weather conditions and the enormous spatial heterogeneity, which ranges from field scale to microscale
77 (Kuffner et al., 2012; Regan et al., 2014; Suriyavirun et al., 2019). All of these intrinsic properties hinder the
78 interpretation of data obtained from the analysis of soil biomes and the measurement of their functional traits.

79 Despite the fundamental methodological advances over the past years, which allow for an in-depth
80 analysis of microbiomes and, to some extent, other soil-living organisms (e.g. Joergensen and Emmerling, 2006;
81 Paul, 2015; Yates et al., 2016), only a limited number of soil biological methods have been standardized (for details
82 see section 3). As a result, large ~~and often significant~~ deviations are observed ~~in the results obtained~~ ~~between~~when
83 non-standardized methods ~~protocols~~ are used (e.g. Strickland and Rousk, 2010). ~~This is especially true for methods~~
84 ~~that are based on high throughput sequencing approaches, where variability and bias in data can occur from the~~
85 ~~“wet-lab” steps right through the various bioinformatics pipeline analysis steps (Quince et al., 2017).~~ Therefore,
86 ~~the~~ comparability between datasets generated by different laboratories using different methods or modified
87 protocols of the same method ~~are~~is problematic. Consequently, the development of quality indices and threshold

88 values, respectively, for assessing soil quality is nearly impossible (Bastida et al., 2008). Presumably, this is why
89 the number of meta-analyses in soil biology remains small.

90 Given that there is a lack of harmonization between existing methods and, at the same time, a proliferation
91 of new methods, there is a need to identify the most promising methods described in the literature that can be
92 standardized to produce reliable indicators for soil quality (e.g. Philippot et al., 2012). At the Annual International
93 Organization for Standardization (ISO) meeting of TC 190 (Soil Quality) in Fukuoka, Japan in October 2013, a
94 decision was made to compile a list of available methods and to identify those that would be suitable for assessing
95 soil quality. Additionally, during a subsequent meeting of ISO TC 190/SC 4/WG 4 (Microbiological Methods)
96 held in Paris, France in March 2014, further discussions focused the criteria for suitable methods to be comprised
97 of microbial functional indicators. In this paper, we summarize the major outcomes of the discussions which took
98 place over the past several years within ISO TC 190/SC 4. Therefore, besides collating a list of criteria for the
99 selection of test methods for the future analysis of microbial functions in soil, the aim of this paper is to present
100 our opinion, as members of the ISO TC 190 committee, to initiate further discussion on possible methods that
101 should be standardized for future soil quality assessments.

102

103 **2 Criteria for the selection of methods**

104 Several papers addressing the task to identify suitable methods to be used as biotic indicators (usually including
105 faunal indicators) were published in the last few years, mainly in the context of EU research projects (e.g. Bispo
106 et al., 2009; Faber et al., 2013; Ritz et al., 2009; Römbke et al., 2010). Here, we propose to base the selection of
107 soil quality methods more on the “Ecosystem Service Approach” (MEA 2005) which is increasingly recognized
108 by both environmental scientists and regulatory agencies (Breure et al., 2012; Galic et al., 2012) and [which takes](#)
109 [into account that](#) ~~that~~ soils have been raised to the rank of a natural resource to be protected. As a consequence,
110 and in addition to method development and application (including the assessment of biodiversity as a prerequisite
111 for soil function), the focus of future activities should be the determination of soil microbial function as
112 recommended endpoints (Kvas et al., 2017; Nienstedt et al., 2012; van der Putten et al., 2010; TEEB, 2010).
113 Consequently, we propose to assess both existing and new methods for the selection of microbial functional tests
114 that support various soil ecosystem services. This structures our approach and simplifies the identification of
115 ecologically relevant methods, as well as, presumably [increasing](#) their acceptance by users, including the
116 regulatory and stakeholder community. The following [soil functions](#) ~~Functions~~ and **ecosystem services** have been

117 defined and are proposed to be used as a starting point for the development of future methods (MEA, 2005;
118 Ockleford et al., 2017):

- 119 (1) Biodiversity, genetic resources, cultural services;
- 120 (2) Food web support;
- 121 (3) Biodegradation of pollutants;
- 122 (4) Nutrient cycling (for example N and P);
- 123 (5) Pest control and plant growth promotion;
- 124 (6) Carbon cycling and sequestration;
- 125 (7) Greenhouse gas emissions; and
- 126 (8) Soil structure affecting soil water, gas balance and filtration function.

127 A second major criterion for selecting methods for standardization is its usability. The method should be applicable
128 in regulations (e.g. European and National agencies registering chemicals or products) and for the evaluation of
129 soil ecology and functioning as a fundamental aspects of soil quality (e.g. by stakeholders and researchers).
130 Moreover, the routine use of methods to inform farmers and site owners on soil quality as continuous assessments
131 of their land and land-use practises; could be an additional condition that would require the choice of easy-to-use
132 methods or possibly encourage the simplification of existing methods. Overall, frequently used methods generate
133 more data, which in turn is of high importance for the validation of threshold values. Therefore, the aim of this
134 process is to identify methods that are scientifically sound and best fulfil the practical needs of future users. The
135 most appropriate new methods, including those proposed in this article, need to be evaluated using the criteria
136 required for the standardization of ISO methods.

137 To assess possible methods, a list of criteria was used based on the 'logical sieve' approach (Ritz et al.,
138 2009). The list of criteria for the identification of functional indicators and associated methodologies (Table 1)
139 was an outcome of the EU FP7 EcoFINDERS project (Faber et al., 2013). The criteria were compiled after sending
140 a questionnaire to 25 partner institutions primarily working in the field of environmental science; mainly
141 representing academia but also regulators and subcontracting laboratories. These criteria are applicable for
142 different kinds of indicators and methods, including those addressing the functions of soil microbial communities.
143 In the following sections, we assume that existing ISO standardized methods partly already fulfil these criteria,
144 but not all relevant endpoints can be measured. Additionally, the most appropriate new methods, including those
145 proposed in this article, need to be evaluated using the same criteria required for the standardization of ISO

146 ~~methods. Therefore, the aim of this whole process is to identify methods which are scientifically sound and that~~
147 ~~best fulfil the practical needs of future users.~~

149 **3 Existing and new methods**

150 Current methods that have already been implemented as ISO standards are found in Table 2, whereas methods that
151 might be considered for future standardization are in Table 3. The compilation in Table 2 comprises methods to
152 quantify microbial biomass (e.g. through fumigation extraction of microbial biomass carbon (MBC) and DNA)
153 (**Function 6**) as well as for (further) analysis of structural microbial diversity (e.g. determination of microbial
154 fingerprints by phospholipid fatty acids (PLFA) analysis) (**Function 1**). Additionally, microbial biomass,
155 measured as respiratory activity, has been included in Table 2, although not directly linked to one of the ecosystem
156 services, as it provides important information on the activity of the complete microbiome (i.e. microflora and
157 microfauna). Soil basal respiration normalized to MBC (ISO 14240-1 and ISO 14240-2, 1997, Table 2) yields the
158 metabolic quotient qCO_2 , which is a sensitive indicator for microbial carbon use efficiency (Anderson and
159 Domsch, 1993). Interestingly, However, its use as an endpoint to assess anthropogenic and natural impacts on the
160 soil microbiome has been controversially discussed in literature (Wardle and Ghani, 1995). The microbial quotient
161 (MBC related to organic carbon content of a soil) is an indicator revealing changes in the microbial dynamic
162 equilibrium of soils in response to exposure to natural or anthropogenic stressors (Pankhurst et al., 2001).

163 The biodiversity function (**Function 1**) addresses parameters related to the structural diversity of the soil
164 microbiome. Here, respective ISO guidelines analysing PLFA, phospholipid ether lipids (PLEL) (ISO/TS 29843-
165 1, 2010; ISO/TS 29843-2, 2011) and DNA (ISO 11063, 2012; ISO 17601, 2016), have already been well
166 implemented into guidelines (Table 2). In addition, high throughput sequencing of barcodes of the ribosomal
167 operon (16S rRNA gene for bacteria and archaea and ITS [internal transcribed spacer] region for fungi) have
168 generated a large amount of data (Schöler et al., 2017). These approaches have been also used successfully for
169 other microbial groups like protists (using the 18S rRNA gene as a target). As it is well accepted that the use of
170 different primer pairs introduce different biases (Ramirez et al., 2018), standards have been recommended by
171 international initiatives. For example, the Earth Microbiome project (www.earthmicrobiome.org) recommended a
172 primer pair targeting the V4 region of the 16S rRNA gene and ITS2 region for bacterial and fungal barcoding,
173 respectively. Bioinformatics pipelines used for barcoding approaches are already well standardized and shared
174 worldwide among the scientists, which makes possible the cross comparison of various datasets from different

175 [labs](#). Unfortunately, this is not yet the case for the pipelines to analyse metagenomics datasets that are still under
176 [constant evolution making difficult cross-comparisons](#).

177 Although microbial diversity, per se, is not strongly correlated with a particular functional capacity, it is
178 clear that the loss of diversity can have an impact on microbial function (Thiele-Bruhn et al., 2012); at least for
179 relatively specific functions performed by narrow microbial guilds or taxa. This applies even more, when certain
180 taxa are closely linked to very specific functions including nitrifiers, methanogens, arbuscular- and ecto-
181 mycorrhizal fungi, and biocontrol microorganisms like *Trichoderma* (e.g. Hartmann et al., 2009; Hayat et al.,
182 2010; Lugtenberg and Kamilova, 2009; Peng et al., 2008; Singh et al., 2007; Xia et al., 2011). Therefore, the
183 interpretation of the outcomes from microbial community-based testing tends to be straightforward and closely
184 linked to **Function 4** and **Function 5**.

185 Food web support (**Function 2**) of higher trophic levels no doubt starts from soil microorganisms and
186 propagates through the trophic levels (e.g. earthworms) that are consumed by birds and mammals (Haynes, 2014;
187 Scheu et al., 2002; Scheu et al., 2005). However, the role of the microbiota in the soil food web is not fully
188 understood, since many eukaryotic organisms can be considered as meta-organisms, which carry their “own
189 microbiome” that itself is essential for life supporting functions. From this, it is unclear if environmental
190 microbiomes and host specific microbiomes complement one another. So far there are no comprehensive methods
191 (especially not those addressing microbial functions) or standards available to [assess-address](#) this problem. [A future
192 considering may employ](#) The use of stable isotope labelling of select carbon [or nutrient](#) sources as a promising
193 approach to follow food webs and degradation pathways [might provide one possibility in the future to assess food
194 webs in soil. This would be accomplished by combining carbon and nitrogen stable isotope fluxes determination
195 with phylogenetic analysis of the microorganisms labelled with the stable isotopes](#) (e.g. Coban et al., 2015;
196 Traugott et al., 2013; [Lueders et al., 2004](#)).

197 Methods to assess the biodegradation of pollutants (**Function 3**), as described above, are already
198 implemented into ISO guidelines (Table 2) and are part of legal frameworks including pesticide directives (EU
199 Regulation 1107/2009/EC; European Commission, 2009). A number of standard methods for the determination of
200 the potential of soils to degrade organic chemicals (**Function 3**) under both aerobic (ISO 14239, 2017) and
201 anaerobic (ISO 15473, 2002) conditions are available. [This emphasizes that](#) However, in the past, the development
202 of standard methods was mainly driven by the need to assess the ecotoxicological effects of anthropogenic
203 activities, such as chemical contamination of soils, rather than to describe and understand the natural properties
204 and functions of soils. [However, d](#) Defining methods for the determination of adverse effects of contaminants on

205 soil biota was not only ~~being done by~~ ISO, but it was also a major task of other organizations such as the
206 Organization for Economic Co-Operation and Development OECD). For example, there are OECD guidelines,
207 tests No. 216 and 217, for testing the long-term effects of single exposure chemicals on soil microbial nitrogen
208 and carbon transformation, respectively (OECD, 2000a; 2000b). As a result, it was decided early that the
209 standardization of methods for toxicity testing ~~sh~~ould not be the primary aim of the ISO sub-committee (ISO TC
210 190/SC 4). Metagenomics lead to the rapid discovery of new genes, which catalyse degradation processes of
211 xenobiotics and, consequently, offer new insight in the study of ecology of microbial degraders. The transfer of
212 this knowledge into operational bioindicators for the estimation of soil filtration capabilities (e.g. by using qPCR
213 or RT-qPCR assays to quantify those genes and their expression in soils, respectively) will be one of the major
214 tasks in soil science in the upcoming years. Jeffries et al. (2018) for example were able to successfully predict the
215 degradation rates of organophosphorus based pesticides in a contaminated soil by using metagenomics based
216 functional profiling.

217 Some of the existing standard methods that are listed in Table 2 focus on the estimation of enzyme
218 activities useful for soil quality assessment, which mainly contribute to **Function 4**. Here, the potential
219 dehydrogenase activity measurement is an indicator for general (potential) oxidoreductase activity in soil. Since
220 this measurement has been frequently used, there are large amounts of baseline data available on the toxic effects
221 of a range of pollutants in soil. Recently, additional potential enzyme activities related to the C, N, P and S cycle
222 have been used and are either standardized or are in the process for standardization.

223
224 The current evolution of molecular methods in microbial ecology has resulted in a large number of new
225 endpoints. It is well known that many of the new endpoints (e.g. using quantitative real-time PCR (qPCR)) are
226 more sensitive than classical methods that had been standardized in the past (Ribbons et al., 2016; Schulz et al.,
227 2016). This new metagenomics approach will be of high importance in the future, as it allows for the
228 implementation of information on new functional traits that can be standardized into an analytical pipeline. Direct
229 sequencing of soil DNA extracts, called metagenomics, is nowadays- a method that can easily be implemented to
230 obtain an overview of in-soil living organisms, including microorganisms. It is recognized that metagenomics is
231 no longer limited by sequencing possibilities but more by the availabilities of large-scale computing clusters to
232 analyse the amount of data generated. The future success of metagenomics in soil surveys will mainly depends
233 on the possibilities to standardize bioinformatics pipelines as well as on the availability of tools for big data
234 analysis and artificial intelligence. However, it must be mentioned that even a well-standardized and automated

235 [workflow will generate only data on the relative abundance of nucleotide sequences and not absolute values](#)
236 [\(Geisen et al., 2019\)](#). Thus at the moment for the assessment of new methods linked to **Functions 4 to 8**, qPCR
237 from soil DNA extracts (ISO 17601, 2016) plays a very important role in determining the abundance of single
238 marker gene sequences, which are indicative of specific transformation processes or soil functions. For example,
239 the quantification of nitrogen fixing microbes, nitrifiers and denitrifiers has been successfully implemented using
240 the *nifH*, *amoA* and *nirS/nirK* genes as markers, respectively (Henry et al., 2004; Hirsch et al., 2010; Ollivier et
241 al., 2010; Sessitsch et al., 2006). Similarly, the quantification of microorganisms involved in the β -ketoacid
242 pathway has been implemented by targeting *pcaH* (El Azhari et al., 2008) and *catA* (El Azhari et al., 2010) gene
243 sequences. Various methods for the assessment of soil microbial **Function 4** (nutrient cycling), **Function 5** (pest
244 control and plant growth promotion) and **Function 7** (greenhouse gas emissions) are proposed based on the qPCR
245 analysis of gene sequences coding for enzymes which trigger the respective function (e.g. Fish et al., 2013; Ribbons
246 et al., 2016; Smith and Osborn, 2009). Additionally, it should be noted that molecular methods based on the
247 assessment of specific marker genes for estimating the degradation potential in soil have already been proposed
248 both for PAHs (e.g. Cebron et al., 2008) and individual pesticides (e.g. Martin-Laurent et al., 2004). These could
249 be interesting for future standardization; however, if a method is very compound-specific and targeted, this could
250 limit its application range. Thus, these specific approaches will not be discussed further in this article.

251
252 Major advantages of qPCR assays to quantify gene sequence numbers, which can be used as proxies for
253 a given microbial process, are that they are: (i) highly standardized, sensitive, selective and reproducible, (ii)
254 designed for high throughput analysis, (iii) available for a wide range of targets, and (iv) methods that are relatively
255 cheap once the necessary analytical devices are on hand. Some training on the method is required, however, once
256 trained the assays are easy to perform. For example, numerous studies have already used the microbial functional
257 genes involved in nitrogen cycling to determine the status and to assess induced changes in the soil microbial
258 community (Levy-Booth et al., 2014; Nannipieri and Eldor, 2009; Wallenstein et al., 2006). Consequently, the
259 number of functional genes that are suited for use as specific indicators of soil function are continuing to grow in
260 the literature as researcher gain experience in this field and data becomes more prevalent. [Also, evidence is](#)
261 [increasing that functional gene abundance and community structure are closely linked to related microbial](#)
262 [activities and their increase or decrease, e.g. through agricultural fertilizer regime or soil contamination \(Levy-](#)
263 [Booth et al., 2014; Ouyang et al., 2018; Xue et al., 2018\)](#). However, also contrasting findings have been reported,
264 [pointing to the fact that functional gene abundance and diversity is less affected by short-term changes, e.g. due to](#)

265 [soil moisture changes \(Zhang et al., 2019\). A critical meta-analysis of existing data and reports, respectively, would](#)
266 [be timely to better identify and generalize the linkage of functional gene abundance and ecosystem services.](#)

267 Disadvantages, on the other hand, are that: (i) the quality of qPCR data depends on soil DNA extracts
268 (PCR inhibition), (ii) primer pairs, even degenerated ones, might not successfully amplify all microbes of the
269 functional group of interest, (iii) only genetic potential is resolved, and (iv) there is no differentiation between
270 active, dormant or dead microorganisms, when working with DNA as a template for the qPCR reaction. The
271 analysis of total RNA and of mRNA, which could help to overcome the latter problem, is currently not a suitable
272 alternative as it is highly dynamic in time and space and needs special care to stabilize the RNA extracted from
273 complex environmental matrices to avoid its degradation. Another problem of DNA analysis is the biological
274 representativeness of the results is solely based on a relatively small amount of soil (from few hundred mg to ten
275 g of soil) from which the DNA extracted. The use of small soil samples (< 1 g) simplifies the sample preparation
276 process for molecular biologists; however, it provides a poor representation of the indigenous soil microbial
277 community in the naturally inhomogeneous soil. Typically, the α -biodiversity declines with sample size while that
278 of β -biodiversity increases (Nicol et al., 2003; Penton et al., 2016). Lastly, it must be noted that the high
279 repeatability and reproducibility of molecular biology methods, including qPCR assays, depends on extraction,
280 purification and amplification of DNA or RNA. This is typically performed using commercial extraction kits;
281 however, by simply changing the commercial supplier of a kit can substantially change the results (Brooks et al.,
282 2015; Feinstein et al., 2009). This clearly challenges standardization since standard methods must not hinge on a
283 specific supplier.

284 Recently, ~~successful examples molecular tools for the assessment of the of~~ microbial phosphorous
285 turnover (**Function 4**) have been published (Bergkemper et al., 2016) where metagenomics data have been used
286 for the construction of primers for P mineralization, transport and uptake. As another example, the relevance of
287 anaerobic ammonium oxidation (anammox) for N cycling in soils has increased (Levy-Booth et al., 2014) along
288 with the development of analytical methods for high throughput analysis. Among the microorganisms in soil that
289 substantially govern pest control and plant growth promotion (**Function 5**), ~~the most numerous organisms are the~~
290 arbuscular mycorrhizae and ectomycorrhizal fungi [are of high relevance](#). These microorganisms are especially
291 abundant in the rhizosphere (Hartmann et al., 2009; Hayat et al., 2010; Lugtenberg and Kamilova, 2009). Methods
292 related to **Function 5** are listed in Table 3.

293 Several options exist for (additional) standardized methods to test **Function 6** (carbon cycling and
294 sequestration) (Table 3). For **Function 6**, there is a need to implement more fungal activity analysis as most tests

295 described ~~this far only~~mostly assess bacterial activities. Thus, the integration of more fungal enzyme activities into
296 the suite of standardized methods for soil quality assessment is essential (for example determining the turnover of
297 complex natural compounds such as lignin) (Baldrian, 2006). The ligninolytic enzymes laccase and Mn-
298 peroxidase, as well as the chitin degrading 1,4- α -N-acetylglucosaminidase, are typical fungal enzymes of interest
299 for ecosystem services (Jiang et al., 2014; Šnajdr et al., 2008). However, since other organisms also produce these
300 enzymes, including bacteria and plants (Bollag, 1992; de Gonzalo et al., 2016), current methods do not specifically
301 target fungal enzyme activities. As a result, the implementation of molecular methods for assessing fungal
302 communities are far less developed than those for bacterial communities (Table 3). [First approaches to close this](#)
303 [gap have been made in recent years. For example, molecular systems to detect genes coding for laccases and other](#)
304 [ligninolytic enzymes as well as other fungal activities in carbon cycling have been published and applied](#) (Edwards
305 [et al., 2011; Chen et al., 2013; Hannula and van Veen, 2016](#)). Also for genes involved in the fungal nitrogen
306 [turnover, primer pairs have been successfully developed](#) (Gorfer et al., 2011). However here, even more than for
307 [bacteria, it is critical to link copy numbers of genes directly to the size of the respective functional population, as](#)
308 [in many cases one organism can harbour multiple operons coding for the same genes involved in a given function.](#)

309
310 The method of community level physiological profiling (CLPP) using the BiologTM system (Biolog,
311 Hayward CA, USA) was first developed in the late 1980s to identify bacteria of clinical importance by assessing
312 the consumption of 95 different carbon sources in a microtiter plate. The technique was then extended to identify
313 bacterial strains from environmental mixed microbial communities samples using select carbon sources (Garland,
314 1997). Currently, the technique is frequently used to assess the effects of contaminants on soil microbial activity
315 (Bloem and Breure, 2003; Schmitt et al., 2004). As such, the CLPP method has become a measure of microbial
316 functional diversity in soil (e.g. Gomez et al., 2006) and was used to distinguish the biodiversity of soil microbial
317 communities in monitoring programs (Rutgers et al., 2016). Even though the method is easy to use, it does have
318 some drawbacks (Winding and Hendriksen, 2007). The technique is based on the utilisation of select carbon
319 sources, which when consumed result in reduction, and thus colour change, of a tetrazolium indicator dye (Garland
320 and Mills, 1991). This reaction is based on the dehydrogenase enzyme activity of cultivable, fast growing, aerobic,
321 eutrophic microorganisms (largely bacteria). Consequently, this technique does not reflect the full spectrum of
322 microbial species within a mixed soil community. Additionally, due to the artificial growth conditions required in
323 the test, it is argued that the method does not reflect the microbial community diversity and its function of a given
324 soil (Glimm et al., 1997). On the other hand, however, standardized conditions allows for direct comparisons

325 between microbial communities in different sites, for example, independent of the abiotic conditions, thus making
326 CLPP a popular method for toxicology testing (Preston-Mafham et al., 2002).

327 Isothermal micro-calorimetry is another technique that involves the direct measurement of energetics in
328 soil and provides a functional link between energy flow and the composition of belowground microbial
329 communities at a high taxonomic level (Herrmann et al., 2014). With this method, an integrative determination of
330 the metabolic activity of soil bacteria and fungi is achieved. The integrated assessment of substances' and energy
331 turnover has high potential to elucidate the regulation of soil ecological functions. However, the substantial costs
332 for the acquisition of this very specific instrumentation is considered as a major drawback. Furthermore, the
333 measurement requires water saturation of the soil and, thus, the samples are modified. Since calorimetry has been
334 rarely used and data and publications are few, this method is considered not ready for standardization.

335 The methods targeting thymidine or leucine incorporation into microbial biomass can be used to
336 determine microbial growth rates and efficiencies (Bååth et al., 2001; Rousk, 2016). Growth rate is a fundamental
337 reference for numerous other microbial properties and functions. For example, it is required to calculate microbial
338 carbon use efficiency (CUE) as a key-parameter describing C-substrate turnover and storage in soil (Liu et al.,
339 2018; Spohn et al., 2016; Takriti et al., 2018). Furthermore, the method can be used to assess the adverse effects
340 of toxic chemicals on the microbial community (Modrzyński et al., 2016; Rousk et al., 2009a). The drawbacks of
341 these two methods are: (i) specific training is required, (ii) laboratories must have a permit to manipulate
342 radioactive isotopes, and (iii) there are higher costs for proper handling and disposal of ³H-labelled radioactive
343 material. As an alternative, the incorporation of the stable isotope ¹⁸O from labelled water into soil microbial DNA
344 can be used to distinguish growing and non-growing microorganisms based on the gradient-separation of
345 [¹⁸O]DNA and [¹⁶O]DNA (Schwartz, 2007). The ¹⁸O stable isotope method has been improved by sequencing a
346 marker gene from fractions retrieved from ultracentrifugation to produce taxon density curves; thus enabling
347 researchers to estimate the percent isotope composition of each microbial taxon's genome (Schwartz et al., 2016).
348 This method continues to be advanced and, although not used often, could have a high potential for future
349 standardization.

350 There are simplistic methods available to determine organic matter decomposition, which are indicative
351 of C cycling (**Function 6**). The tests listed in Table 3 are based on measuring the weight loss of introduced organic
352 materials of different complexity in soil over time. The tests are relatively easy to perform and inexpensive,
353 however, degradation activity is not exclusive to microorganisms but can also include invertebrates. The OECD
354 litter bag test (OECD, 2006) for site specific assessment of organic matter decomposition uses wheat straw as the

355 substrate and provides clear evidence of cellulose degradation. In general, the litter bag tests provides evidence for
356 the degradation of naturally occurring plant material in soil. Results do, however, depend on the mesh size of the
357 litter bags (increasing exclusion of soil animals with decreasing mesh size). On the other hand, plant material or
358 litter is hard to standardise with the results largely depending on the composition of the plant material. As such,
359 artificial cellulose has been successfully used for a laboratory procedure to assess organic matter decomposition
360 (Kvas et al., 2017). Another alternative to the litter bag test is the use of tea bags (Keuskamp et al., 2013). Tea
361 bags can be purchased to contain a consistent quality of material, and so this method is preferred by citizen science
362 (e.g. farmers to assess the soil quality of their land). In order to better distinguish the degrading abilities of different
363 soil microbiomes, the test could be modified to use different types of tea that contain recalcitrant material to a
364 different extent. Another test for future method development is the Bait Lamina test (ISO 18311, 2016) used to
365 assess the degradation of organic matter in field soil by grazing invertebrates (Jänsch et al., 2013; Kvas et al.,
366 2017). It is a simple test that can easily be adapted for use under controlled laboratory conditions (Jänsch et al.,
367 2017).

368 Methods for the determination and assessment of greenhouse gas emissions from soil (**Function 7**) have
369 already been standardized or are well advanced in the standardization process (Table 2). They are mostly focused
370 on measuring concentrations of greenhouse gases, like CO₂, CH₄ and N₂O, as well as their fluxes as endpoints. In
371 addition, molecular biology methods that estimate the relative abundance of functional microbial guilds or taxa
372 gives new insight into the ecology of microorganisms involved in the formation of greenhouse gases. For example,
373 the qPCR measurement of key N₂O functional genes has allowed researchers to link N₂O reduction capacity to
374 reduced greenhouse gas emissions in soil amended with organic matter (Xu et al., 2018). Additionally, the
375 quantification of functional gene sequences related to methane generation and methane oxidation, respectively,
376 yields detailed insights into the functional potential of climate change-affected permafrost soils (Yergeau et al.,
377 2010).

378 For **Function 8** (soil structure affecting soil water, gas balance and filtration function), there is clear
379 evidence that microbial activity and biomolecules substantially contribute to the formation and stability of micro-
380 aggregates, and thus to the structure, pore system and pre-consolidation stress of soils (Six et al., 2004). While
381 existing parameters, such as enzyme activities, are not clearly indicative in this regard (Beck and Beck, 2000),
382 glomalin can be considered as a biochemical marker of soil aggregation. This glycoprotein is produced by
383 microorganisms, especially arbuscular mycorrhiza fungi, and significantly increases aggregate formation and
384 stability (Rillig, 2004; Rillig and Mummey, 2006). The existing protocols for extraction (chemical extraction

385 combined with autoclaving) and determination of glomalin, either by using the Bradford protein assay, enzyme-
386 linked immunosorbent assay (ELISA), or LC-MS method (Bolliger et al., 2008; Janos et al., 2008), open the
387 possibility for its standardization in the near future. It should be noted, however, that a well-equipped and
388 experienced laboratory is required to perform this method.

389

390 **4 Transforming standardized methods into indicators of soil quality**

391 As recently underlined by the European Food Safety Agency (EFSA) in a scientific opinion ‘addressing the state
392 of the science on risk assessment of plant protection products for in-soil organisms’, there is an urgent need to
393 modernize pesticide risk assessment by implementing specific protection goals for in-soil organisms which are
394 key drivers of a wide range of functions supporting ecosystem services (Ockleford et al., 2017). There currently
395 exists a multitude of methods that can potentially be used for this task. Here, we have identified in the body of this
396 paper a number of methods that are presumably suitable for further evaluation and standardization with regard to
397 their scientific value and practical applicability. These prospective standardized methods will not only be useful
398 to identify adverse effects on the soil microbiome, but also to conduct comparable studies in laboratories all over
399 the world to define normal operating ranges of microbial activity in soil and respective quality indices and
400 threshold values.

401 It is clear that all parameters taken together reflect the potential of a microbial community to perform a
402 certain function and not solely a specific (actual) activity. This is important to understand [in order](#) to interpret the
403 values of a given endpoint in relation to both energy fluxes and compound transformation rates, which can largely
404 depend on intrinsic properties such as pedoclimatic conditions, and nutrient availability as well as extrinsic
405 properties such as anthropogenic effects, and soil cultivation measures. To make use of these methods as indicators
406 for soil quality, there are several requirements that need to be included. These involve the assessment of the
407 normal operating range of soil that include natural and dynamic fluctuations of a given endpoint. The methods
408 need to be implemented into a framework, which takes into account site-specific conditions including soil type,
409 pedoclimate and land-use. [Undoubtedly, this requires further joint efforts in order to generate comprehensive
410 databases from which normal operating ranges of values for a given proxy can be read. Such a task calls for
411 standardized methods to obtain comparable data.](#) Additionally, there is a requirement for the assessment of
412 resistance and resilience of a given microbial endpoint to see how much it is affected by a soil disturbance and
413 whether or not it can recover (e.g. return to its original state) after the disturbance has disappeared. [Here the use
414 of DNA based methods, which provide a measure of a microbial community’s potential to perform a given process.](#)

415 might be more useful than using RNA. The RNA rather indicates actual activities, which may highly fluctuate in
416 time and space, and thus are of less significance as an indicator. However, free DNA released from dead microbes
417 is often highly resistant in soil, which might result in an over estimation of a potential function. This needs to be
418 taken into account when interpreting the data. Recently, methods that extract DNA only from living cells have
419 been described, but their use has not been yet introduced into recent standardization activities.

420 Also, the use of a test battery to measure a range of interconnected endpoints is recommended (Ockleford
421 et al., 2017) to integrate different biological and other parameters (e.g. soil pH, organic carbon content) into
422 multiparametric indices (Bastida et al., 2008; Kvas et al., 2017). At present, it appears to be favourable to use a
423 suite of different methods, i.e. functional gene analysis and microbial enzyme and/or degradation activities, for
424 soil quality testing. Finally, to fully understand soil microbial functioning, a task was envisioned to investigate the
425 linkage between the genetic functional potential and the available resources, termed the soil metaphenome (Jansson
426 and Hofmockel, 2018). This will require even further integration and assessment of multiple parameters and test
427 methods. Reaching that goal will surely promote soil ecological research but, at the current stage, may clearly go
428 beyond the applied aim of standardization to release easy-to-use targeted methods.

429 The critical evaluation of existing and non-standardized methods is required to further select and
430 standardize new methods to assess soil quality. For methods linked to the molecular analysis of soil microbiomes,
431 there is a need to ensure that worldwide activities are synchronized to propose important standards that are well
432 accepted by the scientific community. ~~For example, recently, the Earth Microbiome Project~~
433 ~~(www.earthmicrobiome.org) has proposed primer pairs to barcode soil bacteria in a standardized manner.~~
434 ~~Furthermore, new bioinformatic pipelines have been developed that are being used more and more as standard~~
435 ~~procedures. Finally,~~ ~~To improve the reproducibility of data it has been agreed that a complex mixture of~~
436 ~~microorganisms (MOCK) must be implemented as a control in every experiment, but so far no common agreement~~
437 ~~on the composition of a MOCK community has been reached. The exact composition of the MOC is still under~~
438 ~~discussion.~~ However, it is clear that if further developments of microbial bar coding and/or metagenomics
439 methods are to be implemented into ISO guidelines, an MOCK is required.

440 ISO standardization committees are open circles and the presented selection and valuation of methods
441 may not be complete. Environmental scientists are solicited to propose new work items enlarging the current
442 catalogue of biological methods for future standardization. Accordingly, this opinion paper aims at initiating a
443 broader discussion intended to improve the measurement of microbial functions for soil quality assessment. Lastly,
444 it must be noted that standardization of methods is inevitably a balancing act. On one hand, standardization

445 provides defined methods that are essential to obtain comparable data, e.g. for integration in large, joint databases.
446 On the other hand, it requires setting a specific method for several years. Consequently, scientific progress cannot
447 be easily adopted, or at least with a delay, considering that standards are revised every five years, which may be
448 a barrier to the introduction of new approaches resulting from technological evolution, especially in the fast
449 developing field of molecular biology methods. Hence, it is also the aim of this paper to have an open discussion
450 to identify the best suitable methods with an assumed longer period of validity.

451
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Table 1. List of criteria for the selection of indicators for microbial functional indicators, based on Faber et al. (2013) and Pulleman et al. (2012), with slight modifications by the authors.

Criteria	Measured by	Low Score	High Score
a) Practicability	Lab equipment	Very few labs have the equipment needed	All labs would be able to carry out the work
	Skills	Specialist skills are needed	General skills would suffice
b) Cost efficiency	Capital start-up	More than €100 000	Less than €2000
	Cost per sample	More than €100	Less than €2
	Labour needed in the lab	High labour demand	Low labour demand
	Labour needed in the field	High labour demand	Low labour demand
c) Policy relevance	Focus on ecosystem processes and services	Weak links with existing or planned legislation	Strong links with existing or planned legislation
d) Sensitivity	Effect of soil properties	No response or idiosyncratic response	The indicator responds characteristically to change
	Effect of land use	No response or idiosyncratic response	The indicator responds characteristically to change
	Effect of disturbance	No response or idiosyncratic response	The indicator responds characteristically to change
e) Selectivity		Endpoint affected by numerous variables	Endpoint only affected by parameter under investigation
f) Reproducibility		Low or largely varying reproducibility among replicates	Highly reproducible
g) Use as an indicator	Status quo	Not in use already	In use already
h) Handling and availability of organisms ¹		Rare and/or difficult to obtain	Easy to obtain
		Difficult to keep	Easy to keep
		Largely varying quality/fitness	Easy to provide with constant quality/fitness
		Seasonal availability	Year-round availability
i) Fit for use as an indicator	Significance / explanatory power	Weak relationship to ecological function	Strong relationship to ecological function
	Standardized	Methods are not ready for general use or standardization (i.e. low experience, no SOPs ²)	Methods are already in general use, preferably as standard (e.g. OECD)
	Spatio-temporally relevant	Spatio-temporally only relevant for a small plot at one point in time	Representative for more than one site and/or more than one point in time
	Understandable	Difficult to explain in a policy situation	Easily understood in a policy situation
j) Experience	Literature data	Low amount of information on performance and outcome, e.g. <10 publications	High amount of information on the performance and outcome, e.g. >10 publications, existing ring test(s)
k) Data evaluation	Database	No or hardly any existing data available or not freely available	Freely available and sound database for data evaluation

928 ¹ Only relevant for faunal species. Does not apply to soil microorganisms that are tested with their natural
929 abundance in mixed communities.

930 ² Standard operating procedures

931 **Table 2.** Methods already validated and published as ISO standards for determining potential microbial biomass
 932 and activities for soil quality.

Microbial biomass and respiration (some relations to Functions 1 and 6)	
ISO 14240-1	Determination of soil microbial biomass – Part 1: Substrate induced respiration method
ISO 12240-2	Determination of soil microbial biomass – Part 2: Fumigation – extraction method
ISO 16072	Laboratory method for determination of microbial soil respiration
ISO 17155	Determination of the activity of the soil microflora using respiration curves
ISO 11063	Direct soil DNA extraction
ISO 17601	Quantification of the abundance of microbial groups in soil DNA extract
ISO/TS 29843-1	Method by phospholipid fatty acid analysis (PLFA) and phospholipid ether lipids (PLEL) analysis (
ISO/TS 29843-2:	Method by phospholipid fatty acid analysis (PLFA) using the simple PLFA extraction method
(Potential) microbial enzymatic activities: C, N and P turnover (Functions 4 and 6)	
ISO/TS 22939 ¹	Measurement of enzyme activity patterns in soil samples using fluorogenic substrates in micro-well plates
ISO- DIS 20130- EN ²	Measurement of enzyme activity patterns in soil samples using colorimetric substrates in micro-well plates
ISO/TS 23753-1	Determination of dehydrogenase activity in soils — Part 1: Method using triphenyltetrazolium chloride (TTC)
ISO/TS 23753-2	Determination of dehydrogenase activity in soils — Part 2: Method using iodotetrazolium chloride (INT)
ISO 14238	Biological methods – Determination of nitrogen mineralization and nitrification in soils and the influence of chemicals on these processes
ISO 15685	Determination of potential nitrification and inhibition of nitrification — Rapid test by ammonium oxidation
Potential microbial activities: biodegradation of pollutants (Function 3)	
ISO 11266	Guidance on laboratory testing for biodegradation of organic chemicals in soil under aerobic conditions
ISO 14239	Laboratory incubation systems for measuring the mineralization of organic chemicals in soil under aerobic conditions
ISO 15473	Guidance on laboratory testing for biodegradation of organic chemicals in soil under anaerobic conditions
Potential microbial activities: turnover greenhouse gases (Function 7)	
ISO- DIS 20951	Guidance on methods for measuring greenhouse gases (CO ₂ , N ₂ O, CH ₄) and ammonia (NH ₃) fluxes between soils and the atmosphere
ISO/TS 20131-1	Easy laboratory assessments of soil denitrification, a process source of N ₂ O emissions - - Part 1: Soil denitrifying enzymes activities
ISO/TS 20131-2	Easy laboratory assessments of soil denitrification, a process source of N ₂ O emissions - - Part 2: Assessment of the capacity of soils to reduce N ₂ O
Potential microbial activities: organic matter decomposition (Function 6)	
ISO/CD 23265 ³	Test for measuring organic matter decomposition in contaminated soil

933 ¹ Measured enzyme activities: Arylsulfatase E.C. 3.1.6.1; α-glucosidase E.C. 3.2.1.20; β-glucosidase E.C. 3.2.1.21;
 934 β-xylosidase E.C. 3.2.1.37; cellobiosidase E.C. 3.2.1.91; N-acetylglucosaminidase E.C. 3.2.1.52;
 935 phosphodiesterase E.C. 3.1.4.1; phosphomonoesterase E.C. 3.1.3.2; leucine-aminopeptidase E.C. 3.4.11.1;
 936 alanine-aminopeptidase E.C. 3.4.11.12.

937 ² Measured enzyme activities: Arylamidase E.C. 3.4.11.2; arylsulfatase E.C. 3.1.6.1; α-glucosidase E.C. 3.2.1.20;
 938 β-glucosidase E.C. 3.2.1.21; β-galactosidase E.C. 3.2.1.22; N-acetylglucosaminidase E.C. 3.2.1.52; phosphatase
 939 E.C. 3.1.4.1; acid phosphatase E.C. 3.1.4.1; alkaline phosphatase E.C. 3.1.4.1; urease E.C. 3.5.1.5.

940 ³ Degradation of cellulose under laboratory conditions.
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Table 3. Potential new methods for the ISO standardization process and assessment according to the “logical sieve” selection criteria (described in Table 1).

Method	Source	Function addressed	Assessment ¹											
			a	b	c	d	e	f	g	h	i	j	k	
Function 4. Nutrient cycling (N and P)														
Functional genes assessed by real time qPCR			1-2	3	5	5	5	5	5	5	na ²	4	4	3
Ammonium monooxygenase gene (<i>amoA</i>)	Levy-Booth et al., 2014	quantify the abundance of nitrifying microbes	1-2	3	5	5	5	5	5	5	na	4	4	3
Ammonium monooxygenase gene (<i>amoB</i>)	Norton et al., 2002	quantify the abundance of nitrifying microbes	1-2	3	5	5	5	5	5	5	na	4	4	3
Nitrogenase gene (<i>nifH</i>)	Gaby and Buckley, 2012	quantify the abundance of N fixing microbes	1-2	3	5	5	5	5	5	5	na	4	4	3
Various genes driving P turnover	Bergkemper et al, 2016	quantify the abundance of microbes driving P transformation	1-2	3	5	5	5	5	5	5	na	4	4	3
Function 5. Pest control and plant growth promotion														
Specific mtDNA sequences assessed by real time qPCR	Voříšková et al., 2017	quantify the abundance of arbuscular mycorrhiza	1-2	3	5	5	5	5	5	5	na	4	3	2
Specific ITS sequences assessed by real time qPCR	Sakakibara et al., 2002	quantify the abundance of ectomycorrhizal fungi	1-2	3	5	5	5	5	5	5	na	4	3	2
Specific ITS sequences assessed by real time qPCR	Savazzini et al., 2008	quantify the abundance of active Trichoderma fungi	1-2	3	5	5	5	5	5	5	na	4	3	2
Function 6. Carbon cycling and sequestration														
Enzyme activity of fungi	Eichlerová et al., 2012	determine activity of laccases	4	4	5	5	3	5	5	5	na	4	5	4
	Bach et al., 2013	determine activity of phenoloxidases	4	4	5	5	3	5	5	5	na	4	5	4
Community level physiological profiling (CLPP, “Biolog”)	Garland and Mills, 1991	determine degradation of a set of carbon sources	3	4	3	1	1	5	3	na	1	5	3	
Microcalorimetry	Prado and Airoldi, 2001; 2003	quantify microbial energy turnover	1	2	1	3	3	5	1	na	3	3	1	
[³ H]-leucine or [³ H]-thymidine incorporation	Bååth, 1998; Bååth et al., 2001; Rousk et al., 2009b	quantify microbial growth rate and efficiency	1	2	5	5	4	5	5	5	na	4	4	2
[¹⁸ O] incorporation into DNA from labelled water	Schwartz, 2007; Schwartz et al., 2016	quantify microbial growth rate and efficiency	2	3	5	5	4	5	5	5	na	3	2	2
Organic matter decomposition	OECD, 2006; Knacker et al., 2003	assess organic matter degradation and therefore C cycling	5	5	4	4	5	4	5	5	na	5	5	5
Litter bag technique	Bockhorst and Wardle, 2013	assess the degradation of plant litter material	5	5	4	4	5	4	5	5	na	5	5	5
Tea bag technique	Keuskamp et al., 2013	assess the degradation of tea leaves	5	5	4	4	5	4	5	5	na	5	5	5
Funct. genes within C cycle assessed by real time qPCR	El Azhari et al., 2008	quantify the abundance of microbes able to degrade protocatechuate (<i>pcaH</i>) a key intermedi-ary metabolite of the β-ketoadipate pathway	1-2	3	5	5	5	5	5	5	na	4	4	3
	El Azhari et al., 2010	quantify the abundance of microbes able to degrade catechol (<i>cat A</i>) a key intermediary metabolite of the β-ketoadipate pathway	1-2	3	5	5	5	5	5	5	na	4	4	3

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Table 3. Continued

Function 7. Greenhouse gas emissions													
Method	Source	Function addressed	Assessment ¹										
Methyl coenzyme M reductase (<i>mcrA</i>) assessed by real time qPCR	Steinberg and Regan, 2009	quantify the abundance of methane producing microbes	1-2	3	5	5	5	5	na	4	4	3	
N ₂ O reductase gene (<i>nosZ</i>) assessed by real time qPCR	Jung et al., 2013	quantify the abundance of N ₂ O reducing microbes	1-2	3	5	5	5	5	na	4	4	3	
Methane reductase gene (<i>pmoA</i>)	Kolb et al., 2003	quantify the abundance of methane reducing microbes	1-2	3	5	5	5	5	na	4	4	3	
Nitric oxide reductase gene (<i>cnorA</i>) assessed by real time qPCR	Dandie et al., 2007	quantify the abundance of methane reducing microbes	1-2	3	5	5	5	5	na	4	4	3	
Function 8. Soil structure affecting soil water, gas balance and filtration capacity													
Determination of glomalin	Bolliger et al., 2008; Janos et al., 2008; Wright et al., 1998	determine the content of glomalin in soil as a proxy of soil aggregation	3	3	4	4	4	5	3	na	4	4	2

Formatiert: Deutsch (Deutschland)

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¹ Overall scoring in case of several measures for one criterion. Fulfilment of criterion described by numbering (colour code): 1 (red) very low; 2 (orange) low; 3 (yellow) medium; 4 (light green) good; 5 (dark green) very good.

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² na = not applicable

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